The CALIPSO fish and seafood study concerning dietary consumption and biomarker of exposure to trace elements, pollutants and omega 3 was performed at the initiative of the General Directorate for Foods of France's Ministry of Agriculture and Fisheries, the French Institute for Agronomy Research and the French Food Safety Agency. Unlike traditional so-called "indirect" exposure studies based on ingestions, this study enables finer characterisation of the risks and benefits associated with fish and seafood consumption by measuring the actual biological internal levels of individuals as a function of their dietary habits and local provisioning modes.
The CALIPSO study constitutes an important scientific and methodological examination of the risksbenefits question in general and that of fish and seafood consumption in particular, a subject widely debated at national and international levels.
The study shows that French coastal populations, generally high seafood consumers, are well informed and have sound knowledge of these foods. They appreciate information on this subject which is a source of concern, yet they tend regard the public controversy on this issue with some scepticism. The study shows that the contaminant levels measured in provisioned fish and seafood are globally satisfactory relative to currently applicable regulations, with the exception of a few products. For trace elements this "background " contamination level is relatively homogeneous all along the French coast, whereas for persistent organic pollutants a North-South contamination gradient is observed.
From a benefits point of view, the study shows that consuming fish alone at least twice a week (including some oily fish) provides the recommended intake of omega 3 long-chain polyunsaturated fatty acids. As regards risks, the study reveals that although some high consumers exceed the reference toxicological values, the excesses are moderate and moreover difficult to interpret owing to the uncertainties inherent in all indirect exposure studies and the existence of safety factors. Nevertheless these results demonstrate the need to pursue the efforts being made to reduce exposure (by reducing pollution), especially to dioxins and all PCBs.
Finally, concerning the global question of weighing health risks against nutritional benefits, the study results confirm the validity of the recommendations made by various national scientific bodies: that the general population should consume fish at least twice a week, including some oily fish, and that pregnant or breast-feeding women should consume predator fish not more than once a week.
Looking beyond these general recommendations, this study highlights the advantages of diversifying the consumed fish and seafood species in terms of proportions and provisioning origins in order to ensure a rational balance between benefits and risks compatible with nutritional
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| AFSSA: | French Food Safety Agency (Agence Française de Sécurité | IPCS: | International Program on Chemical Safety |
| :---: | :---: | :---: | :---: |
|  | Sanitaire des Aliments) | JECFA: | Joint FAO/WHO Expert Committee |
| ALA: | Alpha-linolenic acid |  | on Food Additives and contaminants |
| AsB: | Arsenobetaine | JMPR: | FAO/WHO Joint Meeting Pesticide |
| AsC: | Arsenocholine |  | Residues |
| CIQUAL: | Informatics Centre for Food Quality | LA: | Linoleic acid |
|  | (Centre Informatique sur la Qualité | LOD: | Limit of detection |
|  | des Aliments) | LOQ: | Limit of quantification |
| CREDOC: | Research Center for the Study and | MBT: | Monobutyltin |
|  | Observation of Living Conditions <br> (Centre de recherche pour l'étude et | MMA: | Monomethylarsonic acid |
|  | l'observation des conditions de vie) | MOT: | Monooctyltin |
| CSHPF: | High Council for Public Health | MPT: | Monophenyltin |
|  | (Conseil supérieur d'hygiène | MRL: | Minimum Risk Level |
|  | publique de France) | MUFA: | Monounsaturated fatty acid |
| CSTEE: | Committee on Toxicity, Ecotoxicity and the Environment | OCA: | Food Consumption Observatory (Observatoire des Consommations |
| CVD: | Cardiovascular diseases |  | Alimentaires) |
| DBT: | Dibutyltin | OPCST: | Parliamentary Office for Evaluation |
| DGAL: | General Food Directorate (Direction Générale de l'Alimentation) |  | of Scientific and Technical Options (Office Parlementaire des Choix |
| DGCCRF: | General Directorate for |  | Scientifiques et Techniques) |
|  | Competition, Consumption and | PAH: | Polycyclic aromatic hydrocarbon |
|  | Fraud Prevention (Direction | PBDE: | Polybromodiphenylether |
|  | Générale de la Concurrence, de la | PCB: | Polychlorobiphenyl |
|  | Consommation et de la Répression des Fraudes) | PCDD: | Polychlorodibenzo-p-dioxin |
| DHA: | Docosahexaenoic acid | PCDF: | Polychlorodibenzofuran |
| DL-PCB: | Dioxin-like polychlorobiphenyl | POP: | Persistent organic pollutant |
| DMA: | Dimethylarsinic acid | PTWI: | Provisional Tolerable Weekly Intake |
| DOT: | Dioctyltin | PUFA: | Polyunsaturated fatty acid |
| DPT: | Diphenyltin | RDA: | Recommended Daily Allowance |
| EPA: | Eicosapentaenoic acid | REGAL: | General Foods Directory (Répertoire Général des Aliments) |
| GAA: | gamma-aminolevulinic acid | SCOOP: |  |
| GEMS/ |  |  | relating to food |
| Food Euro: | Global Environment Monitoring System/ Food Contamination | SFA: | Saturated fatty acid |
|  | Monitoring and Assessment | TBT: | Tributyltin |
|  | Program | TDI: | Tolerable Daily Intake |
| IARC: | International Agency for Research | TDS: | Total Diet Study |
|  | on Cancer | TEF: | Toxic equivalency factor |
| IFREMER: | French Research Institute for Exploitation of the Sea (Institut | TEI: | Total energy intake |
|  | Français de Recherche pour | TEQ: | Toxic Equivalent |
|  | l'Exploitation de la Mer) | TOT: | Trioctyltin |
| INCA: | Individual National Food | TPT: | Triphenyltin |
|  | Consumption Survey | TRV: | Toxicological reference value |
| INRA: | French Institute for Agronomy | WHO: | World Health Organisation |

## Preface

The diversity of environmental pollutants, largely related to unceasing industrial and technological development, presents a permanent problem when verifying food control quality from a health point of view and evaluating the risks for consumers of foods subject to contamination. This is a real challenge for scientists, health experts and the public services responsible for food safety. The problem is particularly important in that chemical pollutants are ubiquitous and extremely varied in their chemical forms and toxicological characteristics. Moreover, due to their properties and variable persistence, they enter the food chains that lead to man through vegetable and animal foods. Consequently environmental pollutants are a constant public health concern, which is why the General Food Directorate (DGAL: Direction Générale de l'Alimentation), depends on the scientific and methodological support of risk assessment experts to face this fast and ever changing challenge.

The present work was delegated by the DGAL to the French Institute for Agronomy Research (INRA: Institut National de la Recherche Agronomique) with the aim of assessing the exposure of high consumers of seafood. The French Food Safety Agency (AFSSA: Agence Française de Sécurité Sanitaire des Aliments) made a substantial methodological contribution to the study and analysed its results. This study has improved our knowledge of the dietary habits of high consumers of fish and seafood in France and of their provisioning practices. It has yielded inventories of the levels of nutritional and toxic substances in these products consumed in different regions, and it has provided exposure data for these populations thanks to a study of the biological impregnation to contaminants. The outcome of the study is therefore an evaluation of the risks relative to the consumption of seafood products, which are then balanced with the nutritional benefits.

The study results shed new light on the relationships between diet and health and will help to better protect and inform the consumer. In addition, they enable France to make a useful contribution to the scientific and regulatory studies whether national, European or international levels.



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# FIRST PART 

## Methodology and

 general presentation
## Introduction

The scientific approach to the evaluation of the nutritional benefits of nutritive elements (minerals, fatty acids, etc.) in food and the health risks related to toxic substances (pesticides, heavy metals, natural toxins and other environmental contaminants) is to estimate the quantity of ingested substances then compare this with nutritional baselines and/or toxicological reference values. These nutritional and toxicological baselines defined by expert scientific committees form part of global public health programmes intended to confirm optimal cover of physiological needs and the absence of adverse effect on consumer health.

The evaluation of the dietary intake of a given nutriment or contaminant, also known as the "dietary exposure" is based on various approaches including the traditional one that consists in proffing consumption data with composition or contamination data ${ }^{1}$. In most cases this so-called "indirect" or "food exposure" approach provides a response to health questions posed by the national authorities responsible for evaluating and managing food risks.

In parallel with this first approach, a "direct" measure of the intakes by exposure biomarkers, complementary to the first, can be made if necessary in order to better characterise the benefits and/or the possible risks of a particular substance as regards consumer health. This method has the advantage of evaluating in situ, in other words in the biological tissue of individuals, the internal level of a nutriment and/or contaminant of interest; it also takes into account exposure channels other than food. However the results are not easy to interpret.

For many years seafood such as fish, molluscs and crustaceans has often been the focus of attention in nutritional and toxicological work. Nutritionists consider these products to be an important source of high-quality proteins, minerals and essential fatty acids such as omega 3, although only half the population follow the recommendation of the National Nutrition and Health Programme (PNNS) to consume fish at least twice a week ${ }^{2}$. Toxicologists tend to regard seafood as a major vector for toxic substances such as metal trace elements and persistent organic pollutants. The scientific reality is more complex and a reconciliation of these two viewpoints requires that we take into consideration both nutritional and toxic substances contained in food products and also consumer behaviour with regard to these products.

Concerning intakes of the omega 3 family of long-chain polyunsaturated fatty acids, today there is very little available data on the fatty acid content of fish and seafood or on biomarkers of exposure to omega 3 fatty acid in the French population, in particular in people consuming large quantities of seafood (other than through food supplements). Many studies have already demonstrated the involvement of the fatty acids eicosapentaenoic acid (EPA) and docosahexaenoic acid (DHA), found mainly in fish and seafood, in the mechanisms protecting against certain pathologies, notably cardiovascular disease.

As regards exposure to trace elements, several recent studies have shown that for the average consumer in the general population the toxicological limits are not exceeded. Nevertheless, the absence of risk for the average consumer does not exclude a risk for heavy consumers, as underlined in many studies (French Total Diet Study (TDS) 2004³, INCA Study 19994, reports of the French Upper Council for Public Health (CSHPF) in $1996^{5}$ and the Parliamentary Office for the Assessment of Scientific and Technical Choices (OPECST) in $2001^{6}$ ). Moreover, the absence of French data on the levels of "indirect" or "direct" exposure to certain substances (omega 3 and pollutants in particular) of populations consuming large quantities of seafood does not enable a quantified assessment of the benefits or risks associated with these dietary habits, a situation that is obviously prejudicial to the global health risk evaluation and management process.

Consequently, it is necessary to push the analysis further by performing, first, a representative study of several French coastal populations that are regular consumers of seafood, and of their local provisioning and self-provisioning practices (beach fishing), and secondly a representative study of the biologicals levels of pollutants and omega 3 in these populations. The choice of coastal regions for the study is justified by their particularly high seafood consumption.

The objective of this study is therefore to make a survey of the nutritional intakes and exposure to trace elements and persistent organic pollutants of high fish and seafood consumers by examining their food consumption habits and analysing the real risks of these habits as well as their nutritional benefits, notably those associated with polyunsaturated fatty acids.

This study is described in three distinct and complementary parts (Figure 1):

- a consumption study focused on heavy fish and seafood consumers in four French coastal regions;
- a study of blood and urinary biomarkers associated with intakes of fatty acids and exposure to contaminants in a sub-sample of consumers having participated in the food consumption survey;
- a study of levels of fatty acids and contamination by trace elements and persistent organic pollutants in the seafood bought and consumed by the populations of the four coastal regions, applying a standardised international methodology of the "total diet study" (TDS) type ${ }^{7}$.

[^0]Figure 1: General outlines of the Calipso study


### 1.1 Present situation

### 1.1.1 Fatty acids

Fish consumption and cardiovascular diseases
Many studies have demonstrated that fish consumption correlates inversely with coronarian mortality. More specifically, such mortality is observed to decrease by $15 \%$ among populations consuming fish at least once a week ${ }^{8}$. A 20 g increase in daily fish consumption reduces the coronary heart disease mortality risk by $7 \%$. Moreover, these trends are accentuated in the case of oily fish.

One of the hypotheses proposed to explain this protective effect of fish consumption is their richness in fatty acids, in particular in polyunsaturated fatty acids of the omega 3 class.

## Fatty acids

Fatty acids are organic molecules composed of a carbonyl chain terminating on a carboxyl group. They are characterised by the length of their carbonyl chain, the number of double bonds and their position on the chain. We can therefore distinguish saturated fatty acids (SFA) with no double bonds, monounsaturated (MUFA) with a single double bond, and polyunsaturated (PUFA) with several double bonds. The PUFAs can be divided into four classes according to the position of the first unsaturation relative to the carbon atom at the methyl end: n-7 (omega 7), $n-9$ (omega 9), $n-6$ (omega 6) and n-3 (omega 3).

While saturated, monounsaturated and some polyunsaturated fatty acids ( $n-7$ and $n-9$ classes) can be synthesised by the organism (Figure 2), the omega 3 and omega 6 precursors (alpha-linolenic and linoleic acid, respectively) must be provided by food; these are referred to as "essential fatty acids".

The physiological role of fatty acids is first and foremost energetic. But the polyunsaturated fatty acids of the essential $n-6$ and $n-3$ classes are above all important constituents of many structures (membrane phospholipids bringing fluidity and their properties to membranes) and some are precursors of oxygen mediators notably involved in the processes of inflammation and blood platelet aggregation (prostaglandins, thromboxanes, leukotrienes, etc.) ${ }^{910}$.

## Fatty acids and cardiovascular diseases

Many epidemiological studies have confirmed the adverse effects of an excess of SFAs in the development of cardiovascular diseases (CVD) and the associated mortality. Excessive SFA consumption correlates positively with mortality due to CVD and with factors of high risk of CVD. On the other hand, the consumption of MUFAs and above all PUFAs correlates negatively with CVD. The protective role of the n-3 PUFAs has been demonstrated in primary prevention and above all in secondary prevention of CVD. Long-chain omega 3 might help to reduce mortality, though not morbidity ${ }^{10}$. Long-chain omega 3 supplements might help to reduce cardiovascular risks by lowering the risk of sudden death of people with a history of cardiovascular problems. However they do not reduce the incidence of non-mortal cardiac infarct. More particularly, the protective role of alpha-linolenic acid (ALA), the precursor of the n-3 long-chain PUFAs, has been demonstrated in several clinical intervention studies, in particular those concerning prevention of sudden death in man ${ }^{9}$. Linoleic acid (LA), precursor of the $n-6$ long-chain PUFAs, tends to lower the cholesterolemy but it does not appear to reduce cardiovascular mortality. Generally speaking, the n-6 PUFAs have a lipid-lowering effect; they reduce the LDL-cholesterol, but they have no effect on the circulating triglycerides, whereas the n-3 PUFAs have a hypotriglyceridemiant effect (at least in certain population groups), although this concerns only the very long-chain compounds, eicosapentaenoic acid (EPA) and docosahexaenoic acid (DHA).

## Fatty acids and cancer

Today only experimental studies on animal models provide some information on the role of fatty acids in cancerous processes. The n-6 PUFAs encourage tumorous growth while the n-3 PUFAs have a protective effect ${ }^{1112}$. However the mechanisms by which fatty acids act on the cellular cycle (modulation of the expression of proteins regulating the cellular cycle and the apoptosis) are not well understood.

[^1]
## Fatty acids in food

The principal dietary sources of the precursors ALA (alpha-linolenic acid) and LA (linoleic acid) are vegetable oils and animal products. For example, rapeseed, nut and soyabean oils are rich in ALA and sunflower and maize oils are rich in LA. Animal products, in particular fish, seafood and breastmilk, provide long-chain n-3 PUFA compounds in substantial quantity. The n-6 PUFAs are found in quantity in terrestrial animal products, in particular meat and eggs and also in breastmilk.

Leaving aside dietary supplements, seafood products remain the major source of long-chain n-3 PUFAs since in humans the conversion of the precursor ALA into these long-chain derivatives is low: it has been shown that less than $1 \%$ of the ALA is converted into DHA ${ }^{13} 14$.

Figure 2 : Conversion capacity of fatty acids in animals and plants


[^2]
## Needs and intakes of the French population

Table 1 presents recommended daily allowances (RDA) determined for adults on the basis of plasmatic parameters for SFAs, MUFAs and PUFAs.

Table 1: National nutritional recommended daily intake of fatty acids in adults ${ }^{9}$

| kcal. $\mathrm{d}^{-1}$ |  | SFA | MUFA | LA | ALA | LC-PUFA | DHA | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Adult man | g. $\mathrm{d}^{-1}$ | 19.5 | 49 | 10 | 2 | 0.5 | 0.12 | 81 |
| 2200 | \% TEI | 8 | 20 | 4.0 | 0.8 | 0.20 | 0.05 | 33 |
| Adult woman | g.d ${ }^{-1}$ | 16 | 40 | 8 | 1.6 | 0.40 | 0.10 | 66.0 |
| 1800 | \% TEI | 8 | 20 | 4.0 | 0.8 | 0.20 | 0.05 | 33 |
| Pregnant woman | g.d ${ }^{-1}$ | 18 | 45.5 | 10 | 2.0 | 1 | 0.25 | 76.5 |
| 2050 | \% TEI | 8 | 20 | 4.4 | 0.9 | 0.4 | 0.1 | 33.7 |
| Breat-feeding woman | g. $\mathrm{d}^{-1}$ | 20 | 50 | 11 | 2.2 | 1 | 0.25 | 84.2 |
| 2250 | \% TEI | 8 | 20 | 4.4 | 0.9 | 0.4 | 0.1 | 33.7 |
| Older subject | g. $\mathrm{d}^{-1}$ | 15 | 38 | 7.5 | 1.5 | 0.40 | 0.10 | 62.5 |
| 1700 | \% TEI | 8 | 20 | 4.4 | 0.9 | 0.4 | 0.1 | 33.7 |

TEI: Total energy intake (lipids, carbohydrates, proteins)
LC-PUFA: Long chain polyunsaturated fatty acids
Ideally, in the diet, the LA ( $18: 2 n-6$ ) / ALA ( $18: 3 n-3$ ) ratio should tend towards 5 . This ratio was determined taking into account the existence of competition between the $n-3$ and $n-6$ PUFA classes in various enzymes (the desaturated $\Delta 6$ and $\Delta 5$ ) involved in the conversion of the precursors ALA and LA into essential long-chain PUFA derivatives (Figure 2).

To obtain the beneficial effects on the cardiovascular system and neurodevelopment some institutions recommend daily consumption of 0.2 g to 0.5 g of $\mathrm{n}-3$ LC-PUFA ${ }^{15}$. The American Heart Association recommends daily consumption of 1 g of EPA and DHA ${ }^{16}$.

## Fatty acids and pregnant women

There have been few studies on the fatty acid needs of pregnant and breast-feeding women. Taking into account the needs of the foetus, the development needs of the placenta and state of pregnancy, the recommended daily allowances for pregnant women are estimated to be 10 g of LA and 2 g of ALA. For breast-feeding women, the daily needs are 11 g of LA and 2.2 g of ALA (Table 1).

DHA and arachidonic acid (AA) are essential in the development of the central nervous system of the foetus, in particular during the last quarter of gestation when the synthesis of brain cells is fastest. These two fatty acids are incorporated in the cellular membranes contributing to their structure and functions. However it appears that the conversion rate of the ALA into DHA is very low, which implies that the DHA content of cellular membranes depends more on dietary intake of DHA than on ALA. The RDA of DHA has been fixed at 0.25 g for pregnant and breast-feeding women.

## Fatty acids and elderly people

In view of the lower energy needs after 65 years of age, the RDAs for elderly people are lower than for younger adults: 7.5 g of LA and 1.5 g of ALA. Concerning the long-chain $\mathrm{n}-3$ PUFAs, there have been

[^3]indications of a reduction of the ability to convert EPA into DHA and/or an alteration of the oxidation of these fatty acids in elderly people ${ }^{9}$, an alteration that has also been demonstrated on animal models ${ }^{1718}$.

Fatty acid intakes in the French population, in particular ALA and LA, were assessed by means of a panel of 5,008 volunteers aged between 35 and 60 in the SU.VI.MAX study (supplementation of vitamins and antioxidant minerals). The consumption data of the study were crossed with ALA and LA composition data supplied by the British Ministry of Agriculture, Fisheries and Food (MAFF), the American Department of Agriculture (USDA), and in France, the Informatics Centre for Food Quality (CIQUAL), the Meat Information Centre (CIV) and the Institute for Fats and Oils (ITERG).

Table 2: Intakes of linoleic and alpha-linolenic acids in France (data from SU.VI.MAX) ${ }^{10}$

|  |  | Min | P5 | Mean | P95 | Max |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18:2 n-6 (LA) | M | 1.53 | 2.81 | 4.26 | 6.21 | 10.54 |
| (\% TEI) | W | 1.62 | 2.91 | 4.38 | 6.31 | 11.63 |
| 18:3 n-3 (ALA) | M | 0.21 | 0.30 | 0,39 | 0.52 | 1.52 |
| (\% TEI) | W | 0.19 | 0.32 | 0.41 | 0.55 | 1.11 |
| Ratio 18:2 n-6 / 18:3 n-3 | M | 5.5 | 7.5 | 11.1 | 16.1 | 33.8 |
|  | W | 4.5 | 7.3 | 10.8 | 15.7 | 34.6 |

This work has shown that ALA intakes are lower than the RDAs ( $0.8 \%$ of total energy intake), regardless of age and sex (Table 2). The 18:2 n-6 / 18:3n-3 ratio is too high, ranging from 5 to 34 with an average of about 11, compared to the RDA of 5 .

On the other hand, to date there are no available data on intakes of LC-PUFA omega 3 (EPA and DHA) in the French population.

### 1.1.2 Trace elements

## Mercury

Mercury (Hg) is a chemical compound used in many industrial activities (batteries, electric equipment, chemical industry, paints, dental amalgams). These sources are both environmental and anthropogenic, notably due to the combustion of fossil fuels, industrial releases and waste incineration.

The organic forms of mercury, in particular methylmercury (MeHg), are more toxic than the inorganic ones. The World Health Organisation (WHO) estimates that $99 \%$ of the MeHg absorbed by the population comes from food ${ }^{19}$. The major source of MeHg exposure for humans is fish ${ }^{2021} 22$. Metallic mercury is transformed into organic mercury by marine bacterial microflora which makes it bioavailable and explains its marked ability to cumulate in shellfish and predator fish near the top of the trophic chain. Following ingestion, methylmercury is rapidly absorbed in the digestive tract and $90 \%$ is found in the blood. It then passes the hemato-encephalic barrier and concentrates essentially in the central nervous system.

[^4]In humans and animals the brain is the main organ targeted by MeHg . This neurotoxicity appears in the adult brain but even more so in the developing brain. Moreover, MeHg diffuses easily through the placenta and is also found in human milk.

In both humans and animals the neurotoxic effects observed after exposure to high MeHg levels (concentrations greater than $100 \mathrm{mg} / \mathrm{kg}$ in human hair) are essentially sensorial, visual, auditory and motor deficiencies. On the other hand, in humans, the effects of exposure to low doses of MeHg are more difficult to identify.

In adults, epidemiological studies in Amazonia have revealed the presence of alterations of visual, somatosensorial and motor functions, and of memory, attention, learning ability and manual dexterity in people whose hair mercury level is $6 \mathrm{mg} / \mathrm{kg}$ or more ${ }^{2223242526}$. The blood and the hair are good markers of MeHg exposure, in particular in conditions of uniform dietary regime, and these concentrations are linked to those of MeHg in the brain, which enables good estimation of the effects of mercury on health and the central nervous system ${ }^{27}$. Nevertheless the authors underline that the measured levels do not enable confirmation of a dose-response effect, since the concentrations measured in the biological matrices at the time of the study do not necessarily correspond to the earlier exposures that caused the observed adverse effects ${ }^{22}$.

In children the main prospective epidemiological studies carried out in the Seychelles, New Zealand and the Faeroe Islands point to a correlation between dietary exposure to mercury during pregnancy and the appearance of neurological symptoms in children ${ }^{28}$. However different populations appear to have different sensitivity to mercury, which can also be affected by dietary habits and exposure to other contaminants.

Apart from its effects on the central nervous system, MeHg also appears to be able to affect the immunitary system of adults and the developing immunitary system ${ }^{293031 \text {. Further experimental studies are necessary }}$ to confirm and clarify the mechanisms of this immunotoxicity.

In 1990 the WHO established a provisional tolerable weekly intake (PTWI) of $3.3 \mu \mathrm{~g}$ of $\mathrm{MeHg} / \mathrm{kg}$ bw (kilograms of body weight) based on evaluations made by the JECFA (Joint FAONWHO Expert Committee on Food Additives and Contaminants) from 1972 to $1989{ }^{19}$. However, in order to assure better protection for foetuses and infants, the WHO issued warnings for pregnant and breast-feeding women. In France, the CSHPF in 1998 and the AFSSA in $2002^{32}$ recognised the existence of sensitive groups (pregnant and breast-feeding women, very young children, fishermen operating in highly contaminated zones) and recommended the provision of specific information to encourage these particular groups to diversify the species of fish they consume. Following the publication of new results, the JECFA lowered the PTWI

[^5]to $1.6 \mu \mathrm{~g} / \mathrm{kg}$ bw in $2003^{22}$. This intake corresponds to a steady state concentration in the mother's blood of $56 \mu \mathrm{~g}$ of MeHg per litre assessed from a NOAEL for hair of $14 \mathrm{mg} / \mathrm{kg}$ (taking into account a hair/blood ratio of 250) that does not have any appreciable adverse effects on the foetus. The PTWI takes into account uncertainty factors: 2 for the variability between individuals of the relationship between the MeHg concentration measured in hair and that measured in the blood, and 3.2 for the inter-individual variability (pharmacokinetic component) of the relationship between the dietary intake of MeHg and the concentration measured in the blood. The AFSSA Opinion dated March $2004{ }^{33}$ confirmed the validity of this PTWI for the most sensitive population groups: pregnant and breast-feeding women and young children. The Food Standards Agency's advisory committees on contaminants in the United Kingdom has stated that in view of these new toxicological data there appears to be no reason to revise the PTWI established previously by the JECFA at $3.3 \mu \mathrm{MeHg} / \mathrm{kg}$ bw for the general public, with the exception of sensitive populations ${ }^{35}$.

In France, exposure studies to date tend to show that values close to or greater than the PTWI of $1.6 \mu \mathrm{~g} / \mathrm{kg}$ $\mathrm{bw} /$ week can be reached by certain categories of high fish and seafood consumer, and notably very young infants and women of child-bearing age ${ }^{33} 36$. Better estimation of the exposure of the more sensitive groups taking into account both the species of fish consumed and their origin is necessary to correctly evaluate the risk run by these groups and, if need be, to enable the provision of better information - or even better recommendations than those made today - on dietary consumption.

## Cadmium

Cadmium (Cd) is a contaminant found in the environment and in particular in the soil, due to erosion and human and agricultural activities. It thereby enters the food chain. In non-smoking individuals the main source of cadmium exposure is food. The most highly contaminated foods are molluscs, offal, leaf vegetables and cereals ${ }^{2021}$. Digestive absorption of cadmium is low (about $5-10 \%$ ). On the other hand, cadmium is a cumulative toxin whose biological half-life is very long (estimated to be 20-30 years in humans). The International Agency for Research on Cancer (IARC) ${ }^{37}$ classifies cadmium as "carcinogenic for man" (category 1).

Cadmium has numerous toxic effects, but the main impact on the organism of prolonged exposure to cadmium in both man and animals is on the renal function. The nephrotoxic effects are characterised by degeneration of the proximal tubules and proteinuria ${ }^{38394041}$. The risk associated with this degeneration starts to increase when the urinary excretion of cadmium exceeds $2.5 \mu \mathrm{~g} / \mathrm{g}$ creatinine. The JECFA Committee considers this to be the value for which there is an absence of prevalence of renal tubular malfunction ${ }^{42}$. In man, these alterations of the renal function can be accompanied by bone damage with osteomalacia

[^6]and demineralisation ${ }^{43}{ }^{44}$. Additionally, relationships exist in man and animals between cadmium exposure and retarded foetal growth ${ }^{45}$; reduced fertility in males has also been reported ${ }^{4647}$. On the other hand, there is no confirmed relationship between dietary exposure to cadmium and arterial hypertension or cancer.

In France, the average daily intake of cadmium was estimated to be $19.6 \mu \mathrm{~g}$ for adults in $1998^{21}, 17 \mu \mathrm{~g}$ in $2000^{48}$ and $3.6 \mu \mathrm{~g}$ in $2003^{49}$. Following the first French total diet study (TDS) ${ }^{3}$, the latest estimations in 2005 indicate an average daily intake of $2.7 \mu \mathrm{~g}$ in people over 15 years old ${ }^{50}$, which represents about $4 \%$ of the PTWI of $7 \mu \mathrm{~g} / \mathrm{kg}$ bw/week established by the JECFA using a theoretical prediction model estimating the relationships between the dietary intake of cadmium, urinary excretion and associated prevalence of renal tubular malfunction ${ }^{42}$.

The most contaminated foods are offal and seafood, notably molluscs. Seafood represents $8 \%$ to $25 \%$ of dietary intake of cadmium ${ }^{51}$. Vegetables, potatoes and similar products, due to their importance in human diets, are also major vectors of dietary exposure ( $23.7 \%$ and $21.2 \%$ respectively) in the general population ${ }^{50}$.

## Lead

Lead $(\mathrm{Pb})$ is an environmental pollutant found in soil and the atmosphere, in particular in the neighbourhood of industrial sites and heavy automobile traffic.

The dietary intake of lead comes mainly from drinks ${ }^{50}$, fresh fruit, vegetables and cereals. The contribution from drinks, which was non-negligible just a few years ago, is tending to decline rapidly as production methods are improved.

One of the major effects of lead on the organism is its hematological toxicity of which anaemia is the most common symptom. The presence of lead in the blood is the principal biomarker of lead exposure. Lead acts on the biosynthesis of the heme, inhibiting two key enzymes, gamma-aminolevulinic acid dehydratase (ALA-D) and ferrochelatase. In adults, urinary ALA-D excretion and free porphyrins of the erythrocytes, whose level is linked to the ferrochelatase, are exploited as biological markers for lead exposure ${ }^{52}$. In 1991 it was shown that lead exposure reduces the erythrocyte's defences against oxidation and shortens its life ${ }^{53}$.

[^7]Many studies have long demonstrated a correlation between hypertension and professional exposure to lead ${ }^{5455}$.

However, the most worrying impact of lead remains the neurotoxic effects (saturnism) it can cause. Lead perturbs the liberation of neuromediators by nerve cells and can pass the hemato-encephalic barrier. Epidemiological studies have shown that exposure of the foetus to small doses of lead can cause congenital abnormalities ${ }^{56}$. Exposure to doses of lead that do not result in the appearance of saturnism symptoms during infancy can nevertheless cause durable neuro-behavioural handicaps (reading difficulties, lower intellectual performance, absenteeism, etc. $)^{57}$.

The annual dietary intake of lead by the French population was estimated between 1978 and 1980 to be 60 mg , equivalent to $30 \%$ to $50 \%$ of the PTVI of $50 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} /$ week fixed by the WHO in 1972. In 1987, the JECFA revised this dose to $25 \mu \mathrm{~g} / \mathrm{kg}$ bw/week for children, then extended it to adults. The particular sensitivity of pre- and post-natal infants to the neurotoxic effect of lead implies that the dietary intake should be reduced for pregnant women, but no modification of the PTWI has been proposed. Dietary intake of lead has fallen substantially in industrialised countries with the disappearance of leadbased solders, the introduction of lead-free gasoline and improvements in food production and transformation practices, notably in vinification and wine making processes and food can soldering. The average daily dietary intake of the adult population was estimated to be $68 \mu \mathrm{~g}$ in $1998^{21}, 52 \mu \mathrm{~g}$ in $2000^{48}$ and $34 \mu \mathrm{~g}$ in $2003^{49}$, and it has been more recently estimated to be $18 \mu \mathrm{~g}^{50}$, equivalent to $7 \%$ of the PTWI.

Seafood consumption (fresh fish, crustaceans and molluscs) accounts for $3 \%$ to $11 \%$ of lead intake via food. As for mercury and cadmium, we can suppose that values close to or even exceeding the PTWI are reached by the highest consumers. An estimation of the exposure of these people is therefore necessary.


#### Abstract

Arsenic

Arsenic (As) is a soil contaminant naturally present in the environment but whose main anthropogenic origins are the use of phytosanitary products, atmospheric releases from incineration installations and industrial activity. The organic forms of arsenic - arsenobetaine (AsB), arsenocholine (AsC), monomethylarsonic acid (MMA), dimethylarsinic acid (DMA), etc. - are the predominant forms in food matrices. Whilst public health organisations still consider inorganic arsenic ( $\mathrm{As}(\mathrm{III})$ and $\mathrm{As}(\mathrm{V})$ ) to be the most dangerous forms and the methylated forms to be less harmful, in recent years scientists have revised this position as analytical speciation techniques continue to improve and in the light of the latest toxicological studies on the different forms of contaminants 585960 . These studies reveal that DMA is probably carcinogenic or promotes carcinogenesis and that the MMA(III) and the DMA(III) are genotoxic ${ }^{58}$. In man, pentavalent arsenic $\mathrm{As}(\mathrm{V})$ is reduced to trivalent arsenic $\mathrm{As}(\mathrm{III})$, itself methylated into methylarsonic and dimethylarsinic acids, most of which are then excreted.


[^8]The IARC classifies arsenic as "carcinogenic for man" (category 1). Severe exposure to arsenic results in vomiting, abdominal pain and diarrhoea; prolonged exposure to low doses of arsenic in drinking water can cause cancer of the skin, lung, bladder or kidney, and cutaneous disorders such as hyperkeratosis and pigmentation modifications ${ }^{6162}$.

In 1989 the WHO ${ }^{38}$ fixed a PTWI of $15 \mu \mathrm{~g} / \mathrm{kg}$ bw/week for inorganic arsenic.
The exposure to total arsenic can be of atmospheric origin and it is also increased by cigarette smoking, but dietary exposure remains the prime source. Much of the arsenic comes from fish and seafood. The scientific literature reports that $0.4 \%$ to $5.3 \%$ of the arsenic present in fish and seafood are in the form of inorganic arsenic ${ }^{63}$. The average daily intake of total arsenic in Europe was estimated in 2003 to be $125 \mu \mathrm{~g}$ in adults, and the contribution of seafood to this exposure exceeded $50 \%{ }^{64}$. However little contamination data on all seafood is available today. More particularly in France, the average daily intake for the adult population was estimated in 2000 to be $109 \mu g^{48}$ and in 2003 to be $147 \mu g^{49}$ of which 135 $\mu \mathrm{g}(95 \%)$ comes from fish. However this report underlines the difficulty in evaluating precisely the exposure to arsenic due to the diversity of its origins (sea fish, river fish, fish farms, etc.). European data provide average contamination levels for fish and seafood (including molluscs, crustaceans and echinoderms) ranging from less than $0.1 \mu \mathrm{~g} / \mathrm{g}$ to $18 \mu \mathrm{~g} / \mathrm{g}^{64}$. A more recent estimation indicates an average daily intake of $62 \mu \mathrm{~g}$ for adults ( $6.2 \mu \mathrm{~g} / \mathrm{kg}$ bw/week for a person weighing 70 kg ), of which $62 \%$ comes from seafood ${ }^{50}$.

## Organotin compounds

Organostannic compounds (OTC) present in the environment are mainly of anthropogenic origin. They are used as stabilizers and catalysts, vermifuges in the composition of plastics, biocides in paints, and in washing products and pesticides. Sludges from sewage treatments plants and industrial and agricultural activities are the main sources of environmental contamination.

Water pollution leads to contamination of living marine organisms almost permanently exposed to organic tin. These active substances are very certainly responsible for the toxic effects observable in marines species at very low doses, such as growth and reproduction disorders in oysters and sex changes in certain gasteropods.

In humans, food - and seafood in particular ${ }^{65}$ - is the principal origin of organic tin absorption. Trisubstituted organic tin - tributyltin (TBT) and triphenyltin (TPT) - appear to be the most toxic. TBT causes endocrine perturbations and TPT affects the reproductive system and development ${ }^{65}$. Generally speaking, the organostannic compounds such as dibutyltin (DBT), tributyltin (TBT) and triphenyltin (TPT) are immunotoxic, causing a drop in the lymphocytes in the thymus and peripheral lymphoid organs ${ }^{6667}$.

[^9]In 2004 the European Food Safety Authority (EFSA) ${ }^{68}$ established a group tolerable daily intake (TDI) of $0.25 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} /$ day for tributyltin (TBT), dibutyltin (DBT), triphenyltin (TPT) and dioctyltin (DOT) based on TBTO molecular mass; this group TDI is $0.1 \mu \mathrm{~g} / \mathrm{kg}$ bw when expressed as Sn content, in line with the CSTEE proposals (Committee on Toxicity, Ecotoxicity and the Environment) in $2003{ }^{69}$ and the WHO recommendations in 1999.

The rarity of the contamination data, in particular in France where they are almost inexistent, makes it difficult to evaluate population exposure. The data published in the SCOOP task 3.2.13 ${ }^{65}$ or in the European OT-SAFE report ${ }^{70}$ are difficult to exploit for purposes of risk evaluation due to their qualitative and quantitative disparity. However these two studies do show that shellfish accumulate more organic tin than fish and crustaceans.

### 1.1.3 Persistent organic pollutants

Persistent organic pollutants are environmental contaminants mainly of anthropogenic origin (industrial and agricultural utilisations over the last 30 years) which, mainly due to their lipophilic nature, accumulate in food chains. Their have multiple effects on human health including genotoxicity, embryotoxicity and endocrine perturbations.

## Dioxins and "dioxin-like" polychlorobiphenyls

Polychlorodibenzo-p-dioxins (PCDD) and polychlorodibenzofurans (PCDF) are compounds of similar molecular structure and physicochemical properties. There are 210 dioxin congeners, 75 PCDDs and 135 PCDFs. However only the toxicity of the 2.3.7.8-tetrachloro-dibenzo-p-dioxin (2.3.7.8-TCDD), the "Seveso dioxin", has been widely verified. This therefore serves as a reference in the calculation of toxicities as so-called Toxic Equivalents (TEQ) by applying Toxic Equivalent Factors (TEF) to the 16 other dioxin congeners and furans similar to 2.3.7.8-TCDD.

The polychlorobiphenyls (PCB) include 209 congeners that differ only in the number and position of the chlorine atoms on the biphenyl molecule. Twelve PCBs have toxicological properties similar to those of dioxins and are therefore referred to as "dioxin-like PCBs" (DL-PCB). TEF weightings are also applied to the DL-PCBs and they form part of the toxicity calculation along with the PCDDs and PCDFs.

In 2001, the JECFA fixed a provisional tolerable monthly intake for PCDDs, PCDFs and DL-PCBs at 70 pg TEQ ${ }_{\text {who }} / \mathrm{kg} \mathrm{bw}^{71}$.

[^10]
## "Indicator" polychlorobiphenyls

The seven congeners called "indicator PCBs" (i-PCB) 28, 52, 101, 118, 138, 153 and 180 have properties different from DL-PCBS. They have antithyroidian and neurotoxic effects. It is estimated that the exposure to i-PCB account for half the exposure to total PCB congeners due to their tendency to cumulate in food matrices and their toxicological impact on man ${ }^{72}$. Recent European studies indicate for adults an average daily intake of 0.01 to $0.045 \mu \mathrm{~g} \mathrm{i} \mathrm{PCB} / \mathrm{kg} \mathrm{bw}^{73}$.

Concerning all the PCBs, in 2002 the WHO proposed a TDI of $0.02 \mu \mathrm{~g} / \mathrm{kg}$ bw, in Aroclor Equivalent ${ }^{74}$. The i-PCB analysis results must be multiplied by two to be expressed in Aroclor Equivalent. The calculated exposure is then compared with the TDI.

## Polybromodiphenylethers

Polybromodiphenylethers (PBDEs) are flame retardants used in plastics and textiles. Since the 1970s they have accumulated in food chains, in aquatic biotopes in particular. These compounds are hepatotoxic, embryotoxic and also have antithyroidian effects, which is particularly worrying in view of the PBDE concentrations found in human milk ${ }^{75}$.

To date no PTWI has been fixed for PBDEs at national, European or international level.
Fish and seafood are major contributors to dietary exposure to persistent organic pollutants: $25 \%$ to $30 \%$ for the 17 congeners of dioxin and furan type ${ }^{7879}, 75 \%$ for PCBs (from the i-PCBs) ${ }^{78}$, and $30 \%$ for the 7 PBDE congeners $(28,47,99,100,153,154,183)^{80}$. In France, data from the INCA survey point to an estimated daily dietary intake of PCDDs and PCDFs in adults of 1.45 pg TEQWHO/kg bw in 200078, and $0.5 \mathrm{pg} / \mathrm{kg}$ bw in 200579. The daily intake of DL-PCBs was estimated to be $1.2 \mathrm{pg} / \mathrm{kg}$ bw in $2006^{79}$.

[^11]
### 1.2 Study methodology

### 1.2.1 Selection of study zones and populations

The four coastal zones selected for the seafood consumption survey are MediterranéeNar, Normandie/Baie de Seine, Bretagne sud, and Gironde/Charente Maritime sud.

The populations in these regions are the highest consumers of fish and seafood, as confirmed by a Food Consumption Observatory study in 1996 (OCA-CREDOC ${ }^{81}$ ). For example, as regards fish the highest consumptions are in Nord-Pas-de-Calais, Picardie, the Parisian Region, Haute-Normandie, Basse-Normandie, Pays de la Loire, Poitou-Charentes, Aquitaine, Languedoc-Roussillon, Provence-Alpes-Côte d'Azur and Corsica where annual consumption per household is 18.5 to 32.7 kg . It appears that fish consumption declines as we move away from the coast, which may be partly explained by widespread self-procurement of seafood by coastal populations. This hypothesis is supported by a CREDOC ${ }^{82}$ survey of 400 representative households in the north of the Cotentin peninsular which reveals that $20 \%$ of the seafood consumed (fish, crustaceans and shellfish) is self-provisioned. Moreover we also note that the coastal regions have the highest number of retail seafood outlets ${ }^{838}$ (Figure 3).

Figure 3: Retail trade for fish in 2002


## Retall trade for flsh in (2002) Number of outlets for 100,000 inhabitants

Light bleue : 0 a 2.5
Middle bleue : 2.5 a 5
Dark bleue : 5 al 30
There are many more fisheries near the coasts. The proximity of the sea explains why consumers are more tempted to buy fish, and when they do, to buy it in a traditional fishery rather than a supermarket.

Source : DGCCRF, 2002.
For each of the four zones one nearby fishing port enabling direct provisioning was selected where beach fishing is possible. Individuals were recruited within a radius of 20 to 25 km around this point. Between 20 and 27 towns were visited in each zone and the number of individuals questioned per town was calculated on the basis of the number of inhabitants published in the 1999 INSEE survey. The four large regions finally selected are Le Havre in Normandie/Baie de Seine, Lorient in Bretagne sud, La Rochelle in Gironde/Charente Maritime sud, and Toulon in Mediterranee/Var. The existence in each of these zones of an environmental source of a contaminant of interest was also a selection criterion, notably PCBs in the Baie de Seine and cadmium in the Gironde estuary.

[^12]A representative consumer population sample was ensured by random recruitment (apart from the quotas applied) by door-to-door canvassing every five doors, using the so-called "random route" method.

About one thousand individuals were recruited, or 250 per zone. The people satisfying the inclusion criteria listed below were questioned.

## Inclusion criteria

- Adult (18 years or older),
- Consumes seafood at least twice a week, a criterion defined in the 1999 INCA study and the recommendations of the PNNS. The median consumption frequency calculated from the individual seafood consumption data in the population of the INCA study was twice a week (CREDOC-AFSSADGAL, 2000 ${ }^{85}$ ),
- Permanent resident in the zone for several years.


## Exclusion criteria

- Refusal to participate: during the pilot survey about $42 \%$ of the people contacted refused to participate in the survey ${ }^{86}$.

Over and above these inclusion criteria, the female population aged 18 to 44 years corresponding to women of child-bearing age was over-represented in order to obtain a sample representative of this population and a larger volume of consumption and biological data in view of the health risks associated with methylmercury. Based on an evaluation of the health risks of methylmercury consumed in seafood, in 2004 the AFSSA recommended that pregnant and breast-feeding women should consume no more than 150 g of predator fish per week, in addition to their usual consumption of non-predator fish ${ }^{32}$.

The dietary study covered all foods consumed by adults (18 years or older): seafood and other foods. This study is based on a questionnaire concerning consumption frequency validated during the pilot survey by means of 7 -day consumption diaries. The portion sizes usually consumed were estimated by means of a book of sample photographs ${ }^{87}$. The survey involved a single interview during which the following points were covered:

- Presentation of the study to obtain the consent of the adult person to participate;
- Questionnaire concerning the frequency of food consumption in general and seafood in particular. Information was also collected on purchasing methods (fresh, frozen, canned, etc.) and on the usual origin of the seafood consumed (commercial and self-provisioning), the socio-demographic profile of the respondent and about ten closed questions on the perception of the food risks associated with seafood;
- Presentation of the biological part to obtain the informed consent of the respondent to participate if eligible (see below the exclusion criteria for the biological part).

[^13]
## Exclusion criteria for the biological part

- Refusal to participate;
- People suffering from pathologies that could have repercussions on the biological level of omega 3, lead, mercury, arsenic or cadmium (renal disease, arterial hypertension, urinary incontinence).


### 1.2.2 Consumption survey

In order to have data on habitual consumptions, we gave priority to the food frequency questionnaire (FFQ). However, and in view of the fact that FFQs are less precise than questionnaires concerning short periods, such as consumption diaries or 24 h recall, we first performed a FFQ validation study ${ }^{86}$.

This pilot survey was intended to prepare for the full-scale survey and was performed in two coastal zones (La Rochelle on Atlantic and Toulon on Mediterranean). The field survey involved 61 people aged 15 or over. The consumptions were recorded using a consumption diary and a food frequency questionnaire. Owing to the heavy work implied by a consumption diary and its poor representativeness of dietary habits over the year, this solution was excluded for the full-scale survey. It was therefore necessary to validate the lighter and less restrictive food frequency questionnaire for the main surveys. The results of this survey reveal poor correlation (correlation coefficient ranging from -0.1 to 0.15 ) between the consumption of some seafood products and the two collection methods. Several reasons could explain this disagreement. First, the seasonality of the consumption of certain seafood products. Indeed, since the diary does not reflect the dietary habits over the year, a large number of fish in the list on the FFQ had not been consumed when filling in the diary. Moreover some confusion was observed for several categories of products:

- Seafood products consumed fresh but also canned and/or smoked: these were therefore well separated and well identified in the FFQ used in the full-scale survey to avoid risk of confusion.
- Products for which several denominations are sometimes used (cod, grenadier/hoki, etc) or which have local names: for these we tried to be as exhaustive as possible by indicating all the known names or by grouping the names designating the same product.

These two points led us to review the classification of products subject to confusion in the statistical analysis of the pilot survey data.

On the other hand, satisfactory correlations were observed for well identified seafood products (salmon, skate, perch and trout for which the correlation coefficient ranged from 0.35 to 0.5 ) and for the other food categories.

The FFQ was therefore validated and improvements were made prior to the full-scale survey.

The survey was carried out in all the selected zones between October and December 2004. In all, 6,379 people were contacted by door-to-door recruitment and 43\% agreed to participate. Of these 2,768 people, 1,757 (almost 66\%) failed to meet the inclusion criteria previously defined and were therefore ineligible:

- $24 \%$ said they do not eat any fish or seafood,
- 34\% did not consume seafood at least twice a week,
- 2\% did not reside permanently in the town concerned,
- 3\% were under 18 years old.

In the four zones a total of 1,011 interviews were carried out.
Tables 3 and 4 present the detailed results concerning the acceptability of the food survey part for each of the four zones.

Table 3 and 4: Participation in the study and distribution by zone

|  | Refusal | Agreement |
| :--- | :---: | :---: |
| Le Havre | 1,028 | 777 |
| Lorient | 950 | 742 |
| Lotal |  |  |
| La Rochelle | 804 | 564 |
| 1,805 |  |  |
| Toulon | 829 | 685 |
| Total | 3,611 | 2,768 |


|  | Eligible | Non eligible |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Interview | City | Age | Consumption | Consumption frequency | Other |  |
| Le Havre | 251 | 7 | 13 | 256 | 250 | - | 777 |
| Lorient | 249 | 14 | 31 | 229 | 219 | - | 742 |
| La Rochelle | 253 | 10 | 19 | 21 | 261 | - | 564 |
| Toulon | 258 | 33 | 26 | 153 | 214 | 1 | 685 |
| Total | 1,011 | 64 | 89 | 659 | 944 | 1 | 2,768 |

In order to assure consistency between the exploitation of the data and the national and/or international recommendations concerning omega 3 intake and exposure to contaminants, the population was divided into three categories and one sub-category (Table 5):

- Male adults: men aged 18 to 64 years,
- Female adults: women aged 18 to 64 years,
- In order to acquire information on the risks or benefits of fish consumption by women of child-bearing age, a sub-category was also defined: women from 18 to 44 years old.
- Elderly people: the population aged 65 or over without distinction of sex.

The data on consumption, nutritional intakes and exposure to contaminants were analysed for each of these population categories in each of the zones.

Table 5: Distribution of the respondents by survey region

| Category | Le Havre | Lorient | La Rochelle | Toulon | Total |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Adult men (aged 18-64) | 45 | 53 | 88 | 60 | 246 |
| Adult women (aged 18-64) | 180 | 159 | 125 | 177 | 641 |
| Women of childbearing age (aged 18-44) | 98 | 77 | 79 | 96 | 350 |
| Older subjects (aged 65 and more) | 26 | 37 | 40 | 21 | 124 |
| Total | 251 | 249 | 253 | $\mathbf{2 5 8}$ | $\mathbf{1 , 0 1 1}$ |

The female population is effectively over-represented (about 2.5 times more than men) and in particular women aged 18 to 44 , which enabled us to better exploit the data for this particular target population. The data for pregnant women were not exploited owing to the very small sample size ( $n=14$ ).

Tables 6 and 7 present the professional situations and the professional categories of the respondents who were still working at the time of the study. More than the half the respondents had a profession, the other half being inactive (unemployed, homemaker, invalid, etc.), retired or military. All social categories are represented.

Table 6: Professional situation of the respondents

| Present professional situation | Le Havre | Lorient | La Rochelle | Toulon | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Exercises a profession | 133 | 111 | 148 | 131 | 523 | 51.7\% |
| Student | 27 | 7 | 4 | 14 | 52 | 5.1\% |
| Job seeker | 1 | 3 | 2 | 5 | 11 | 1.1\% |
| Annuitant | 1 | 6 | 4 | 3 | 14 | 1.4\% |
| Retiree | 43 | 66 | 55 | 29 | 193 | 19.1\% |
| Long-duration illness or invalid | 3 | 6 | 6 | 10 | 25 | 2.5\% |
| Housewife, Homemaker | 31 | 38 | 12 | 47 | 128 | 12.7\% |
| Unemployed worker | 12 | 12 | 21 | 18 | 63 | 6.2\% |
| Military | 0 | 0 | 1 | 1 | 2 | 0.2\% |
| Total | 251 | 249 | 253 | 258 | 1011 | 100\% |

Table 7: Distribution of professional situation among active individuals

| Professionnal situation | Le Havre | Lorient | La Rochelle | Toulon | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Farmer | 0 | 1 | 0 | 0 | 1 | 0.2\% |
| Artisan, trader or contractor | 8 | 13 | 13 | 8 | 42 | 8.0\% |
| Manager or high intellectual profession | 8 | 13 | 15 | 13 | 49 | 9.4\% |
| Intermediate profession | 26 | 20 | 41 | 35 | 122 | 23.3\% |
| Employee | 68 | 42 | 54 | 53 | 217 | 41.5\% |
| Worker | 16 | 21 | 20 | 19 | 76 | 14.5\% |
| No reply | 7 | 1 | 5 | 3 | 16 | 3.1\% |
| Total | 133 | 111 | 148 | 131 | 523 | 100\% |

### 1.2.3 Biological part

This is a study of biomarkers of exposure to fatty acids and environmental contaminants, exploiting biological samples taken from consumers participating in the food consumption survey. Despite the general interest shown in the subject of this study, $52.8 \%$ of the people questioned did not participate in the biological part: $39.1 \%$ did not wish to, for the reasons shown below, $13.2 \%$ were not eligible for health reasons (see the exclusion criteria for the biological part) and $0.5 \%$ ( 5 people) did not reply (Table 8). The $55 \%$ acceptance rate among the eligible people is high for this type of public health study. Figure 4 shows the distribution of the reasons for refusal (several replies were possible; the results correspond to the aggregated replies).

Table 8: Participation in the biological part

|  | Agreement | Refusal | Non eligible | No reply |
| :--- | :---: | :---: | :---: | :---: |
| Le Havre | 103 | 109 | 38 | 1 |
| Lorient | 126 | 80 | 41 | 2 |
| La Rochelle | 119 | 106 | 28 | - |
| Toulal | 251 |  |  |  |
| Total | 129 | 100 | 27 | 2 |

Figure 4: Reasons for refusal to participate in the biological part


[^14]Of the eligible respondents 477 people agreed to participate in this biological part; 83 desisted and 394 were sampled (Table 9).

Table 9: Summary of sampling

|  | Agreement | Desistance | Sampling | among |  | Age |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Females | Males |  |  |
| Le Havre | 103 | 15 | 84 | 66 | 18 | $44 \pm 15 y$ |
| Lorient | 126 | 13 | 115 | 91 | 24 | $46 \pm 14 y$ |
| La Rochelle | 119 | 25 | 100 | 57 | 43 | $47 \pm 14 y$ |
| Toulon | 129 | 30 | 95 | 75 | 20 | $44 \pm 13 y$ |
| Total | 477 | 83 | 289 | 105 |  |  |

From October to December 2004, the samples were taken by selected medical laboratories in each of the four zones. They also performed the lipids profile for each individual. The samples were stored at $-20^{\circ} \mathrm{C}$ in suitable containers until their analysis. This storage temperature was maintained throughout the transportation of the samples to the analytical laboratories.

The biological samples were analysed for a profile of fatty acids in the erythrocyte membrane, and also for lead, mercury, arsenic and cadmium and the toxic forms of arsenic and mercury (inorganic arsenic and methylmercury).

The exposure of the subjects to fatty acids was evaluated by performing a fatty acids profile of the phospholipids in erythrocyte membranes, since with a 120 days half-life these phospholipids provide a good marker of the long-term dietary regime ${ }^{88}$, unlike adipose tissue or plasma which could only reflect the last meal. Plasma being constituted essentially by triglycerides, it does not provide a good marker for long-chain n-3 PUFAs intake. Similarly a profile based on adipose tissue takes into account only the latest variations of the fatty acid levels in the diet. The level of the erythrocyte membrane phospholipids integrates all these fluctuations over the preceding three months.

Concerning trace elements, total blood and urine constitute the most appropriate biomarkers for evaluating the levels of individuals chronic exposure 8990919293 .

## Analysis of the biological samples

Lipids profiles were performed using 4 ml blood samples:

- Lipids profile (total cholesterol, HDL cholesterol, LDL cholesterol and triglycerides) based on serum using the classic analysis methods of medical laboratories.
- Fatty acids profile on the pellet fraction of the erythrocyte membrane phospholipids, notably longchain PUFAs: EPA (eicosapentaenoic acid) and DHA (docosahexaenoic acid).

[^15]
## Fatty acids

The fatty acid profile of erythrocytes was performed on the blood collected in heparin tubes and immediately centrifuged ( $3,500 \mathrm{~g}, 10 \mathrm{~min} ., 4^{\circ} \mathrm{C}$ ) after elimination of the plasma. The erythrocytes were frozen at $-80^{\circ} \mathrm{C}$.

The erythrocyte lipids were first extracted ${ }^{94}$. After evaporation of the solvents, the total lipids were saponified then esterified. The methylic esters were extracted then separated and quantified by gas phase chromatography coupled to a flame ionisation detector. The identification of the methylic esters of fatty acids is based on the retention times obtained for standard methylic esters.

## Trace elements

The trace elements analyses were based on total blood and urine.
The total lead, mercury, cadmium and arsenic were determined by Inductively Coupled Plasma-Mass Spectrometry (ICP-MS), a sensitive, multi-element detection system. To ensure optimal reliability of the blood analysis results, the lead, cadmium and mercury were measured by two different laboratories on all the samples taken. The lead, cadmium and total arsenic were measured in the urine samples. The laboratories' detection limits for these trace elements are presented in Table 10.

The speciation of arsenic was performed on a hundred samples presenting the highest total arsenic concentration ( $>75 \mu \mathrm{~g} / \mathrm{g}$ creatinine). The inorganic arsenic, the most toxic forms, was measured after elimination of the AsB and AsC by extraction in liquid-liquid phase. The remaining arsenic, inorganic As, $\mathrm{As}(\mathrm{III})$ and $\mathrm{As}(\mathrm{V})$ and their intermediate mono- and dimethylated derivatives, MMA and DMA of degree $V$ was determined by graphite furnace atomic absorption spectroscopy using Zeeman effect background correction.

Table 10: Standards and analytical limits for trace elements ${ }^{95}$ :

| Trace element | Biological matrice | Limit of detection | Limit of quantification | Standard* |
| :---: | :---: | :---: | :---: | :---: |
| Lead (Pb) | Total blood | 0.2-0.3 $\mu \mathrm{g} / \mathrm{L}$ | 0.7-10 $\mu \mathrm{g} / \mathrm{L}$ | $<90 \mu \mathrm{~g} / \mathrm{L}$ total blood (man) |
|  |  |  |  | $<70 \mu \mathrm{~g} / \mathrm{L}$ (woman) |
|  | Urine | $0.15 \mu \mathrm{~g} / \mathrm{L}$ | $0.5 \mu \mathrm{~g} / \mathrm{L}$ | $<25 \mu \mathrm{~g} / \mathrm{g}$ créatinine |
| Cadmium (Cd) | total blood | 0.03-0.3 $\mu \mathrm{g} / \mathrm{L}$ | 0.1-1 $\mu \mathrm{g} / \mathrm{L}$ | <2 $\mu \mathrm{g} / \mathrm{L}$ (smoker) |
|  |  |  |  | $<1 \mu \mathrm{~g} / \mathrm{L}$ (no smoker) |
|  | Urine | $0.1 \mu \mathrm{~g} / \mathrm{L}$ | $0.5 \mu \mathrm{~g} / \mathrm{L}$ | $<2 \mu \mathrm{~g} / \mathrm{g}$ créatinine |
| Mercury (Hg) | Total blood | $0.3 \mu \mathrm{~g} / \mathrm{L}$ | $1.0 \mu \mathrm{~g} / \mathrm{L}$ | <10 $\mu \mathrm{g} / \mathrm{L}$ |
| Methylmercury (MeHg) | Total blood | $0.3 \mu \mathrm{~g} / \mathrm{L}$ | $1.0 \mu \mathrm{~g} / \mathrm{L}$ |  |
| Total Arsenic (As) | Urine | $1.3 \mu \mathrm{~g} / \mathrm{L}$ | $5 \mu \mathrm{~g} / \mathrm{L}$ |  |
| Arsenic forms** | Urine |  | $10 \mu \mathrm{~g} / \mathrm{L}$ | <10 $\mu \mathrm{g} / \mathrm{g}$ créatinine |
| * correspond to the P95 for the general French population which is not professionally exposed <br> ** $\operatorname{As}(\mathrm{III})$, $\mathrm{As}(\mathrm{V}), \mathrm{MMA}(\mathrm{V})$ and DMA (V), considered as more toxic |  |  |  |  |

[^16]
### 1.2.4 "Total Diet Study" part

This part of the study is intended to estimate the intakes of nutriments and environmental contaminants of the adult population studied by means of a local total diet study (TDS) covering consumed fish and seafood. This study is based on the methodology developed in the total diet study of the French population ${ }^{3}$. It consists in sampling the fish and seafood mainly consumed by the population studied, taking into account the form of purchase (fresh, frozen, canned, etc.) and provisioning (bought or self-procured) and seasonal effects (consumption and contamination). However in this study seasonal effects were not taken into account in view of the fact that all the sampling was performed between January and April 2005. In a traditional total diet study foods are analysed "as-consumed" (with the exception of conserves), in other words precooked, cooked, raw, etc. In the present case, the foods were not cooked and the analyses were made on raw samples conserved from one to four weeks at $-20^{\circ} \mathrm{C}$, which does not modify the fatty acids composition ${ }^{96}$ or the content of trace elements or persistent organic pollutants.

## Preparation of the list of food samples

The list of sampled food is based on an analysis of the individual dietary consumptions of the respondents. It is composed of the seafoods (fish, molluscs, crustaceans, etc.) mainly consumed by the respondents. The list is designed to cover:

- Sea fish whose daily consumption per person is at least 1 g ;
- All predator fish, whatever their level of consumption: tuna, ray, ling, seabass, seabream, grenadier, halibut, eel, anglerfish, catshark, swordfish, emperor;
- Crustaceans and molluscs whose average daily consumption per person is at least 1 g ;
- All other canned or smoked seafood or seafood-based products, whatever their level of consumption.

The final list included 138 fresh and frozen products ( 32 for Le Havre, 38 for Lorient, 35 for La Rochelle, 33 for Toulon), plus 21 canned products, smoked products or prepared seafood-based dishes mentioned in the food frequency questionnaire, making 159 products in all (Appendix 1).

As shown in Table 11, these cover $88 \%$ to $100 \%$ of the total consumption of fish and seafood.
Table 11: Coverage by sampling of the total fish and seafood consumption in each region

| Site | Fish | Mollusc, crustacean | Other seafood * |
| :--- | :---: | :---: | :---: |
| Le Havre | $89.2 \%$ | $89.7 \%$ | $100 \%$ |
| Lorient | $96.1 \%$ | $89.2 \%$ | $100 \%$ |
| La Rochelle | $89.0 \%$ | $90.9 \%$ | $100 \%$ |
| Toulon | $93.0 \%$ | $88.1 \%$ | $100 \%$ |
| *canned food, smoked fish and seafood-based dish |  |  |  |

## Sampling

The following parameters were taken into account:

- quantities consumed,
- consumption frequencies,
- purchase methods (fresh, semi-fresh, frozen, canned, etc.),
- provisioning place (beach fishing, purchase on the fish dock, at the market, from a fish merchant, in another type of shop, or consumption outside the home),
- product origins (preferentially local, regional, etc.).

The sampling has not taken into account the proportion of wild fish and farmed fish insofar as this information did not appear in the food frequency questionnaire.

A sample of about $1,000 \mathrm{~g}$ was taken for each fresh product, allowing five 200 g sub-samples. The origin and distribution of these sub-samples were determined according to the place of purchase, based on the purchase frequency data of the consumption survey, which were weighted by the consumption frequency and quantity consumed. Figure 5 presents the manner in which the samples were constituted. For example, if $10 \%$ to $29 \%$ of a fish consumed is bought in a supermarket, a sub-sample of the composite sample comes from a supermarket. If $30 \%$ to $49 \%$ of the same fish is bought from a fishmonger, two sub-samples of the composite sample come from fishmongers, and so on. The purchase form (percentage of purchases of fresh and/or frozen products) was taken into account in the number of samples corresponding to purchases made in supermarkets.

Figure 5: Composition of a food composite sample according to the place of purchase (Example of seabass)


The samples of canned and smoked products and prepared seafood dishes were composed taking into account the market shares of the different brands, based on the purchasing data of households in the Secodip panel (2001). These are not composed of five sub-samples, as for the fresh products, but of $x$ samples of different brands covering the market shares as widely as possible, as presented in Table 12 for a product shared between 5 different brands.

Table 12: Illustration of the representative sampling for a product

| Brand | Found on the market | Market share |
| :--- | :---: | :---: |
| A | Yes | $40 \%$ |
| B | Yes | $30 \%$ |
| C | No | $10 \%$ |
| D | Yes | $10 \%$ |
| E | No | $10 \%$ |
| Total | $100 \%$ | $-37.5 \%$ |

Taking into account the effective presence of products on the market, the market coverage was $37 \%$ to $80.4 \%$ for canned products, $42.9 \%$ to $91.5 \%$ for smoked products and $50.3 \%$ to $72.3 \%$ for other products (prepared seafood-based dishes).

For each product in the list, the 5 sub-samples were mixed, ground and remixed to obtain a single homogeneous composite sample of the product. The sub-samples were composed only of the comestible parts of the products. More precisely, fish were filleted and skinned (notably the smoked fish). The canned foods were drained, particularly when they contained oil; for shellfish only the soft content was ground (plus the coral in the case of scallops); crustaceans were peeled in order to sample only the flesh (notably legs and claws of crabs and lobsters); mollusc and crustacean samples were composed of raw and/or cooked sub-samples.

The use of intermediate recipients was not allowed during the preparation of the samples. The mixing equipment used was made of stainless steel (K55 Dito Cutter/Mixer). In compliance with good laboratory practices, the hardware used to prepare the composite samples was thoroughly washed (RBS. 25 detergent) between each preparation in order to avoid cross-contamination between samples.

After preparation, the samples were stored at $-20^{\circ} \mathrm{C}$ in suitable containers until the time of analysis. This storage temperature was maintained throughout the transportation of the samples to the analytical laboratories.

## Analysis of the food samples

The analyses involved both nutritional and toxic elements in the products sampled. Concerning the toxic elements, the measurements concerned total lead, cadmium, arsenic and mercury, the various forms of speciation of arsenic ( $\mathrm{As}(\mathrm{III}), \mathrm{As}(\mathrm{V})$, MMA, DMA, AsB), of mercury (methylmercury) and of organostannic compounds (monobutyltin (MBT), dibutyltin (DBT), tributyltin (TBT), monophenyltin (MPT), diphenyltin (DPT), triphenyltin (TPT), monooctyltin (MOT), dioctyltin (DOT) and trioctyltin (TOT)), and persistent organic pollutants (POPs): 17 dioxins and furans, 12 dioxin-like PCBs, 7 indicator PCBs and 7 PBDEs, most of them found ( $28,47,99,100,153,154,183$ ).

The same samples were also analysed for saturated, monounsaturated and polyunsaturated fatty acids (48 fatty acids in all).

Table 13 shows the detection limits (LOD) and quantification limits (LOQ) of the various analyses.

## Fatty acids

The principle consists in extracting, purifying and esterifying the free fat in the samples according to the AFNOR NF V04-403 standard. After drying the sample, the fat is extracted, filtered through a column, then weighed and esterified. The esters are analysed using a gas phase chromatograph equipped with a flame ionisation detector. The identification of the esters in fatty acids is based on the retention times obtained for standard esters. The concentrations, calculated relative to a standard, are based on the areas of the corresponding peaks.

## Trace elements

The trace elements in the food matrices were measured by ICP-MS. Sixteen replicas were analysed. The speciation of mercury (methylmercury) was performed by coupling gas phase chromatography and ICPMS, and that of arsenic (AsB, MMA, DMA, As(III) and As(V)) by coupling liquid phase chromatography and ICP-MS. The compounds MMA and DMA detected in the tissue of fish and seafood are of redox potential V . The organostannic compounds were analysed by gas phase chromatography coupled with a microwave-induced plasma and an atomic emission detector.

## Persistent organic pollutants

The samples are first lyophilised then ground. For the analysis of the dioxin congeners, furans, DL-PCBs, i-PCBs and PBDEs, markers preimplanted with ${ }^{13} \mathrm{C}$ are added before extraction. The fat is then extracted by accelerated solvent extraction (ASE) using a toluene/acetone mixture under high pressure and temperature $\left(\mathrm{P}=100 \text { bar, } \mathrm{T}=120^{\circ} \mathrm{C}\right)^{97}$. The solvents are evaporated in order to determine the quantity of fat extracted. The extract is finally purified in three successive open chromatographic columns.

After these fat extraction and purification steps, a quantification standard is added in order to evaluate the recovery yields.

The four fractions obtained corresponding to each of the pollutant classes are analysed by gas phase chromatography coupled to a high-resolution mass spectrometer (GC-HRMS).

Table 13: Analytical limits for fatty acids, trace elements and persistent organic pollutants in food samples

|  | Limit of detection | Limit of quantification |
| :---: | :---: | :---: |
| Fatty acids (mg/g lipid) |  | Limitof 1 |
| Trace elements ( $\mu \mathrm{g} / \mathrm{g}$ fresh weight) |  |  |
| Lead (Pb) | 0.0004 | 0.001 |
| Cadmium (Cd) | 0.0004 | 0.001 |
| Mercury ( Hg ) | 0.0008 | 0.003 |
| MeHg | 0.0007 | 0.002 |
| Arsenic (As) | 0.002 | 0.005 |
| AsB | 0.005 | 0.002 |
| As(III) | 0.002 | 0.008 |
| DMA | 0.002 | 0.008 |
| MMA | 0.007 | 0.02 |
| As(V) | 0.01 | 0.03 |
| Organotin |  |  |
| MBT | 0.0002 | 0.0005 |
| DBT | 0.0001 | 0.0004 |
| TBT | 0.00008 | 0.0003 |
| MPT | 0.0002 | 0.0008 |
| DPT | 0.0002 | 0.0005 |
| TPT | 0.0002 | 0.0005 |
| MOT | 0.0002 | 0.0005 |
| DOT | 0.0002 | 0.0008 |
| TOT | 0.0002 | 0.0008 |
| Persistent organic pollutants (pgTEQ $\mathrm{who}^{\text {/g fresh weight - } \mathrm{ng} / \mathrm{g} \mathrm{fw} \text { ) }}$ |  |  |
| PCDD |  | - |
| PCDF | - | - |
| PCB-DL | - | - |
| iPCB | 0.001 | 0.001 |
| PBDE | 0.001 | 0.001 |

### 1.3 Presentation and interpretation of the results

The results are described in five parts. The first treats the results of the fish and seafood consumption survey, the second the composition in nutriments and contamination in contaminants of these same products, the third the nutritional intakes and exposure to contaminants. Two methodological approaches of evaluation of nutriment intakes and exposure to contaminants are presented, first a so-called "direct" or "biomarkers of exposure" approach, then an "indirect" or "food exposure" approach. The fourth part presents the perception of risks by participating consumers, and the fifth part presents some general remarks for discussion.

In the sections entitled "Seafood composition and contamination" and "Nutritional intakes and exposure to contaminants" the following substances are treated successively: fatty acids, trace elements and persistent organic pollutants.

### 1.3.1 Estimation of consumptions and concentrations

Concerning fish and seafood consumption, the results are presented in the form of three national tables, the first one for fish, the second for molluscs and crustaceans, and the last one for other seafood (canned, smoked, prepared dishes). Each table describes the average quantities consumed, the 95 th percentile of consumption and the consumption rate (the \% of consumers consuming a given product), for each product and each group of individuals (male adults, female adults, women of child-bearing age and people over 65 years old).

Concerning the composition of the products, and as in the "consumption" part of the study the results are presented in the form of three national tables: one for each category of foods. Each table describes the average quantities of each element analysed in the food samples listed in alphabetic order. The total lipids (total and unsaponifiable fatty acids) expressed in $\mathrm{g} / 100 \mathrm{~g}$ corresponds to an average of two analyses on the same samples, but made two different laboratories using two methods (NF VO4-403 standard and accelerated extraction by solvents). The unsaponifiable parts not having been analysed, the levels were deduced from the literature when possible (USDA database for the common species). For fish not covered by the literature, the unsaponifiable levels were estimated as recommended by the FAO and applied by the CIQUAL ${ }^{98}$. For molluscs, crustaceans and other products the average unsaponifiable levels were estimated from values found in the literature.

The consumption data are expressed in $\mathrm{g} /$ week, the fatty acid levels in $\mathrm{mg} / 100 \mathrm{~g}$ of food, the trace elements in $\mu \mathrm{g} / \mathrm{g}$ of food, and the persistent organic pollutants in $\mathrm{pg} \mathrm{TEQ}_{\mathrm{WHO}} / \mathrm{g}$ of food for dioxins, furans and DL-PCBs and in $\mathrm{ng} / \mathrm{g}$ of food for $\mathrm{i}-\mathrm{PCBs}$ and PBDEs.

### 1.3.2 Estimation of missing or censured data

Missing composition data for the products not sampled in the study zones were completed on a case-by-case basis. In some cases, a data item missing in one zone was replaced by the average of the data available for the other zones. In others cases a product was not sampled at all (not found on the markets at the time of sampling or very little consumed) so no measurement could be made ( 7 fish and 12 molluscs or crustaceans). The fatty acid level was then estimated either by exploiting information available in the national and/or international literature or by applying the calculated average for products of the same family, class or order.

The trace elements that were undetectable (<LOD) or unquantified (<LOQ) were taken to be equal to half these limits, in line with international recommendations ${ }^{99}$. For the persistent organic pollutants, since it is generally accepted that at high resolution (GC-HRMS) the LOD is equal to the LOQ, half the LOD was taken for undetectable concentrations. For the censured fatty acid measurements, the levels were not taken into consideration.

### 1.3.3 Estimation of fatty acids and contaminants intakes

## Food exposure (Indirect approach)

The average intakes of fatty acids and contaminants were calculated by crossing the individual consumption data from the food consumption survey with the individual composition and/or contamination data obtained by analysis of the representative food samples in the consumption/provisioning sets selected in each study zone. The estimation of these intakes takes into account a coverage exceeding $90 \%$ of the individual consumptions of fish and seafood declared by the populations studied in each of the four zones (Table 11).

The fatty acid intakes via fish and seafood are expressed in mg/day; the trace element intakes are expressed in $\mu \mathrm{g} / \mathrm{kg} \mathrm{bw} /$ week; the persistent organic pollutants intakes are expressed in $\mathrm{pg} \mathrm{TEQ}_{\mathrm{who}} / \mathrm{kg}$ bw/week for dioxins, furans and DL-PCBs or in ng/kg bw/day for i-PCBs and PBDEs.

[^17]These results are presented in the form of a table for each region showing the average quantities and high quantiles (P95) of fatty acids and contaminants contained in food for the four groups of individuals (mean $\pm$ standard deviation, or SD). The exposure value for the high quantiles does not correspond to the sum of the high quantiles of exposure of each group of foods taken into account since the high consumers associated with each group are not the same. We should remember that the exposure is calculated for each individual on the basis of his/her declared real body weight. Eighteen individuals out of the 1,011 respondents did not state their body weight, so this was taken to be the average weight of individuals of the same sex and age group.

Moreover, when necessary the text indicates the main vector(s) contributing to the total exposure and/or toxicological reference value (TRV), expressed here as a poucentage, following the methodology of fixing food standards as recommended by the Codex Alimentarius Committee on Food Additives and Contaminants ${ }^{100}$.

## biomarker of exposure (direct approach)

One table for each region describes for each group of individuals the average and high-quantile (P95) results of the levels of biological exposure of the 394 subjects of the study, for fatty acids measured in the erythrocytes, and for trace elements in urine and blood.

Fatty acid and trace element levels that were undetectable (<LOD) or unquantified (<LOQ) were taken to be equal to the half these limits. The composition of the fatty acid profile of the erythrocytes is expressed as a percentage of the total lipids. The trace element concentration is expressed in $\mu \mathrm{g} / \mathrm{L}$ for blood or in $\mu \mathrm{g} / \mathrm{L}$ or $\mu \mathrm{g} / \mathrm{g}$ creatinine for urine.

## Characterisation of risks/benefits

This final stage of the risk evaluation aims to describe the breakdowns of intakes or biomarker and to compare the average level, the P95 and even the P97.5 of the exposed population with reference nutritional or toxicological values established by national, European or international expert scientific committees (AFSSA, EFSA, JECFA). The results obtained are expressed as equivalents or as contribution to the reference values.

Unlike the food exposure approach which in this study concerned only the contribution of fish and seafood products but not other foods, the biomarker approach enables us to characterise the real levels of omega 3 and contaminant biological exposure of populations. In theory it therefore offers a suitable risk/benefit approach for fish and seafood consumption. However a prerequisite is the availability of a common physiological target for which beneficial and adverse effects have been described and linked to an effect or state of health of a population studied. The comparison of these two information sources is indispensable for an objective analysis in this type of approach. In reality, from a methodological point of view the scientific approach is complex and necessitates further in-depth work later.

At this stage, only a descriptive analysis of the benefits of daily consumption of omega 3 as regards cardiovascular diseases and an analysis of the risks of daily intake of methylmercury was performed, based on existing recommendations and published epidemiological data, but without correlating the impact on the health of our population. To do this, we start from the assumption that beneficial nutritional
effects may appear in individuals whose omega 3 intakes conform to the recommendations and that, inversely, toxicological effects may appear in individuals whose exposures exceed the reference toxicological values. A quantification of these risk/benefit effects in relation to fish and seafood consumption is thereby obtained empirically simply by estimating the number of individuals that match the nutritional recommendations and those that exceed the reference toxicological values.

Moreover, to better explain or explore any differences in dietary habits, in the composition of the provisionings, in the estimated intakes of nutriments and contaminants and in biological levels, a statistical analysis was performed between the regions. Additionally, correlations were calculated between the different variables, fatty acid intakes, exposure to contaminants evaluated by the indirect approach and the direct approach and also using socio-demographic variables using SAS. 7 in order to validate the selected exposure markers from a methodological point of view.

## Interpretation of the results

It is necessary here to make some important remarks regarding the interpretation of the results:

- The use of the average concentration of composition or contamination in the indirect approach represents a realistic and appropriate estimation of the long-term omega 3 intakes and exposure to contaminants from fish and seafood products in that these data are based on representative sets of consumptions and provisionings of the populations studied in the four zones, and also in that these data are compared with reference lifelong nutritional and toxicological levels established by European or international scientific committees ${ }^{101}$.
- Toxicologists generally agree that as regards chronic toxic effects, occasional exposure exceeding the reference toxicological values for short periods of the life of an individual does not necessarily induce a significant human health risk due to the fact that the toxicological reference values (TRV) contain a safety margin ${ }^{102}$.
- Nutritionists generally agree that omega 3 fatty acids have a beneficial effect on prevention of CVDs and development of the cerebral nervous system. Nevertheless at present the respective involvement of eicosapentaenoic acid (EPA) and docosahexaenoic acid (DHA) LC-PUFAs is debatable. In France the ratio of the precursors of omega 6 and omega 3, linoleic acid (LA) / alpha-linolenic acid (ALA), is recommended to tend to 5 in adults. The recommended daily intakes vary depending on scientific authority. In France the recommended daily intakes are 2 g ALA and $500 \mathrm{mg} n-3$ and $n-6$ LC-PUFAs including 120 mg DHA ${ }^{15}$. At international level, the International Society for the Study of Fatty Acids and Lipids (ISSFAL) recommends a minimal daily intake of 500 mg n-3 LC-PUFAs (EPA and DHA) concerning the prevention of CVDs ${ }^{103}$.
- Epidemiologists generally agree that validated exposure biomarkers constitute good indicators of the total long-term dietary (or other) exposure of individuals and are consequently valid indicators for interpreting the level of toxicological risk or nutritional benefit to which consumers are exposed.

[^18]Moreover, it is important to bear in mind that this study is representative of the methodology applied and that there is some "background noise" in exposures of dietary origin of high fish and seafood consumers in the four coastal areas. Consequently, a priori it does not take into account special situations of overexposure, for example due to possible local environmental sources of contaminated food (e.g. consumers who do not respect interdiction of beach fishing) or to atypical consumer behaviour (e.g. intake of food supplements).

Finally, in view of our adoption of a composite food sampling approach (Total Diet Study method), one should bear in mind that the comparative statistics calculated on the composition or contamination of food and for which trends are observed, not significant differences, necessitate complementary investigations in view of the relative smallness of our sampling compared to statistical sampling of monitoring plan type.

# SECOND PART 

## Fish and seafood

## x. <br> consumption

## Seafood consumption by high consumers

On examination of the results relating to weekly seafood consumption, and despite our verifications of the completed questionnaires, some data appeared absurd. It therefore appeared reasonable to eliminate individuals consuming more than 5 kg of seafood per week (corresponding to a consumption of 714 g per day or 357 g per meal) and individuals consuming less than 200 g of seafood per week ( 29 g per day), since these are not representative of high seafood consumption. In all 15 individuals were excluded for this reason.

Consequently, the results concerning the seafood consumption are presented for 996 individuals.

For the questions relating to provisioning habits or perception of food risks, the calculations include the entire sample of 1,011 individuals.

### 2.1.1 Fresh and frozen fish

## National level (4 zones)

We were able to verify that the selected consumers were indeed high seafood consumers. The average consumption of fresh and frozen fish (excluding conserved or smoked products) was found to be $633 \mathrm{~g} /$ week, or $90 \mathrm{~g} /$ day in men aged 18 to 64 years with a 95th percentile of $1,491 \mathrm{~g}$ (Table 14). For women in the same age group the average is $637 \mathrm{~g} /$ week, or $91 \mathrm{~g} / \mathrm{day}$, with a 95 th percentile of $1,522 \mathrm{~g} /$ week. Finally, the average consumption of subjects over 65 years old is $788 \mathrm{~g} / \mathrm{week}$, or $112 \mathrm{~g} / \mathrm{day}$, with a 95 th percentile of $1,783 \mathrm{~g} /$ week.

In all the zones and all the population groups (male adults, female adults and elderly people) cod is the fish most consumed, not only in terms of quantity but also in percentage of consumers: the average quantity consumed is about $93 \mathrm{~g} /$ week, and the consumer rate is between $81 \%$ and $88 \%$. Women of child-bearing age conform to the same trend as female adults, but with an even higher consumer rate.

Salmon, saithe and sole are also among the most consumed fish. Women consume more salmon and saithe ( 67 and 59 g/week) than men ( 56 and $54 \mathrm{~g} /$ week) and elderly people ( 57 and $50 \mathrm{~g} / \mathrm{week}$ ). On the other hand, for the latter group sole is the second most consumed fish ( $63 \mathrm{~g} / \mathrm{week}$ ) while it is only fourth in the other groups ( $50 \mathrm{~g} /$ week for men and $35 \mathrm{~g} /$ week for women).

Apart from these four fish, the distribution of the consumed species differs between groups. For example, elderly people consume much more herring than the other groups ( $32 \mathrm{~g} /$ week versus 8 g and $9 \mathrm{~g} /$ week for male and female adults, respectively). Elderly people also consume more ray ( $42 \mathrm{~g} / \mathrm{week}$ versus 25 and 26 g/week for under-65 men and women, respectively), with a consumers rate of around 56\%.

## Regional level

The regional differences are presented in Figure 6. Generally, male adults in Le Havre consume significantly less fish than those in Lorient ( $p<0.01$ ) and La Rochelle ( $p<0.05$ ). The average fish consumption in Lorient is also higher than that in Toulon for this same age group ( $p<0.05$ ). In the other groups, no difference in fish consumption was observed between the different zones.
Table 14: Consumptions of fresh and frozen fish by high consumers ( $\mathrm{g} / \mathrm{week}$ )


Figure 6: Mean consumptions of fish per zone, age group and sex (g/week)


### 2.1.2 Molluscs and crustaceans

## National level (4 zones)

The average consumption of molluscs and crustaceans is 270 g per week for men aged 18 to 64 , with a 95th percentile of $703 \mathrm{~g} ; 260 \mathrm{~g}$ for women in the same age group, with a 95th percentile of 665 g ; 279 g for people over 65 , with a 95 th percentile of 649 g (Table 15).

In male adults, the highest average consumption is of oysters, shrimps and great scallops, with respectively $41 \mathrm{~g}, 36 \mathrm{~g}$ and 34 g per week. Shrimps have the highest consumers rate of the three ( $92 \%$ versus $70 \%$ for great scallops and $67 \%$ for oysters). Mussels are also very widely consumed: $88 \%$ of all men aged 18 to 44 years say they consume on average 23 g per week.

Among women in the same age group, shrimps, great scallops, oysters and mussels are the most consumed seafood, with respectively $42 \mathrm{~g}, 40 \mathrm{~g}, 28 \mathrm{~g}$ and 22 g per week. Shrimps and mussels have the highest consumers rates ( $92 \%$ and $86 \%$ respectively).

Women consumers aged 18 to 44 years (i.e. women of child-bearing age) display the same trends both in terms of average consumption ( 235 g per week, with a 95th percentile of 607 g ) and similar of molluscs and crustaceans consumption profiles.

Among elderly people, oysters and great scallops ( 51 g and 43 g per week) stand out from other crustaceans and molluscs. Shrimps and mussels, consumed less than oysters ( 37 g and 24 g respectively) are consumed by more individuals ( $85 \%$ and $88 \%$ respectively).
Table 15: Consumptions of molluscs and crustaceans by high consumers (g/week)

| Molluscs, , rustaceans | Adult men (18-64 y) |  |  | Adult women (18-64 y) |  |  | Older subjects (65 y and more) |  |  | Women of childbearing age (18-44 y) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{n}=243$ |  |  | $\mathrm{n}=630$ |  |  | $\mathrm{n}=123$ |  |  | $\mathrm{n}=344$ |  |  |
|  | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** |
| Abalone | 0.60 | 0.00 | 1.20 | 0.30 | 0.00 | 1.10 | 1.70 | 0.00 | 1.60 | 0.40 | 0.00 | 1.20 |
| Calico scallop | 14.6 | 46.9 | 21.4 | 11.5 | 45.0 | 26.5 | 22.5 | 113 | 333 | 13.8 | 45.0 | 24.7 |
| Carpet shell | 3.50 | 20.0 | 26.3 | 2.20 | 10.0 | 21.4 | 2.90 | 12.0 | 33.3 | 1.90 | 10.0 | 18.3 |
| Clam | 0.30 | 0.00 | 4.50 | 0.20 | 0.00 | 4.00 | 0.20 | 0.00 | 1.60 | 0.10 | 0.00 | 4.10 |
| Codkle | 2.40 | 12.5 | 23.5 | 3.20 | 17.5 | 26.2 | 3.00 | 17.5 | 22.0 | 2.50 | 12.5 | 23.3 |
| Common periwinkle | 3.80 | 15.0 | 47.7 | 4.20 | 25.0 | 47.6 | 5.10 | 25.0 | 44.7 | 3.60 | 12.5 | 43.0 |
| Crab | 8.80 | 40.0 | 61.7 | 8.30 | 40.0 | 58.7 | 8.00 | 25.0 | 512 | 7.50 | 25.0 | 57.8 |
| Crayfish | 1.80 | 12.5 | 12.8 | 1.60 | 12.5 | 13.7 | 0.40 | 0.00 | 4.10 | 1.80 | 12.5 | 16.0 |
| Donax clam | 0.10 | 0.00 | 0.80 | 0.40 | 0.00 | 1.90 | 0.20 | 0.00 | 1.60 | 0.20 | 0.00 | 0.90 |
| Great scallop | 34.0 | 125 | 69.5 | 39.8 | 156 | 73.2 | 42.6 | 156 | 67.5 | 34.0 | 125 | 70.1 |
| Grooved sea squirt | 1.30 | 0.00 | 2.50 | 0.90 | 0.00 | 3.20 | 0.70 | 0.00 | 1.60 | 0.70 | 0.00 | 2.60 |
| Hard dam | 1.40 | 9.40 | 16.9 | 1.10 | 630 | 14.4 | 2.60 | 12.5 | 203 | 0.70 | 6.30 | 9.90 |
| Limpet | 0.10 | 0.00 | 0.40 | 0.30 | 0.00 | 1.40 | 0.40 | 0.00 | 0.80 | 0.30 | 0.00 | 1.20 |
| Lobster | 4.10 | 22.5 | 13.6 | 5.50 | 45.0 | 18.4 | 3.20 | 22.5 | 122 | 6.20 | 45.0 | 19.5 |
| Mussel | 23.1 | 70.0 | 88.1 | 21.6 | 70.0 | 85.6 | 24.1 | 70.0 | 87.8 | 21.2 | 70.0 | 84.0 |
| Octopus | 7.10 | 32.5 | 18.5 | 4.10 | 26.3 | 13.7 | 1.60 | 16.3 | 7.30 | 3.60 | 26.3 | 13.4 |
| Oyster | 40.9 | 144 | 67.1 | 27.9 | 90.0 | 61.4 | 51.3 | 144 | 71.5 | 23.8 | 90.0 | 55.5 |
| Queen scallop | 1.10 | 0.00 | 2.50 | 0.00 | 0.00 | 0.20 | 0.20 | 0.00 | 1.60 | 0.00 | 0.00 | 0.30 |
| Razor clam | 0.80 | 0.00 | 3.30 | 0.30 | 0.00 | 1.10 | 0.00 | 0.00 | 0.00 | 0.60 | 0.00 | 1.70 |
| Scampi | 19.4 | 90.0 | 54.3 | 17.4 | 75.0 | 52.7 | 25.9 | 113 | 58.5 | 14.4 | 60.0 | 49.4 |
| Sea urchin | 8.20 | 52.5 | 11.9 | 13.0 | 52.5 | 10.8 | 8.30 | 43.8 | 10.6 | 7.60 | 52.5 | 10.2 |
| Seiche | 9.90 | 50.0 | 27.2 | 6.70 | 32.5 | 19.8 | 5.90 | 32.5 | 17.9 | 6.40 | 32.5 | 20.1 |
| Shrimp | 36.0 | 100 | 91.8 | 41.5 | 125 | 91.7 | 37.1 | 100 | 84.6 | 43.0 | 140 | 92.7 |
| Slipper lobster | 0.00 | 0.00 | 2.10 | 0.00 | 0.00 | 0.60 | 0.00 | 0.00 | 0.80 | 0.00 | 0.00 | 0.60 |
| Spider crab | 4.50 | 20.0 | 29.2 | 4.30 | 20.0 | 29.0 | 4.00 | 25.0 | 31.7 | 2.70 | 10.0 | 24.4 |
| Spiny lobster | 1.30 | 12.5 | 13.6 | 1.60 | 630 | 17.0 | 1.50 | 6.30 | 17.9 | 1.90 | 12.5 | 17.4 |
| squid | 20.9 | 103 | 52.7 | 16.1 | 80.0 | 51.6 | 10.9 | 50.0 | 35.8 | 15.1 | 81.3 | 47.4 |
| swimcrab | 7.80 | 50.0 | 15.6 | 8.30 | 50.0 | 16.0 | 8.20 | 50.0 | 15.4 | 4.20 | 25.0 | 10.8 |
| Whelk | 12.6 | 62.5 | 35.4 | 17.3 | 93.8 | 33.3 | 6.70 | 37.5 | 19.5 | 17.0 | 93.8 | 33.4 |
| TOTAL | 270 | 703 | 98.4 | 260 | 665 | 99.7 | 279 | 649 | 100 | 235 | 607 | 99.4 |
| ** Consumers rate |  |  |  |  |  |  |  |  |  |  |  |  |

## Regional level

Concerning the geographic differences, the people in La Rochelle consume more molluscs and crustaceans (Figure 7). The average consumption of adult women in this town is significantly different from that in Le Havre ( $p<0.05$ ) and Toulon ( $p<0.01$ ). People in Lorient consume more crustaceans and molluscs than those in Toulon ( $p<0.05$ ). For men in the same age group, those in La Rochelle consume significantly more molluscs and crustaceans than those in Le Havre ( $p<0.05$ ). Among the elderly, people in Toulon consume significantly less molluscs and crustaceans than those in La Rochelle and Lorient ( $p<0.05$ ),

Figure 7: Mean consumptions of molluscs and crustaceans per zone, age group and sex (g/week)


### 2.1.3 Other seafood

## National level (4 zones)

This category of products includes canned and smoked fish and the other seafood-based products. The consumptions of these foods must be interpreted with care since the recipes of products such as fish soup and paella are very variable between individuals and commercial brands; their fish or crustacean content can differ. Table 16 includes the consumption of a few complete dishes without taking into account the recipe or the proportion of seafood they contain. Consequently these data do not reliably indicate the actual consumption of seafood and overestimate it

The weekly consumption of canned seafood is 125 g for men aged 18 to $64,102 \mathrm{~g}$ for women of the same age and 72 g for elderly people. Tuna is the most widely consumed canned product in all the groups, except for elderly people whose average consumption of anchovy is almost the same as that of tuna ( $20 \mathrm{~g} /$ week for anchovy, $22 \mathrm{~g} /$ week for tuna). While the average consumptions of sardine and mackerel are lower, these are consumed by large numbers of people.

The weekly consumptions of smoked products by men, women and the elderly are $22 \mathrm{~g}, 19 \mathrm{~g}$ and 13 g respectively. Salmon is the most-consumed smoked product with an average quantity of about $10 \mathrm{~g} / \mathrm{week}$ for adult male and female and $7 \mathrm{~g} /$ week for the elderly.

Paella and fish soup are widely consumed in all the population groups, though with a marked preference for soup among elderly people.
Table 16: Mean consumptions of canned food, smoked fish and seafood-based dish by high consumers (g/week)

| Other seafood | Adult men (18-64 y)$n=243$ |  |  | Adult women (18-64 y)$n=630$ |  |  | Older subjects ( 65 y and more)$n=123$ |  |  | Women of childbearing age (18-44 y)$\mathrm{n}=344$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** |
| Canned food |  |  |  |  |  |  |  |  |  |  |  |  |
| Anchovy | 31.7 | 180.0 | 39.1 | 22.8 | 120.0 | 36.2 | 19.5 | 93.8 | 35.8 | 16.9 | 75.0 | 28.8 |
| Crab | 4.7 | 22.5 | 25.5 | 6.4 | 22.5 | 33.2 | 2.5 | 15.0 | 27.6 | 8.6 | 37.5 | 32.0 |
| Mackerel | 17.6 | 65.6 | 59.3 | 14.1 | 60.0 | 54.9 | 10.0 | 37.5 | 54.5 | 16.3 | 60.0 | 55.8 |
| Pilchard | 2.8 | 15.0 | 10.3 | 1.4 | 7.5 | 10.2 | 1.0 | 7.5 | 73 | 1.3 | 7.5 | 7.8 |
| Sardine | 15.7 | 60.0 | 68.7 | 12.0 | 40.0 | 65.2 | 17.5 | 60.0 | 813 | 11.4 | 40.0 | 59.9 |
| Tuna* | 52.3 | 180.0 | 91.8 | 45.0 | 120.0 | 91.6 | 21.8 | 90.0 | 78.9 | 55.3 | 180.0 | 93.6 |
| Total canned food | 124.8 | 381.3 | 98.8 | 101.8 | 302.5 | 97.9 | 72.3 | 182.5 | 93.5 | 109.8 | 360.0 | 98.8 |
| Smoked fish |  |  |  |  |  |  |  |  |  |  |  |  |
| Haddock | 1.0 | 7.5 | 10.3 | 1.8 | 7.5 | 12.5 | 0.5 | 3.8 | 5.7 | 1.3 | 7.5 | 11.0 |
| Herring | 7.9 | 46.9 | 37.9 | 4.6 | 22.5 | 30.0 | 4.5 | 22.5 | 32.5 | 4.4 | 22.5 | 27.9 |
| Mackerel | 3.1 | 15.0 | 16.0 | 2.1 | 11.3 | 12.9 | 0.9 | 5.0 | 8.1 | 2.8 | 15.0 | 15.7 |
| Salmon | 10.1 | 37.5 | 78.6 | 10.3 | 37.5 | 81.3 | 7.1 | 25.0 | 63.4 | 10.3 | 40.0 | 81.1 |
| Total smoked fish | 22.1 | 75.0 | 86.8 | 18.8 | 70.0 | 85.7 | 13.0 | 40.0 | 79.7 | 18.8 | 66.3 | 85.8 |
| Seafood-based dish |  |  |  |  |  |  |  |  |  |  |  |  |
| Fish soup | 65.2 | 250.0 | 57.6 | 56.9 | 250.0 | 51.1 | 61.1 | 250.0 | 66.7 | 62.2 | 250.0 | 48.3 |
| Paella | 68.9 | 175.0 | 70.4 | 52.3 | 200.0 | 62.1 | 27.0 | 125.0 | 35.8 | 60.5 | 200.0 | 66.6 |
| Surimi | 26.3 | 70.0 | 70.8 | 35.7 | 140.0 | 77.0 | 11.6 | 43.8 | 46.3 | 40.9 | 140.0 | 81.4 |
| Tarama | 5.0 | 31.3 | 35.8 | 6.8 | 31.3 | 40.5 | 2.6 | 12.5 | 17.1 | 8.7 | 31.3 | 41.3 |
| Total seafood-based dish | 165.4 | 450.0 | 92.2 | 151.7 | 472.5 | 95.1 | 102.3 | 256.3 | 85.4 | 172.2 | 522.5 | 95.3 |
| Total | 312.3 | 798.8 | 99.6 | 272.2 | 742.5 | 99.5 | 187.7 | 472.5 | 99.2 | 300.8 | 795.0 | 99.1 |
| *Predatory fish as described in the Commission Regulation (EC) No 7822005 of 19 January $2005 \times *$ Consumers rate |  |  |  |  |  |  |  |  |  |  |  |  |

## Regional level

Significant differences between the average consumptions are observed between the study zones for all the groups studied. In particular the consumptions in Toulon are higher than in the other regions (Figure 8).

Figure 8: Mean consumptions of canned food, smoked fish and other seafood per zone, age group and sex (g/week)


### 2.2. Comparison of consumption data from the CALIPSO study and the Individual National Food Consumption Survey (INCA 99)

We compared the consumption data of fish, molluscs and crustaceans, smoked, canned or other products, taken from the INCA 99 survey and collected by means of a 7-day consumption diary with the results obtained in the present survey.

As often observed in this type of comparison, the values obtained by means of a food frequency questionnaire are greater than those obtained using a consumption diary, although we should remember that the CALIPSO survey concerned only high consumers, in other words those consuming seafood products twice a week or more, which tends to amplify the observed differences. We therefore decided to compare the results with the consumption data for seafood consumers alone.

The consumptions in the CALIPSO survey are about 2.5 times higher for consumption of fish, molluscs and crustaceans, and about 1.5 times higher for the other products. For the total consumption of these products, a factor of about 3.5 is observed between the two studies, which demonstrates that the CALIPSO study's objective of targeting high fish and seafood consumers was effectively reached (Appendix 3a).

### 2.3 Seafood provisioning

One question relative to the provisioning of fresh seafood was included in the survey for each product consumed. Several replies were possible, the results presented in the figures are expressed as a percentage of the replies.

In addition to the question on the provisioning method, when several methods were used the respondents were asked to decompose them (beach fishing, port, market, fishmonger, large and mid-size shops, consumption outside the home) in order to have a more detailed distribution of the provisioning methods for each seafood product consumed.

Only purchases of fish (fresh and frozen), molluscs and crustaceans are detailed below, since the other products (canned, smoked and seafood-based products) are procured only in retail shops.

The declared provisioning frequency at each place of purchase was weighted by the quantities consumed in order to calculate a distribution of the provisioning for each seafood product and for all consumers. For each seafood product $i$ and each individual $j$, the provisioning share of a given method $k$ (called AP) weighted by the consumption was calculated as follows:

$$
\mathbf{A P} \mathbf{P}_{\mathrm{ijk}}=\mathbf{P c t} t_{\mathrm{ijk}} \mathbf{X} \mathbf{C}_{\mathrm{ij}}
$$

where Pct is the purchase share of product i declared by the consumer j for the provisioning method k , and C is the consumed quantity of the product i consumed by the consumer j , all provisioning methods included.

For the whole population, the distribution of the provisioning method $k$ (called RP) was calculated as follows:

$$
R P_{i k}=\frac{\sum_{j} A P_{i j k}}{\sum_{j} C_{i j}} \times 100
$$

where C is the total quantity of the seafood product i consumed by the entire population, all provisioning methods included.

Two important points should be made here:

- Since the towns of Lorient and La Rochelle both have large covered fish markets, the replies "I buy at the market" and "I buy from a fishmonger" could possibly refer to the same supplier.
- Direct sales to the public are forbidden on the port in Le Havre and Toulon, so nearby markets have appeared where fishermen sell their products. Consequently, the replies "I buy at the market" and "I buy at the port" could refer to the same place of purchase in these two towns.

These two facts may explain the observed differences relative to the other zones.

### 2.3.1 Fresh and frozen fish

Figure 9 shows that, all species included, fresh and frozen fish is generally purchased in large stores (supermarket or hypermarket), in all the study zones, in particular in Le Havre and Hyères-Toulon where this provisioning method accounts for more than half of all purchases ( $61 \%$ and $54 \%$, respectively).

In Lorient and La Rochelle, purchases from fishmongers and the market account for almost half of all purchases, and large stores for one third.

Purchases on the port are mentioned by 10\% of respondents in Le Havre and Lorient, 4\% in La Rochelle and $3 \%$ in Hyères-Toulon. Very few people consume fish they have caught themselves; these are most numerous in the region of Lorient and La Rochelle (8\% and 6\% respectively).

Figure 9: Provisioning shares of fresh and frozen fish at different places of purchase in the 4 zones (\% of purchases)

LE HAVRE



LORAENT


Fished or collected
$\square$ Bought at the port
$\square$ Bought at the market


Bought from
a fishmonger


Bought from a supermarket
$\square$ Consumed only outside the home


LA ROCHEUE
Fished or collected


Bought at the market


Bought from
a fishmonger
a fishmonger


Bought from
a supermarket


Consumed only outside the home
$21.7 \%$


### 2.3.2 Molluscs and crustaceans

Molluscs and crustaceans are bought more often in large stores (supermarket, hypermarket) in Le Havre and Hyères-Toulon (Figure 10): half or more of all provisioning is by this method. In these regions about a quarter of all purchases are made at fishmongers. In the Mediterranean region, $14 \%$ of the respondents consume molluscs and crustaceans only outside their home.

In Lorient and La Rochelle almost 75\% of total provisioning is via the market, fishmongers and large stores, although the distribution varies between the regions.

Provisioning by beach fishing is much more prominent in Lorient and La Rochelle (15\% to 10\% compared to $3 \%$ to $5 \%$ in the other regions). However this remains a minority method.

Figure 10: Distribution of mollusc and crustacean provisioning in the 4 zones (\% of purchases)



LORIENT


Bought from
a fishmonger

Bought from a supermarket


Consumed only outside the home


# THIRD PART 

## Seafood composition

## and contamination

### 3.1 Fatty acid composition

As described in the first part, 159 fish and seafood products sampled in the four study zones were analysed for their fatty acids composition: 95 fish, 43 molluscs and crustaceans and 21 other products (canned, smoked and seafood-based products). Tables 17 to 19 present the results of the analysis of total lipids (in $\mathrm{g} / 100 \mathrm{~g}$ ) and fatty acids in these foods (in $\mathrm{mg} / 100 \mathrm{~g}$ ), averaged for the four regions, representing 30 species of fish, 17 species of molluscs and crustaceans and 14 other products.

### 3.1.1 Fresh and frozen fish

Total Lipids : The fish containing most lipids are eel, salmon, swordfish and halibut, in all the regions, with respective average levels of 20.4, 13.5, 12.4 and 11.7 g of lipids for 100 g of fish (Table 17). However the lipids profile of eel must be interpreted with care since the composite sample came from a single batch imported from the Netherlands. Mackerel and sardine are also rich in lipids ( 7.1 and $5.7 \mathrm{~g} / 100 \mathrm{~g}$ ) but unlike the fish mentioned previously the levels are not homogeneous between the different zones. The composite sample of sardine in Toulon is particularly low in fat ( $0.8 \mathrm{~g} / 100 \mathrm{~g}$ ), as is the composite sample of mackerel in La Rochelle ( $2.3 \mathrm{~g} / 100 \mathrm{~g}$ ) (results not presented). These differences may be explained by the difference of size observed during the sampling, a seasonal effect, different provisioning origins or a reproduction period dependent on the region. A regional effect may also explain the difference observed for the sardine sample in Toulon given that this comprised $80 \%$ of Mediterranean sardines (diet, etc.).

Angler fish, pout and cod display the lowest fat levels: 0.2 to 0.3 g of lipids $/ 100 \mathrm{~g}$.
These results are consistent with the data of the national CIQUAL database ${ }^{104}$.
LC-PUFAs : The fish richest in n-3 LC-PUFAs (EPA, DPA, DHA) are mainly the oiliest fish and also fresh anchovy ( $3,241 \mathrm{mg} / 100 \mathrm{~g}$ including $1,365 \mathrm{mg}$ DHA/ 100 g ). Logically, the fish the least rich in $\mathrm{n}-3 \mathrm{LC}$-PUFA are the least oily: gurnard, angler fish and pout ( 46,66 and $77 \mathrm{mg} / 100 \mathrm{~g}$ respectively), although some low-oil fish are also found to be rich in n-3 LC-PUFAs, for example cod, whiting and pollack for which respectively $55 \%, 52 \%$ and $51 \%$ of fatty acids are $n-3$ LC-PUFAs, which are relatively high levels (results not presented).

Data in the literature are absent or incomplete for two thirds of all fish ${ }^{104} 105106$. For the majority of products, our data (lipid level and lipids profile) are consistent with the literature. The differences when they exist may be explained by the large variability of the oil level in the fish flesh, this being dependent on the period of the year, age, size, sex, reproduction period, fishing zone or breeding method, the specific species, food and even particularly large individual variability. Our results for fresh tuna for example indicate a total lipids level of $0.73 \mathrm{~g} / 100 \mathrm{~g}$ versus $6.2 \mathrm{~g} / 100 \mathrm{~g}$ in the CIQUAL database, $15.5 \mathrm{~g} / 100 \mathrm{~g}$ in the German database and $4.90 \mathrm{~g} / 100 \mathrm{~g}$ in the American database. This may be explained by the provisioning period of the samples analysed (end of January to April) and by the provisioning methods: tuna sold in France at this time of year are "fattened" tuna weighing 10 to 35 kg , not necessarily the same as those analysed to constitute the French ${ }^{104}$, German ${ }^{105}$ and American ${ }^{106}$ databases.

[^19]
### 3.1.2 Molluscs and crustaceans

Total lipids : In general molluscs (Table 18) contain much less fat than fish ( 0.4 to 6.7 g of lipids/100 g). Crustaceans have a higher lipids level than molluscs with $6.7 \mathrm{~g} / 100 \mathrm{~g}$ for the common crab, $4.4 \mathrm{~g} / 100 \mathrm{~g}$ for the spider crab and $3.8 \mathrm{~g} / 100 \mathrm{~g}$ for the swimcrab.

LC-PUFAs : Crustaceans are also richer than molluscs in n-3 LC-PUFAs and more particularly in DHA: $714 \mathrm{mg} / 100 \mathrm{~g}$ for the common crab, for example. The molluscs with the least lipids are the cephalopods, in particular octopus ( $0.4 \mathrm{~g} / 100 \mathrm{~g}$ ) and cuttlefish ( $0.9 \mathrm{~g} / 100 \mathrm{~g}$ ).

There is little available composition data on this type of product in the French and international literature ( 9 products compared), but our results are consistent with the data in the CIQUAL' ${ }^{104}$ and German ${ }^{105}$ databases.

### 3.1.3 Other seafood

Total lipids : Concerning canned and smoked products and prepared seafood-based dishes, the lipid and fatty acid levels obtained are very variable (Table 19) despite our standardised preparation protocol of the comestible part, notably as regards draining. Mackerel (conserved in oil or smoked) has, as we might expect, the highest fat level ( 13.2 and 17.0 g of lipids $/ 100 \mathrm{~g}$ ).

LC-PUFAs : Once again the products the least rich in total lipids are not necessarily those the least rich in omega 3: of the total fatty acids, we find $20 \%$ of DHA in fish soup, canned pilchard and smoked salmon compared to $1.3 \%$ in tarama (which is very rich in total lipids - the results are not presented).

The results for these products must be interpreted with care. The total lipids level corresponds to the average of two analyses made on the same sample and the difference between the two levels could be as much as $21.4 \mathrm{~g} / 100 \mathrm{~g}$ for tarama or $14.7 \mathrm{~g} / 100 \mathrm{~g}$ for canned tuna. These differences, much smaller for fish, molluscs and crustaceans, can perhaps be explained by the difficulty of homogenising these products.

Tuna, sardine and mackerel contain two to eight times more fat when they are canned (including conservation in oil) compared to fresh or frozen forms. For sardine and tuna, this difference may be partially explained by the presence of oil in cans, despite draining during preparation of the samples or by different fishing periods for canned and fresh product. We always used preparation methods as close as possible to real consumption habits. Oleic (C18:1 n-9) and linolenic (C18:2 n-6) fatty acids, the most common in many oils, appear in higher concentrations in canned products than in fresh ones. Moreover, in the case of tuna the samples of fresh tuna were constituted only of red tuna whereas the samples of canned tuna (including the sub-samples of conserved in oil) are constituted of different species (yellowfin, albacore and skipjack). For mackerel, the differences in total lipids appear to be due to a concentration of arachidonic acid (C20:4 n-6) ten times higher in the composite sample of canned mackerel (including the sub-samples conserved in oil) than in fresh fish, and a concentration of parinaric acid (C18:4 n-3) four times higher in canned and smoked mackerel than in fresh mackerel. We are unable to explain these differences.

Smoked mackerel contains more fat (total lipids) than fresh mackerel. In this case the difference may be due to cooking prior to smoking which could concentrate the lipids. Smoked mackerel is much richer in n-3 PUFAs, whether as a precursor (ALA) or as long-chain derivatives. Smoked mackerel is also richer in saturated fatty acids (C18:0 and C20:0). These differences are perhaps due to the different fishing zones according to the final use of the product, direct consumption or transformation.

On the other hand, the compositions of fresh and smoked salmon are relatively close, a homogeneity that is perhaps explained by the fact that cultivated salmon accounts for $90 \%$ of salmon consumption in France, most of it coming from Norway ${ }^{107}$. In addition, the salmon is smoked cold $\left(25^{\circ} \mathrm{C}\right)$ without cooking, so there is no water loss and the fatty acids are not concentrated by the smoking.

### 3.1.4 Regional variations

There is no significant difference in the average compositions of fish in the different study zones in terms of the total lipidic composition or the level of LC-PUFAs omega 3. Similarly, there is no significant difference between the compositions of molluscs and crustaceans sampled in these zones.
Table 17：Mean fatty acid composition of fresh and frozen fish（ $\mathrm{mg} / 100 \mathrm{~g}$ ）

| Fish |  | $\begin{aligned} & \text { © } \\ & \frac{0}{5} \\ & \frac{0}{6} \\ & \frac{0}{3} \\ & \end{aligned}$ |  | $\stackrel{\stackrel{\ominus}{\dot{E}}}{\dot{U}}$ |  | $\stackrel{\circ}{i}$ |  |  | $\stackrel{\circ \circ \dot{\circ}}{\underset{\sim}{\circ}}$ |  |  |  |  | $\begin{aligned} & \text { ị } \\ & \underset{\sim}{\infty} \\ & \stackrel{\infty}{心} \end{aligned}$ |  |  | 웅 | Ni |  | $\begin{aligned} & \text { 区 } \\ & \text { 总 } \\ & \stackrel{1}{c} \\ & \text { ì } \\ & \text { d } \end{aligned}$ |  |  | $\stackrel{8}{4}$ | $\begin{aligned} & \stackrel{2}{\mathbf{L}} \\ & \stackrel{\rightharpoonup}{\Sigma} \end{aligned}$ | $\begin{aligned} & \text { 䘚 } \\ & \stackrel{\rightharpoonup}{2} \end{aligned}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchovy | 1 | 7.51 |  | 134 | － | － | 964 | 97 | 422 |  | 840 | － | 172 |  | 211 |  |  |  | 510 | 701 | 964 | 1，365 | 1，520 | 937 | 3，923 | 3，241 | 682 |
| Angler fish＊ | 4 | 021 |  | 8 | － | － | 32 | 6 | 12 | 1 | 18 | 2 | 1 |  | － | 1 |  |  | 6 | 26 | 2 | 37 | 52 | 33 | 71 | 66 | 7 |
| Catshark＊ | 4 | 0.55 | － | － | － | － | 71 | 5 | 20 | － | 42 | 1 | 2 | － |  | 1 |  | － | 22 | 113 | 15 | 66 | 93 | 70 | 219 | 195 | 23 |
| Cod | 4 | 0.30 | － | 1 | － | － | 45 | 2 | 10 | 2 | 19 | 2 | 1 | － | － | 5 | － | － | 7 | 28 | 4 | 75 | 57 | 24 | 121 | 112 | 8 |
| Common dab | 4 | 0.72 |  | 14 | － | 1 | 119 | 25 | 27 | 1 | 83 | － | 3 |  | 1 | 13 | 5 | － | 28 | 84 | 22 | 131 | 174 | 120 | 281 | 250 | 30 |
| Eel＊ | 1 | 20.4 | 56 | 675 | 112 | 31 | 3，525 | 1，759 | 847 | － | 7，379 | － | 618 | － | 151 | 296 | － |  | 666 | 432 | 286 | 716 | 5，583 | 9，591 | 3，205 | 1，880 | 1，285 |
| Emperor＊ | 3 | 5.78 | 78 | 26 | － | － | 541 | 412 | 77 | 30 | 1，616 | 106 | 132 | 3 | 6 | 667 | 8 |  | 65 | 471 | 53 | 742 | 847 | 2，212 | 2，141 | 1，940 | 201 |
| Plaice＊ | 2 | 037 | － | 10 | － | － | 60 | 20 | 12 | － | 22 | 9 | 1 | － | 1 | 2 | 4 |  | 14 | 46 | 6 | 41 | 97 | 53 | 111 | 97 | 15 |
| Goattish | 3 | 3.75 | － | 54 | 3 | 5 | 673 | 173 | 151 | － | 592 | － | 19 | 2 | 4 | 50 | 10 |  | 127 | 348 | 77 | 669 | 908 | 801 | 1，295 | 1，147 | 148 |
| Grenadier／hoki＊ | 4 | 0.44 | － | 4 | － | － | 50 | 6 | 16 | 1 | 45 | 2 | 3 | － | 2 | 30 | － | － | 4 | 41 | 8 | 78 | 74 | 66 | 168 | 160 | 8 |
| Gumard | 1 | 0.73 | － | 26 | － | － | 194 | － | 68 | 53 | 66 | － | 10 | － | － | － | 49 |  | － | 43 | － | 3 | 336 | 120 | 54 | 46 | 10 |
| Haddock | 2 | 025 | － |  | － | － | 44 | 1 | 11 | 4 | 12 | － | 2 | － | 2 |  | － |  | 8 | 18 | 5 | 60 | 60 | 18 | 94 | 84 | 10 |
| Hake | 4 | 0.59 | － | 7 | － | － | 89 | 13 | 21 | － | 73 | － | 3 | 1 | 1 | 9 | 11 |  | 10 | 28 | 19 | 123 | 130 | 87 | 193 | 180 | 14 |
| Halibut＊ | 4 | 11.7 | － | 366 | － | － | 1，832 | 808 | 295 | 109 | 1，544 | 329 | 75 | － | 9 | 1，296 | 143 | － 1 | 116 | 969 | 285 | 1，400 | 3，041 | 3，069 | 4，186 | 3，960 | 191 |
| John dory | 2 | 0.59 | － | 5 | － | － | 93 | 5 | 27 | － | 57 | － | 2 | － | 2 | 5 | 2 |  | 10 | 32 | 8 | 156 | 129 | 63 | 217 | 203 | 13 |
| Ling＊ | 4 | 0.33 | － | － | － | － | 47 | 4 | 16 | 1 | 22 | 3 | 2 | － | － | 2 | 1 |  | 3 | 45 | 1 | 65 | 65 | 45 | 117 | 112 | 4 |
| Mackerel | 4 | 7.07 | － | 179 | － | 10 | 1，198 | 145 | 327 | 15 | 1，258 | － | 145 | － | 58 | 343 | 120 | 1 | 114 | 662 | 118 | 1，404 | 1，867 | 1，436 | 2，845 | 2，585 | 259 |
| Pollack | 3 | 0.27 | － | － | － | － | 45 | 2 | 11 | 1 | 18 | 3 | 1 | － | － | 1 | － |  | 4 | 15 | 4 | 76 | 64 | 24 | 103 | 97 | 5 |
| Pout | 1 | 029 | － | 4 | － | 1 | 43 | 6 | 20 | 1 | 6 | － | 1 | － | 1 | 4 | 1 |  | 13 | 36 | 6 | 30 | 89 | 20 | 91 | 77 | 14 |
| Ray＊ | 4 | 0.61 | － | － | － | － | 105 | 2 | 20 | 1 | 71 | － | － | － | － | 7 | － | － | 23 | 17 | 15 | 156 | 129 | 77 | 216 | 195 | 23 |
| Saithe／coalfish | 4 | 1.04 | 17 | 10 | － | － | 144 | 5 | 32 | 1 | 162 | 5 | 120 | － | 2 | 9 | 2 | － | 7 | 71 | 7 | 173 | 206 | 174 | 391 | 262 | 126 |
| Salmon | 4 | 13.5 | － | 640 | 7 |  | 2，472 | 576 | 360 | 248 | 2，204 | 74 | 577 | 13 | 174 | 672 | 105 | － | 81 | 1，112 | 349 | 2，164 | 4，006 | 3，237 | 5，146 | 4，472 | 671 |
| Sardine | 4 | 5.72 | － | 179 | － | 11 | 1，200 | 197 | 276 | － | 808 | 3 | 37 | 4 | 40 | 146 | 42 |  | 89 | 638 | 178 | 1，269 | 1，79 | 1，081 | 2，407 | 2,270 | 130 |
| Scorpionfish | 1 | 227 | － | 52 | － | － | 342 | 72 | 51 | 25 | 265 | － | 13 | － | － | 205 | 16 | － | 45 | 121 | 57 | 507 | 464 | 399 | 949 | 890 | 57 |
| Seabass＊ | 4 | 2.99 | 2 | 44 | 1 | 3 | 502 | 124 | 154 | 3 | 380 | 27 | 47 | 4 | 12 | 40 | 7 |  | 72 | 357 | 65 | 617 | 732 | 561 | 1，221 | 1，090 | 125 |
| Sea bream＊ | 4 | 4.89 | － | 93 | － | － | 887 | 249 | 234 | 287 | 605 | 37 | 195 | 6 | 27 | 70 | 28 |  | 149 | 497 | 140 | 773 | 1，308 | 1，233 | 1，859 | 1，507 | 350 |
| Sole | 4 | 0.40 | － | 6 | － | － | 62 | 12 | 21 |  | 30 | 7 | 3 | － | 1 |  |  | － | 16 | 14 | 22 | 72 | 101 | 52 | 128 | 109 | 19 |
| Swordfish＊ | 4 | 12.4 | － | 241 | － | 30 | 2，035 | 488 | 651 | 37 | 2，108 | 469 | 70 | 24 | 39 | 367 | 191 | － 4 | 447 | 1，265 | 343 | 1，750 | 3，398 | 3，328 | 4，331 | 3，764 | 541 |
| Tuna＊ | 4 | 0.73 | － | 6 | － | 1 | 109 | 10 | 53 | － | 94 | 1 | 12 | 1 | 3 | 4 | 3 | － | 18 | 35 | 7 | 131 | 187 | 113 | 211 | 179 | 31 |
| Whiting | 4 | 025 | － | 1 | － | － | 37 | 2 | 9 | － | 22 | － | 1 | － | 1 | 6 | 1 | － | 3 | 15 | 3 | 69 | 53 | 29 | 98 | 93 | 5 |
| $-:<L O Q(1 \mathrm{mg} / \mathrm{g}$ lipid）；a ： Nb composite samples．Each sample is composed by 5 subsamples of the same species，representative of the provisioning methods in each zone（por MUFA，PUFA and LCPUFA indude all the analysed fatty adids，some of which are minority and are not presented in the table．＊Predatory fish as described in the Commission R |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 18: Mean fatty acid composition of molluscs and crustaceans ( $\mathbf{m g} / \mathbf{1 0 0} \mathrm{g}$ )

| Mollusc, crustacean |  | $\begin{aligned} & \text { ō } \\ & \frac{0}{0} \\ & \frac{0}{6} \\ & \frac{0}{3} \end{aligned}$ | $\begin{aligned} & \text { 벛 } \end{aligned}$ | $\stackrel{\stackrel{\ominus}{\dot{j}}}{\dot{~}}$ | $\frac{\stackrel{n}{\dot{E}}}{\stackrel{\dot{d}}{\dot{j}}}$ | $\stackrel{\circ}{\dot{H}}$ | $\begin{aligned} & \text { 압 } \\ & \hline \end{aligned}$ | $\stackrel{\grave{i}}{\stackrel{i}{6}}$ | $\stackrel{\circ}{\dot{\infty}}$ |  | $\begin{aligned} & \stackrel{9}{\dot{C}} \\ & \underset{\sim}{\ddot{0}} \\ & \underset{\sim}{6} \end{aligned}$ |  |  | $\begin{aligned} & \dot{L} \\ & \stackrel{L}{亡} \\ & \dot{\infty} \\ & \dot{\sim} \end{aligned}$ |  | $\begin{aligned} & \stackrel{m}{\dot{c}} \\ & \dot{\vdots} \\ & \dot{\oplus} \\ & \hline \end{aligned}$ | O멍 |  |  |  | $\text { ( } \forall d \mathrm{~d}) \varepsilon \text {-u } \varsigma: 乙 \supsetneq$ |  | $\stackrel{\stackrel{n}{4}}{\Delta}$ | $\stackrel{\sim}{\text { ® }}$ | $\stackrel{\stackrel{1}{4}}{\substack{2}}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calico scallop | 1 | 1.06 | - | 21 | 7 | 4 | 130 | 43 | 45 | 19 | 39 | 1 | 8 | 3 | 8 | 35 | 8 | 2 | 32 | 110 | 6 | 119 | 232 | 123 | 326 | 278 | 45 |
| Common periwinkle | 3 | 1.84 | - | 36 | 26 | 1 | 201 | 24 | 65 | 3 | 89 | - | 36 | - | 34 | 138 | 10 | - | 114 | 210 | 42 | 77 | 375 | 155 | 650 | 500 | 149 |
| Cockle | 2 | 0.40 | - | 4 | - | 1 | 51 | 8 | 27 | - | 15 | - | 1 | - | 1 | 19 | 6 | - | 15 | 49 | 6 | 33 | 102 | 28 | 123 | 108 | 16 |
| Crab | 3 | 6.66 | - | 46 | 5 | 18 | 525 | 619 | 54 | - | 915 | - | 15 | 19 | 7 | 113 | 18 | - | 327 | 1,160 | 41 | 714 | 706 | 1,552 | 2,404 | 2,035 | 360 |
| Cuttle fish | 2 | 0.85 | - | 11 | - | - | 134 | 2 | 35 | - | 18 | 2 | - | - | - | 27 | - | - | 22 | 69 | 9 | 156 | 193 | 33 | 282 | 261 | 22 |
| Great scallop | 4 | 0.80 | - | 15 | - | 4 | 128 | 12 | 33 | 4 | 21 | - | 4 | 3 | 4 | 19 | 10 | - | 13 | 105 | 9 | 105 | 201 | 48 | 263 | 242 | 20 |
| Lobster | 1 | 1.53 | - | 19 | - | 6 | 166 | 31 | 62 | - | 320 | 21 | 9 | - | 2 | 64 | 2 | - | 19 | 28 | 30 | 178 | 289 | 389 | 334 | 303 | 27 |
| Mussel | 4 | 1.09 | - | 15 | 1 | 4 | 158 | 49 | 29 | 3 | 25 | 11 | 9 | 1 | 10 | 33 | 12 | - | 32 | 162 | 12 | 151 | 240 | 102 | 410 | 368 | 42 |
| Octopus | 1 | 0.36 | - | 3 | - | - | 59 | - | 19 | - | 3 | 6 | - | - | - | 6 | - | - | 16 | 37 | 3 | 56 | 84 | 12 | 122 | 103 | 16 |
| Oyster | 4 | 0.55 | - | 12 | - | - | 99 | 12 | 16 | 10 | 38 | 5 | 6 | - | 8 | 22 | 15 | - | 17 | 82 | 4 | 64 | 153 | 67 | 208 | 180 | 24 |
| Scampi | 3 | 0.63 | - | 6 | - | 3 | 79 | 19 | 20 | - | 77 | 9 | 4 | 1 | 5 | 11 | 5 | - | 16 | 76 | 8 | 72 | 121 | 116 | 197 | 172 | 21 |
| Sea urchin | 1 | 0.81 | - | 62 | 2 | 5 | 104 | 17 | 20 | 9 | 14 | - | 3 | 4 | 6 | 42 | 2 | - | 68 | 98 | 6 | 6 | 219 | 63 | 234 | 158 | 75 |
| Shrimp | 4 | 0.76 | - | 3 | - | 4 | 69 | 48 | 61 | - | 81 | 7 | 62 | 1 | 3 | 4 | 1 | - | 21 | 71 | 6 | 66 | 162 | 149 | 235 | 150 | 84 |
| Spider crab | 1 | 4.39 | - | 80 | 13 | 39 | 376 | 209 | 160 | - | 566 | - | 18 | 31 | 14 | 338 | 28 | - | 154 | 536 | 39 | 294 | 779 | 807 | 1,445 | 1,221 | 202 |
| Squid | 4 | 1.33 | - | 14 | 1 | 3 | 186 | 8 | 46 | - | 77 | - | 136 | 1 | 1 | 23 | - | 1 | 12 | 68 | 31 | 167 | 255 | 103 | 440 | 290 | 150 |
| Swimcrab | 2 | 3.78 | - | 69 | - | - | 340 | 314 | 48 | - | 483 | - | 5 | - | 9 | 76 | - | - | 109 | 633 | 93 | 290 | 478 | 844 | 1,285 | 1,100 | 113 |
| Whelk | 3 | 0.88 | - | 10 | - | 2 | 95 | 12 | 54 | - | 51 | 18 | 8 | 1 | 3 | 29 | 12 | - | 25 | 88 | 37 | 71 | 201 | 102 | 262 | 228 | 34 |

Table 19：Mean fatty acid composition of other seafood（mg／100 g）

| Other seafood |  |  | 어̇ | $\stackrel{\circ}{\dot{f}}$ |  | $\stackrel{\circ}{\dot{H}}$ | نٌ | $\begin{gathered} \hat{E} \\ \stackrel{i}{U} \end{gathered}$ | OOO | $\begin{aligned} & \stackrel{n}{\tilde{y}} \\ & \stackrel{y}{0} \\ & \stackrel{0}{0} \end{aligned}$ |  |  |  |  |  |  | 음 |  |  |  |  | $\begin{aligned} & \text { T} \\ & \frac{1}{0} \\ & \text { ¢ } \\ & \dot{む} \\ & \text { did } \end{aligned}$ | 苍 | $\stackrel{\stackrel{\circ}{4}}{\underline{1}}$ | 发 | $\begin{aligned} & \text { M } \\ & \stackrel{c}{c} \\ & \text { 딘 } \\ & \text { 문 } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Canned food |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Anchovy | 2 | 7.36 | － | 101 |  | － | 1，294 | 156 | 860 |  | 1，145 | － | 814 | － | 127 | 54 | 3 |  | 106 | 466 | 238 | 886 | 2，259 | 1，301 | 2，694 | 1，771 | 919 |
| Crab | 1 | 0.57 |  | 3 |  | 3 | 59 | 16 | 42 |  | 43 | 13 | 7 |  | 3 | 1 | 1 |  | 33 | 73 | 4 | 77 | 122 | 80 | 199 | 159 | 40 |
| Maderel | 1 | 13.22 | － | 546 |  |  | 2，224 |  | 569 | 37 | 1，313 | 278 | 346 | 22 | 162 | 1，263 | － | 129 | 1，090 | 791 | 152 | 1，985 | 3，730 | 1，794 | 5，976 | 4，353 | 1，587 |
| Pilchard | 1 | 9.80 | － | 365 | 10 | 32 | 1，551 | 283 | 182 |  | 751 | 245 | 149 |  | 68 | 634 | － |  | 797 | 773 | 83 | 1，591 | 2,219 | 1，427 | 4，095 | 3，149 | 946 |
| Sardine | 1 | 11.85 | － | 340 | － | 30 | 2，322 | 365 | 522 | － | 2，579 | － | 590 | 16 | 53 | 338 | － |  | 206 | 1，108 | 156 | 1，999 | 3，328 | 3，093 | 4，484 | 3，653 | 813 |
| Tuna＊ | 5 | 6.03 | － | 56 | 1 | 4 | 404 | 44 | 717 | 8 | 1，683 | 21 | 106 | － | 88 | 4 | 1 | － | 63 | 177 | 52 | 408 | 1，197 | 1，760 | 901 | 729 | 168 |
| Smoked fish |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haddock | 1 | 0.24 | － | 1 |  | － | 39 | 1 | 19 | － | 23 | － | 3 | － | 15 |  |  |  | 5 | 23 | 5 | 24 | 59 | 25 | 75 | 67 | 8 |
| Herring | 1 | 10.04 | － | 52 |  | － | 908 | 192 | 641 | － | 768 | 66 | 138 | 7 | 540 | 513 | 30 |  | 906 | 1，508 | 691 | 1，640 | 1，643 | 1，152 | 5，943 | 4，892 | 1，051 |
| Mackerel | 1 | 16.95 | － | 318 | － | － | 1，525 | 298 | 1，376 | － | 1，160 | 111 | 230 | － | 1，299 | 1，149 | － | － | 1，565 | 2，329 | 948 | 2，883 | 3，219 | 1，658 | 9，871 | 8，008 | 1，795 |
| Salmon | 1 | 9.79 | － | 165 | － |  | 1，612 | 255 | 1，615 | － | 636 | 49 | 227 |  | 471 | 317 | － | － | 462 | 902 | 471 | 1，829 | 3，391 | 940 | 4，678 | 3，990 | 689 |
| Seafood－based dish |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Fish soup | 2 | 0.98 |  | 12 |  |  | 131 | 22 | 118 |  | 76 | 2 | 16 |  | 37 | 15 |  |  | 32 | 55 | 40 | 139 | 260 | 101 | 332 | 285 | 48 |
| Paella | 1 | 4.06 | － | 62 |  |  | 938 | 114 | 283 |  | 1，303 | 106 | 726 |  | 76 | 30 |  |  | 101 | 110 | 18 | 145 | 1，300 | 1，552 | 1，205 | 378 | 827 |
| Surimi | 1 | 2.29 |  | 28 |  |  | 327 | 28 | 64 | － | 662 | 64 | 219 | － | 88 | 60 |  |  | － | 152 |  | 307 | 415 | 754 | 821 | 606 | 219 |
| Tarama | 1 | 38.66 | － | 109 | － | － | 2，173 | 129 | 115 | 69 | 23，119 | － | 6，435 | 135 | 624 | 279 | － | － | 411 | 755 | 255 | 525 | 2，539 | 23，513 | 12，609 | 2，437 | 6，981 |
| $-:<$ LOQ（ $1 \mathrm{mg} / \mathrm{g}$ lipid）；a ：Nb composite samples．Each sample is composed by 5 subsamples of the same species，representative of the provisioning methods in each zone for SFA，MUFA，PUFA and LCPUFA include all the analysed fatty acids，some of which are minority and are not presented in the table．＊Predatory fish as described in the January 2005 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

### 3.2 Contamination by trace elements

Table 20 indicates the proportion of censured data, in other words less than the detection limit (LOD).
For all the products sampled - fish, molluscs, crustaceans and other products - arsenic is present mainly in the form arsenobetaine (organic arsenic) considered as non-toxic. The forms of speciation of inorganic toxic arsenics, $\mathrm{As}(\mathrm{III})$ and $\mathrm{As}(\mathrm{V})$, are in a minority, but the $\mathrm{As}(\mathrm{III})$ is detected in $91.2 \%$ of the samples whereas the $\mathrm{As}(\mathrm{V})$ never is. The total inorganic arsenic is therefore constituted mainly by the species As(III) considered to be the most toxic.

The samples contain mainly butyltins along with some phenyltins. Apart from monooctyltin (MOT), detectable in $10 \%$ of the samples, the presence of octyltins is sporadic.

In all the foods, mercury is mainly found in its toxic methyl form, methylmercury.
Finally, more than $54 \%$ of the samples contain lead and cadmium in detectable quantities.
Table 20: Proportion of censured data (<LOD) in analysing the trace elements in food samples.

| Trace element | \% censured data (<LOD) |
| :--- | :---: |
| As total | 0.00 |
| AsB | 0.00 |
| MMA | 85.5 |
| DMA | 16.3 |
| As(III) | 8.80 |
| As(V) | 100 |
| Hg total | 0.00 |
| MeHg | 0.00 |
| Cd total | 45.3 |
| Pb total | 30.2 |
| MBT | 15.1 |
| DBT | 10.7 |
| TBT | 12.6 |
| MPT | 49.1 |
| DPT | 46.5 |
| TPT | 52.2 |
| MOT | 90.0 |
| DOT | 99.4 |
| TOT | 100 |

Tables 21 to 24 present the results of the trace elements (in $\mu \mathrm{g} / \mathrm{g}$ fresh weight) in the sampled foods, averaged for the four regions.

### 3.2.1 Fresh and frozen fish

Arsenic : The species of fish containing the highest concentrations of total arsenic are bottom fish (plaice, pout, sole, goatfish, ray, common dab and catshark). Their levels are between 12 and $34 \mu \mathrm{~g} / \mathrm{g}$ fresh weight. Fish containing the highest levels of toxic inorganic arsenic (As(III) and As(V)) are pout, ray and goatfish with respectively $0.077,0.073$ and $0.072 \mu \mathrm{~g} / \mathrm{g}$. The least contaminated by inorganic arsenic are eel with $0.009 \mu \mathrm{~g} / \mathrm{g}$, ling, grenadier and emperor with $0.012 \mu \mathrm{~g} / \mathrm{g}$. Inorganic arsenic in fresh and frozen fish represents $0.1 \%$ to $3.5 \%$ of total arsenic, which is consistent with the literature ${ }^{63}$. Our results are particularly consistent with the British FSA study in October $2005{ }^{108}$.

Mercury : As regards total mercury, the data shown in Table 21 are particularly consistent with the results of MAP monitoring plan ${ }^{109}$, for all species. Predator fish are found to contain the highest levels of methylmercury (MeHg): swordfish ( $0.94 \mu \mathrm{~g} / \mathrm{g}$ ), emperor ( $0.57 \mu \mathrm{~g} / \mathrm{g}$ ), tuna ( $0.33 \mu \mathrm{~g} / \mathrm{g}$ ) and eel ( $0.32 \mu \mathrm{~g} / \mathrm{g}$ ). These species also have the highest levels of total mercury, although they do not exceed the maximum authorised limit of $1 \mathrm{mg} / \mathrm{kg}^{110}$, which is reassuring when we consider that our results are obtained from composite samples. None of the non-predator species exceeds the maximum level of $0.5 \mathrm{mg} / \mathrm{kg}^{110}$. The fish the least contaminated by MeHg are anchovy ( $0.020 \mu \mathrm{~g} / \mathrm{g}$ ), salmon ( $0.038 \mu \mathrm{~g} / \mathrm{g}$ ) and saithe $(0.041 \mu \mathrm{~g} / \mathrm{g})$. MeHg represents $67 \%$ to $100 \%$ of the total mercury in fish, within the limits of measuring uncertainty of the two analysis techniques employed.

Cadmium : Some species have cadmium levels exceeding $0.30 \mu \mathrm{~g} / \mathrm{g}, 0.10 \mu \mathrm{~g} / \mathrm{g}$ or $0.05 \mu \mathrm{~g} / \mathrm{g}$, the maximum authorised species to species ${ }^{110}$ : saithe ( $0.07 \mu \mathrm{~g} / \mathrm{g}$ ), swordfish ( $0.07 \mu \mathrm{~g} / \mathrm{g}$ ), and catshark with the highest observed cadmium level ( $0.42 \mu \mathrm{~g} / \mathrm{g}$ ). These high average values are due to the exceptional contamination of the composite sample from Le Havre ( $1.65 \mu \mathrm{~g} / \mathrm{g}$ ), the concentrations of the composite samples of the three other zones do not exceed the maximal authorised level.

Lead: Halibut is the fish the most contaminated by lead $(0.1 \mu \mathrm{~g} / \mathrm{g})$. However no species exceeds the maximum authorised limits ( 0.2 to $0.4 \mu \mathrm{~g} / \mathrm{g})^{110}$. Our data are generally consistent with the MAP monitoring plans ${ }^{109}$, both for lead and cadmium, considering that many levels in these plans are less than the analytical limits of our study.

Organic tin : Concerning the organostannic compounds (OTC), many of our data are censured: $11 \%$ to $15 \%$ for the butyltin, $47 \%$ to $52 \%$ for the phenyltin and more than $90 \%$ for the octyltin. The fish presenting the highest levels are halibut $(0.023 \mu \mathrm{~g} / \mathrm{g})$ and swordfish ( $0.019 \mu \mathrm{~g} / \mathrm{g}$ ), all forms of organic tin included. These results are consistent with the data of the SCOOP task 3.2.13. ${ }^{65}$.

### 3.2.2 Molluscs and crustaceans

Arsenic: Octopus is the species with the highest level of total arsenic ( $42.3 \mu \mathrm{~g} / \mathrm{g}$ ), as shown in Table 22. However it should be underlined once again that we had only one composite sample ( 5 sub-samples) coming from the Toulon region. Otherwise crustaceans are found to be the most contaminated by arsenic, with $37.2 \mu \mathrm{~g} / \mathrm{g}$ for the spider crab and $16.8 \mu \mathrm{~g} / \mathrm{g}$ for crabs including the common crab; these also have very high levels of toxic inorganic arsenic ( 0.188 and $0.257 \mu \mathrm{~g} / \mathrm{g}$ respectively). We also note a high inorganic arsenic level in Mediterranean sea urchins ( $0.222 \mu \mathrm{~g} / \mathrm{g}$ ). Unlike fish, the total arsenic and inorganic arsenic levels in our mollusc and crustacean samples exceed those found in the FSA study in 2005, which is perhaps explained by different provisioning (local fishing, etc.).

Mercury : Octopus and crab are also the species the most contaminated by mercury, in particular by MeHg ( 0.219 and $0.175 \mu \mathrm{~g} / \mathrm{g}$ respectively). However none exceed the limit fixed at $0.5 \mu \mathrm{~g} \mathrm{Hg} / \mathrm{g}$ for fish products (excluding predator fish) ${ }^{110}$. The sea urchin is the least contaminated species, with less than $0.003 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{g}$.

Cadmium : The maximum authorised limits ${ }^{110}$ are exceeded by several species, in particular crab (4.1 versus $0.5 \mu \mathrm{~g} / \mathrm{g}$ ), shrimp ( 1.1 versus $0.5 \mu \mathrm{~g} / \mathrm{g}$ ) and calico scallops ( 1.1 versus $1.0 \mu \mathrm{~g} / \mathrm{g}$ ). The other bivalve molluscs display lower cadmium levels not exceeding $0.04 \mu \mathrm{~g} / \mathrm{g}$. These average contaminations are greater than the mean levels measured in the French monitoring plans: $0.46 \mu \mathrm{~g} / \mathrm{g}$ for crab or $0.05 \mu \mathrm{~g} / \mathrm{g}$ for shrimp.

[^20]These differences are due to high contamination of our composite crab sample in Lorient ( $12 \mu \mathrm{~g} / \mathrm{g}$ versus less than $1 \mu \mathrm{~g} / \mathrm{g}$ in the other sampling zones) and of our composite shrimp sample in Le Havre ( $4 \mu \mathrm{~g} / \mathrm{g}$ versus less than $0.05 \mu \mathrm{~g} / \mathrm{g}$ on the other zones).

Lead : None of the species sampled exceeds the maximum authorised limits ${ }^{110}$. For the common samples, in the case of both lead and cadmium our levels are generally lower than those found in the FSA study ${ }^{108}$.

Organic tin : The organic tin levels are relatively low with a lower maximum value than for fish ( 0.01 $\mu \mathrm{g} / \mathrm{g}$ for squid and swimcrab). As for fish, these results are consistent with the data found in the SCOOP task 3.2.13. ${ }^{65}$

### 3.2.3 Other seafood

Among other seafood, the canned products are the most contaminated by trace elements (Table 23). We find total arsenic levels of $3.54 \mu \mathrm{~g} / \mathrm{g}$ in canned sardine and $2.23 \mu \mathrm{~g} / \mathrm{g}$ in canned crab; more particularly, for inorganic arsenic the average level is $0.07 \mu \mathrm{~g} / \mathrm{g}$ for canned crab and pilchard. The same is true for mercury, the maximum level being found in canned tuna ( $0.2 \mu \mathrm{~g} / \mathrm{g}$ ), but without exceeding the authorised limits ${ }^{110}$.

On the other hand, canned or bottled anchovy and canned sardines reveal cadmium concentrations higher than the authorised limits ( 0.35 and $0.22 \mu \mathrm{~g} / \mathrm{g}$ versus $0.1 \mu \mathrm{~g} / \mathrm{g}$ ). However these results must be interpreted with care in view of the homogenisation problems encountered during the sampling of canned products.

The other products, smoked fish or prepared seafood-based dishes, contain very low levels of trace elements.

Finally, the organic tin levels are equivalent to those measured in the composite samples of molluscs and crustaceans ( $0.01 \mu \mathrm{~g} / \mathrm{g}$ ).

### 3.2.4 Regional variations

Despite the deliberate choice of zones contrasted by the existence of old local environmental pollution, no significant differences are found between the trace element contaminations measured in the different study zones, for all the trace elements except inorganic arsenic (Table 24). For the species common to two zones, fish sampled in Lorient contain significantly more inorganic arsenic than fish in Toulon ( $p<0.05$, significant differences not presented). They also appear to be more contaminated than fish sampled in Le Havre and La Rochelle, although the difference is not significant.

As regards molluscs and crustaceans, it appears that the samples from Le Havre are more contaminated by organic tin than those in the other zones. However, this result must be interpreted with care in view of the small number of samples ( 10 to 12 species or 40 to 48 sub-samples depending on the zone), the large number of censured data and the fact that the sampled species are different from one zone to another (molluscs, crustaceans, etc.).

This absence of regional differences can be largely explained by the low representation of explicitly local origins in the provisioning of seafood products mentioned earlier and by the general respect of local fishing interdictions when these exist.

| Fish | ${ }^{\text {a }}$ | Ast $^{\text {b }}$ | AsB | MMA | DMA | As(III) | As(V) | $\mathrm{Hg}_{\mathbf{T}}$ | MeHg | Cd | Pb | отс | MBT | DBT | TBT | MPT | DPT | TPT | M | DOT | TOT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchovy | 1 | 0.94 | 0.72 | 0.025 | 0.003 | 0.014 | 0.005 | 0.012 | 0.020 | 0.0295 | 0.0075 | 0.0039 | 0.0013 | 0.0006 | 0.0015 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Angler fish* | 4 | 6.00 | 6.64 | 0.004 | 0.004 | 0.012 | 0.005 | 0.147 | . 131 | 0.0002 | 0.0031 | 0.0031 | 0.0006 | 0.0006 | 0.0008 | 0.0002 | 0.0003 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Catshark* | 4 | 34.3 | 31.2 | . 004 | 014 | . 041 | 0.005 | 0.251 | 0.232 | 0.4183 | 0.0106 | 0.0063 | 0.0017 | 0.0014 | 0.0014 | 0.0003 | 0.0007 | 0.0005 | 0.0002 | 0.0001 | 0.0001 |
| Cod | 4 | 5.2 | 5.31 | 0.004 | 0.010 | 0.016 | 0.005 | 0.0 | 0.059 | 0.0004 | 0.0019 | 0.0034 | 0.0007 | 0.0007 | 0.0007 | 0.0004 | 0.0004 | , 002 | 0.0001 | . 001 | 0.0001 |
| Common dab | 4 | 21.8 | 19.9 | 0.004 | 0.0 | 0.029 | 0.005 | 0.10 | 0.098 | 0.0002 | 0.001 | 0.00 | 0.0006 | 0.000 | 0.0002 | 0.000 | 0.000 | 0.000 | 0.0001 | . 0001 | 0.0001 |
| Eel* | 1 | 0.7 | 0.58 | 0.004 | 0.004 | . 004 | 0.005 | 0.32 | . 315 | 0.0033 | 0.0205 | 0.0047 | 0.0001 | 0.0011 | 0.0016 | 0.0006 | 0.0004 | 0.0006 | 0.0001 | 0.0001 | 0.0001 |
| Emperor* | 3 | 1.19 | 0.83 | 0.004 | 0.002 | 0.007 | 0.005 | 0.609 | 0.574 | 0.0048 | 0.0176 | 0.0052 | 0.0005 | 0.0008 | 0.0016 | 0.000 | 0.0008 | 0.0007 | 0.00 | 0.0001 | . 0000 |
| Goatfish | 3 | 16.1 | 15.6 | . 38 | 0.041 | 67 | 0.005 | 0.120 | . 130 | 0.0005 | 0.0036 | 0.0013 | 0.0001 | 0.0002 | 0.0003 | 0.000 | . 0001 | 0.0001 | 0.00 | 0.0001 | . 0001 |
| Grenadier / hoki* | 4 | 3.90 | 4.16 | 0.004 | 0.002 | 0.007 | 0.005 | 0.109 | 0.112 | 0.0036 | 0.0041 | 0.0046 | 0.0014 | 0.0008 | 0.0011 | 0.0003 | 0.0005 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Gurrard | 1 | 6.28 | . 91 | 0.004 | 0.0 | 0.017 | 0.005 | 0.179 | 0.143 | 0.0002 | 0.0002 | 0.0011 | 0.0001 | 0.0003 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Haddock | 2 | 6.52 | 6.57 | 004 | 0.009 | 0.012 | 0.005 | 0.086 | 0.102 | 0.0036 | 0.0035 | 0.0028 | 0.0010 | 0.0005 | 0.0000 | 0.0004 | 0.0004 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Hake | 4 | 4.28 | 4.21 | 0.004 | 0.035 | . 017 | 0.005 | 0.148 | 0.15 | 0.0002 | 0.0085 | 0.0045 | 0.0011 | 0.0008 | 0.0012 | 0.0003 | 0.0005 | 0.0003 | 0.0001 | 0.0001 | 0.0001 |
| Halibut* | 4 | 5.69 | 4.98 | 0.004 | 0.053 | 0.012 | 0.005 | 0.07 | 0.082 | . 335 | 0.1001 | 0.0232 | 0.002 | 0.0059 | 0.010 | 0.000 | 0.0005 | 0.002 | 0.00 | . 000 | 0.0001 |
| John dory | 2 | 1.12 | 0.81 | 29 | 0.037 | . 01 | 0.005 | 0.07 | 0.092 | 0.044 | 0.013 | 0.008 | 0.001 | 0.001 | 0.002 | 0.000 | 0.00 | 0.000 | 0.000 | . 000 | 0.0001 |
| Ling* | 4 | 4.70 | 4.36 | . 004 | 0.004 | 0.007 | 0.005 | 0.310 | 0.305 | 0.0043 | 0.0003 | 0.0030 | 0.0010 | 0.0004 | 0.0007 | 0.0001 | 0.0004 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Mackerel | 4 | 2.41 | 1.70 | 0,078 | 0.138 | 0.028 | 0.005 | 0.047 | 0.072 | 0.0002 | 0.0022 | 0.0088 | 0.0021 | 0.0016 | 0.0024 | 0.0006 | 0.0008 | 0.0008 | 0.0003 | 0.0001 | 0.000 |
| Plaice* | 2 | 12.4 | 13.0 | . 004 | 0.022 | 0.014 | 0.005 | 0.069 | 0.059 | 0.0002 | 0.0036 | 0.0046 | 0.0015 | 0.0005 | 0.0004 | 0.0002 | 0.0007 | 0.0010 | 0.0001 | 0.0001 | 0.0001 |
| Pollack | 3 | 3.65 | 3.36 | 0.004 | . 006 | 39 | 0.005 | 0.081 | 83 | . 0006 | 0.000 | 0.0035 | 0.0004 | 0.0005 | 0.0005 | 0.0003 | 0.0008 | 0.0006 | 0.0001 | 0.0001 | 0.0001 |
| Pout | 1 | 13.8 | 16.8 | 0.004 | 0.070 | 0.072 | 0.005 | 0.149 | 0.158 | 0.0002 | 0.0024 | 0.0028 | 0.0006 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |  | 0.0015 | 0.0001 | 0.0001 |
| Ray* | 4 | 21.8 | 17.6 | 0.004 | 0.065 | 0.068 | 0.005 | 0.096 | 0.097 | 0.0388 | 0.0269 | 0.0026 | 0.00 | 0.000 | 0.00 | 0.00 | 0.0002 | 0.0001 | 0.00 | 0.0001 | 0.00 |
| Saithe/coalfish | 4 | 1.40 | 1.41 | 0.014 | 0.01 | 0.01 | 0.005 | 0.02 | 0.041 | 0.0719 | 0.03 | 0.0 | 0.00 | 0.00 | 0.002 | 0.00 | 0.00 | 0.0006 | 0.00 | 01 | 0.0001 |
| Salmon | 4 | 1.66 | 1.32 | 0.015 | 0.003 | 0.023 | 0.005 | 0.040 | 0.038 | 0.0002 | 0.0010 | 0.0059 | 0.0018 | 0.0009 | 0.0007 | 0.0004 | 0.0006 | 0.0005 | 0.0007 | 0.0001 | 0.0001 |
| Sardine | 4 | 6.02 | 5.81 | 0.025 | 0.100 | 0.047 | 0.005 | 0.070 | 0.099 | 0.0019 | 0.0194 | 0.0064 | 0.0012 | 0.0017 | 0.0015 | 0.0004 | 0.0004 | 0.0005 | 0.0005 | 0.0001 | 0.0001 |
| Scorpionfish | 1 | 1.92 | 1.85 | 0.004 | 0.014 | 010 | 0.005 | 0.172 | 0.196 | 0.0002 | 0.000 | 0.0018 | 0.0001 | 0.0006 | 0.0006 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Seabas* | 4 | 90 | 1.70 | . 004 | 0.007 | 0.021 | 0.005 | 0.144 | 0.149 | 0.0005 | 0.011 | 0.0110 | 0.0011 | 0.0014 | 0.004 | 0.0011 | 0.0015 | 0.0012 | 0.0003 | 0.0001 | 0.0001 |
| Sea bream* | 4 | 3.30 | 2.68 | 0.054 | 0.075 | 0.050 | 0.005 | 0.109 | 0.098 | 0.0002 | 0.0008 | 0.0051 | 0.0011 | 0.0010 | 0.0010 | 0.0005 | 0.0005 | 0.0004 | 0.0003 | 0.0001 | 0.0001 |
| Sole | 4 | 14.3 | 14.5 | 0.004 | 0.044 | 0.010 | 0.005 | 0.112 | 0.126 | 0.0014 | 0.0041 | 0.0024 | 0.0004 | 0.0004 | 0.00 | 0.0002 | 0.0004 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Swordfish* | 4 | 1.00 | 0.70 | . 013 | 0.026 | 0.03 | 0.00 | 0.84 | 0. 94 | . 0671 | 0.0002 | 0.0192 | 0.0066 | 0.0029 | 0.0077 | 0.0006 | 0.0006 | 0.0004 | 0.0001 | 0.0001 | 0.0001 |
| Tuna* | 4 | 2.45 | 1.79 | 0.029 | 0.016 | 0.008 | 0.005 | 0.33 | 0.330 | 0.0132 | 0.0004 | 0.0073 | 0.001 | 0.0018 | 0.001 | 0.000 | 0.000 | 0.000 | 0.00 | 0.000 | 0.0001 |
| Whiting | 4 | 3.85 | 3.75 | 0.011 | 0.006 | 0.027 | 0.005 | 0.2 | 0.170 | 0.0011 | , 013 | . 0037 | . 0009 | 0.0007 | 0.0005 | 0.000 | . 0004 |  | 0.0003 | 0.0001 | 0.0001 |
| a: Nb composite samples. Each sample is composed by 5 subsamples of the same species, representative of the provisioning methods in each zone (port, market, superr <br> b:The total for the forms of speciation of arsenic is not equal to the total arsenic ( $A S_{T}$ ) for all species since the concentrations supplied correspond to individual analyse <br> c : The column of OTC results corresponds to the sum of the 9 organostannic compounds <br> * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 22: Mean contamination by trace elements of molluscs and crustaceans ( $\mu \mathrm{g} / \mathrm{g}$ fresh weight)

| Mollusc, crustacean | $\mathrm{n}^{\text {a }}$ | $A s s_{T}{ }^{\text {b }}$ | AsB | MMA | DMA | As(III) | As(V) | $\mathrm{Hg}_{\mathrm{T}}$ | MeHg | Cd | Pb | OTC ${ }^{\text {c }}$ | MBT | DBT | TBT | MPT | DPT | TPT | MOT | DOT | TOT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calico scallop | 1 | 2.42 | 1.90 | 0.004 | 0.190 | 0.003 | 0.005 | 0.011 | 0.007 | 1.1391 | 0.0931 | 0.0040 | 0.0005 | 0.0004 | 0.0025 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Codkle | 2 | 1.78 | 1.50 | 0.004 | 0.024 | 0.105 | 0.005 | 0.018 | 0.016 | 0.0358 | 0.0437 | 0.0074 | 0.0006 | 0.0013 | 0.0046 | 0.0002 | 0.0002 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Common periwinkle | 3 | 6.39 | 4.08 | 0.020 | 0.055 | 0.185 | 0.005 | 0.011 | 0.009 | 0.1890 | 0.0901 | 0.0033 | 0.0009 | 0.0007 | 0.0006 | 0.0004 | 0.0002 | 0.0003 | 0.0001 | 0.0001 | 0.0001 |
| Crab | 3 | 16.8 | 13.1 | 0.004 | 0.051 | 0.252 | 0.005 | 0.176 | 0.175 | 4.0954 | 0.0189 | 0.0087 | 0.0011 | 0.0014 | 0.0028 | 0.0003 | 0.0004 | 0.0011 | 0.0014 | 0.0001 | 0.0001 |
| Cuttle fish | 2 | 5.59 | 5.30 | 0.014 | 0.035 | 0.030 | 0.005 | 0.040 | 0.048 | 0.0559 | 0.0921 | 0.0035 | 0.0014 | 0.0009 | 0.0007 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Great scallop | 4 | 2.96 | 2.41 | 0.004 | 0.049 | 0.096 | 0.005 | 0.025 | 0.034 | 0.2695 | 0.0665 | 0.0098 | 0.0009 | 0.0010 | 0.0067 | 0.0004 | 0.0002 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Lobster | 1 | 7.08 | 5.35 | 0.004 | 0.094 | 0.041 | 0.005 | 0.073 | 0.092 | 0.4326 | 0.0039 | 0.0012 | 0.0002 | 0.0002 | 0.0002 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Mussel | 4 | 6.61 | 5.55 | 0.023 | 0.174 | 0.089 | 0.005 | 0.041 | 0.038 | 0.0329 | 0.1073 | 0.0033 | 0.0008 | 0.0005 | 0.0011 | 0.0002 | 0.0002 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Octopus | 1 | 42.3 | 32.0 | 0.004 | 0.015 | 0.226 | 0.005 | 0.340 | 0.219 | 0.0324 | 0.0598 | 0.0053 | 0.0013 | 0.0005 | 0.0029 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Oyster | 4 | 2.20 | 1.61 | 0.005 | 0.080 | 0.109 | 0.005 | 0.007 | 0.007 | 0.0343 | 0.0298 | 0.0075 | 0.0012 | 0.0012 | 0.0042 | 0.0002 | 0.0001 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Scampi | 3 | 8.75 | 7.08 | 0.004 | 0.006 | 0.089 | 0.005 | 0.084 | 0.087 | 0.1077 | 0.0314 | 0.0063 | 0.0010 | 0.0013 | 0.0006 | 0.0004 | 0.0009 | 0.0019 | 0.0001 | 0.0001 | 0.0001 |
| Sea urchin | 1 | 3.25 | 2.90 | 0.004 | 0.030 | 0.217 | 0.005 | 0.006 | 0.003 | 0.0643 | 0.1488 | 0.0053 | 0.0008 | 0.0006 | 0.0033 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Shrimp | 4 | 1.31 | 1.12 | 0.004 | 0.020 | 0.009 | 0.005 | 0.033 | 0.031 | 1.0915 | 0.0072 | 0.0021 | 0.0003 | 0.0003 | 0.0007 | 0.0003 | 0.0001 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Spider crab | 1 | 372 | 27.4 | 0.159 | 0.067 | 0.183 | 0.005 | 0.034 | 0.036 | 0.4606 | 0.0583 | 0.0016 | 0.0003 | 0.0002 | 0.0000 | 0.0004 | 0.0001 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Squid | 4 | 5.92 | 5.06 | 0.004 | 0.008 | 0.006 | 0.005 | 0.049 | 0.055 | 0.0511 | 0.0071 | 0.0133 | 0.0009 | 0.0014 | 0.0100 | 0.0003 | 0.0002 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Swimcrab | 2 | 10.1 | 8.37 | 0.004 | 0.157 | 0.053 | 0.005 | 0.073 | 0.069 | 0.1274 | 0.1254 | 0.0140 | 0.0052 | 0.0036 | 0.0027 | 0.0005 | 0.0002 | 0.0015 | 0.0001 | 0.0001 | 0.0001 |
| Whelk | 3 | 15.8 | 14.5 | 0.004 | 0.018 | 0.077 | 0.005 | 0.051 | 0.034 | 0.7807 | 0.0603 | 0.0054 | 0.0010 | 0.0011 | 0.0020 | 0.0002 | 0.0003 | 0.0005 | 0.0001 | 0.0001 | 0.0001 |
| a: Nb composite samples. Each sample is composed by 5 subsamples of the same species, representative of the provisioning methods in each zone (port, market, superm <br> $b$ : The total for the forms of speciation of arsenic is not equal to the total arsenic (As $s_{T}$ ) for all species since the concentrations supplied correspond to individual analyses c: The column of OTC results corresponds to the sum of the 9 organostannic compounds |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 23: Mean contamination by trace elements of other seafood ( $\mu \mathrm{g} / \mathrm{g}$ fresh weight)

| Other seafood | $n^{\text {a }}$ | $\mathrm{As}^{\text { }}{ }^{\text {b }}$ | AsB | MMA | DMA | As(III) | As(V) | $\mathrm{Hg}_{\text {T }}$ | MeHg | Cd | Pb | OTC ${ }^{\text {c }}$ | MBT | DBT | TBT | MPT | DPT | TPT | MOT | DOT | TOT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Canned food |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Anchovy | 2 | 0.80 | 0.57 | 0.019 | 0.036 | 0.018 | 0.005 | 0.022 | 0.016 | 0.3506 | 0.0571 | 0.0093 | 0.0033 | 0.0013 | 0.0041 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Crab | 1 | 2.23 | 2.05 | 0.004 | 0.003 | 0.067 | 0.005 | 0.053 | 0.136 | 0.1713 | 0.0109 | 0.0050 | 0.0021 | 0.0013 | 0.0010 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Mackerel | 1 | 0.70 | 0.18 | 0.082 | 0.186 | 0.006 | 0.005 | 0.031 | 0.027 | 0.0446 | 0.0058 | 0.0130 | 0.0075 | 0.0021 | 0.0028 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Pilchard | 1 | 1.91 | 1.59 | 0.004 | 0.139 | 0.064 | 0.005 | 0.020 | 0.021 | 0.0126 | 0.0059 | 0.0041 | 0.0008 | 0.0004 | 0.0022 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Sardine | 1 | 3.54 | 2.14 | 0.004 | 0.018 | 0.030 | 0.005 | 0.043 | 0.023 | 0.2159 | 0.2882 | 0.0100 | 0.0058 | 0.0017 | 0.0019 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Tuna* | 5 | 0.81 | 0.94 | 0.004 | 0.011 | 0.003 | 0.005 | 0.190 | 0.210 | 0.0178 | 0.0023 | 0.0139 | 0.0050 | 0.0058 | 0.0022 | 0.0002 | 0.0002 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
| Smoked fish |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haddock | 1 | 1.06 | 1.16 | 0.004 | 0.001 | 0.016 | 0.005 | 0.007 | 0.010 | 0.0002 | 0.0097 | 0.0009 | 0.0002 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Herring | 1 | 1.25 | 1.62 | 0.004 | 0.015 | 0.008 | 0.005 | 0.022 | 0.037 | 0.0002 | 0.0053 | 0.0033 | 0.0008 | 0.0007 | 0.0012 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Mackerel | 1 | 2.06 | 1.58 | 0.004 | 0.307 | 0.030 | 0.005 | 0.025 | 0.042 | 0.0027 | 0.0045 | 0.0063 | 0.0020 | 0.0009 | 0.0028 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Salmon | 1 | 2.00 | 1.48 | 0.004 | 0.049 | 0.001 | 0.005 | 0.027 | 0.037 | 0.0002 | 0.0053 | 0.0056 | 0.0009 | 0.0008 | 0.0032 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Seafood-based dish |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Fish soup | 2 | 0.68 | 0.89 | 0.004 | 0.003 | 0.019 | 0.005 | 0.009 | 0.007 | 0.0151 | 0.0070 | 0.0012 | 0.0001 | 0.0001 | 0.0005 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Paella | 1 | 0.11 | 0.27 | 0.004 | 0.007 | 0.001 | 0.005 | 0.000 | 0.000 | 0.0128 | 0.0119 | 0.0013 | 0.0001 | 0.0002 | 0.0004 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Surimi | 1 | 0.39 | 0.33 | 0.004 | 0.001 | 0.006 | 0.005 | 0.021 | 0.018 | 0.0083 | 0.0033 | 0.0083 | 0.0020 | 0.0046 | 0.0011 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| Tarama. terrine ou mousse | 1 | 0.18 | 1.18 | 0.004 | 0.118 | 0.001 | 0.005 | 0.001 | 0.001 | 0.0002 | 0.0002 | 0.0008 | 0.0001 | 0.0001 | 0.0000 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
| a : Nb composite samples. Each sample is composed by 5 subsamples of the same species, representative of the provisioning methods in each zone (port, market, superm <br> b: The total for the forms of speciation of arsenic is not equal to the total arsenic (Ass) for all species since the concentrations supplied correspond to individual analyses <br> c: The column of OTC results corresponds to the sum of the 9 organostannic compounds <br> * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 24: Mean contamination by trace elements of fish (excluding eel), molluscs and crustaceans per site ( $\mu \mathrm{g} / \mathrm{g}$ fresh weight)

|  |  |  | $\mathrm{As}^{\text {T }}{ }^{\text {b }}$ | AsB | MMA | DMA | As(III) | As(V) | $\mathrm{Hg}_{\text {T }}$ | MeHg | Cd | Pb | OTC ${ }^{\text {c }}$ | MBT | DBT | TBT | MPT | DPT | TPT | MOT | DOT | TOT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean | 7.67 | 7.40 | 0.020 | 0.028 | 0.021 | 0.005 | 0.137 | 0.141 | 0.0816 | 0.0055 | 0.0078 | 0.0017 | 0.0015 | 0.0024 | 0.0005 | 0.0005 | 0.0010 | 0.0001 | 0.0001 | 0.0001 |
|  |  | SD | 8.36 | 7.79 | 0.056 | 0.040 | 0.018 | 0.000 | 0.155 | 0.174 | 0.3606 | 0.0097 | 0.0069 | 0.0016 | 0.0011 | 0.0039 | 0.0004 | 0.0004 | 0.0018 | 0.0001 | 0.0000 | 0.0000 |
|  | Mollusc. | Moy | 7.19 | 5.43 | 0.011 | 0.053 | 0.120 | 0.005 | 0.079 | 0.078 | 0.6311 | 0.0694 | 0.0164 | 0.0026 | 0.0026 | 0.0080 | 0.0005 | 0.0006 | 0.0014 | 0.0005 | 0.0001 | 0.0001 |
|  | Crustacean | SD | 5.50 | 4.10 | 0.025 | 0.052 | 0.101 | 0.000 | 0.140 | 0.136 | 1.3307 | 0.0650 | 0.0126 | 0.0028 | 0.0021 | 0.0103 | 0.0002 | 0.0006 | 0.0018 | 0.0012 | 0.0000 | 0.0000 |
| $\begin{aligned} & \stackrel{H}{0} \\ & \frac{0}{\partial} \end{aligned}$ |  | Mean | 7.78 | 7.51 | 0.037 | 0.039 | 0.043 | 0.005 | 0.175 | 0.168 | 0.0179 | 0.0092 | 0.0039 | 0.0010 | 0.0005 | 0.0007 | 0.0003 | 0.0005 | 0.0003 | 0.0003 | 0.0002 | 0.0001 |
|  |  | SD | 734 | 7.26 | 0.116 | 0.076 | 0.045 | 0.000 | 0.232 | 0.217 | 0.0569 | 0.0298 | 0.0042 | 0.0014 | 0.0007 | 0.0020 | 0.0003 | 0.0004 | 0.0002 | 0.0004 | 0.0003 | 0.0000 |
|  | Mollusc. | Mean | 10.90 | 8.68 | 0.023 | 0.058 | 0.102 | 0.005 | 0.043 | 0.047 | 1.1964 | 0.0428 | 0.0033 | 0.0004 | 0.0005 | 0.0014 | 0.0004 | 0.0002 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
|  | Crustacean | SD | 11.18 | 8.43 | 0.048 | 0.082 | 0.130 | 0.000 | 0.025 | 0.031 | 3.6751 | 0.0363 | 0.0025 | 0.0002 | 0.0006 | 0.0018 | 0.0002 | 0.0001 | 0.0001 | 0.0000 | 0.0000 | 0.0000 |
|  |  | Mean | 6.40 | 5.57 | 0.012 | 0.023 | 0.016 | 0.005 | 0.177 | 0.180 | 0.0109 | 0.0103 | 0.0055 | 0.0013 | 0.0010 | 0.0015 | 0.0004 | 0.0006 | 0.0004 | 0.0002 | 0.0001 | 0.0001 |
|  | Fish | SD | 8.11 | 5.76 | 0.023 | 0.033 | 0.016 | 0.000 | 0.175 | 0.165 | 0.0329 | 0.0238 | 0.0063 | 0.0024 | 0.0014 | 0.0024 | 0.0003 | 0.0007 | 0.0005 | 0.0003 | 0.0000 | 0.0000 |
|  | Mollusc. | Mean | 5.90 | 4.96 | 0.004 | 0.075 | 0.070 | 0.005 | 0.038 | 0.034 | 0.2779 | 0.0597 | 0.0038 | 0.0008 | 0.0006 | 0.0017 | 0.0002 | 0.0001 | 0.0002 | 0.0001 | 0.0001 | 0.0001 |
|  | Crustacean | SD | 5.57 | 5.15 | 0.000 | 0.143 | 0.089 | 0.000 | 0.041 | 0.037 | 0.5127 | 0.0527 | 0.0020 | 0.0007 | 0.0004 | 0.0015 | 0.0002 | 0.0001 | 0.0002 | 0.0000 | 0.0000 | 0.0000 |
| $\begin{aligned} & \text { 들 } \\ & \text { 응 } \end{aligned}$ |  | Mean | 829 | 7.57 | 0.008 | 0.030 | 0.020 | 0.005 | 0.211 | 0.220 | 0.0129 | 0.0200 | 0.0080 | 0.0013 | 0.0019 | 0.0031 | 0.0004 | 0.0005 | 0.0003 | 0.0003 | 0.0001 | 0.0001 |
|  |  | SD | 13.29 | 12.55 | 0.021 | 0.049 | 0.029 | 0.000 | 0.274 | 0.300 | 0.0339 | 0.0828 | 0.0108 | 0.0018 | 0.0039 | 0.0055 | 0.0003 | 0.0004 | 0.0003 | 0.0006 | 0.0000 | 0.0000 |
|  | Mollusc. | Mean | 8.36 | 6.98 | 0.006 | 0.048 | 0.093 | 0.005 | 0.059 | 0.050 | 0.1196 | 0.0479 | 0.0036 | 0.0007 | 0.0005 | 0.0018 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | 0.0001 |
|  | Crustacean | SD | 12.63 | 9.77 | 0.006 | 0.047 | 0.093 | 0.000 | 0.101 | 0.065 | 0.1654 | 0.0667 | 0.0025 | 0.0005 | 0.0003 | 0.0018 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | a: Nb composite samples. Each sample is composed by 5 subsamples of the same species, representative of the provisioning methods in each zone (port, market, superm <br> b: The total for the forms of speciation of arsenic is not equal to the total arsenic (Ass) for all species since the concentrations supplied correspond to individual analyses c: The column of OTC results corresponds to the sum of the 9 organostannic compounds |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Figure 11 clearly shows that the species with the highest MeHg levels are predator fish (swordfish, emperor, tuna, eel, ling and catshark). However, with the exception of swordfish, these are not necessarily the fish containing the most omega 3 fatty acids. Some fish with much lower concentrations of MeHg ( $<0.1 \mu \mathrm{~g} / \mathrm{g}$ ) have much higher levels of omega 3; examples include halibut, mackerel, salmon, anchovy and sardine. In other words, we find large disparities in the MeHg/Omega 3 ratio between species.

Those differences are less obvious for molluscs and crustaceans (Figure 12).
Nevertheless we note that for all the products (excepting canned and smoked products and prepared dishes) the lipid level correlates positively with the methylmercury level (Pearson correlation coefficient $r=0.27, p=0.01$ ). Similarly the level of $n-3$ LC-PUFAs (EPA, DPA and DHA) correlates positively with the methylmercury level ( $r=0.23, p=0.03$ ). This may be explained by the fact that the fish with the highest MeHg levels are at the end of the food chain: the MeHg accumulates along the chain. In parallel, some fatty acids including the precursors and long-chain $n-3$ and $n-6$ derivatives also accumulate. Kainz and his team propose a regulation of the assimilation of these fatty acids by marine organisms in order to optimise their physiological performance ${ }^{111}$.

We also note that while the MeHg/Omega 3 profile is homogeneous for fish (tuna, sardine, mackerel, anchovy, salmon), whatever its form (fresh, smoked, canned), this is not true for crab. The omega 3 concentration of canned crab is much lower than that of fresh and frozen crab despite a comparable MeHg level (not shown in Figures 11 and 12).
Figure 11: Mean Omega 3 and MeHg concentrations of fish ( $\mathrm{mg} / 100 \mathrm{~g}$ fresh weight)


* Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005
Figure 12: Mean Omega 3 and MeHg concentrations of molluscs and crustaceans ( $\mathrm{mg} / 100 \mathrm{~g}$ fresh weight)



### 3.3 Contamination by persistent organic pollutants

Tables 25 to 28 present the results of the analyses of persistent organic pollutants in sampled food, averaged for the four regions.

### 3.3.1 Fresh and frozen fish

PCDD/F et PCB-DL : Table 25 shows that the fish the most contaminated by dioxins (PCDD/Fs) and dioxinlike polychlorobiphenyls (DL-PCBs) are eel ( $88.3 \mathrm{pg} \mathrm{TEQ}_{\text {wHo }} / \mathrm{g}$ fresh weight) and sardine ( $10.6 \mathrm{pg} \mathrm{TEQ}_{\text {who }} / \mathrm{g}$ fresh weight). These are followed by the predators emperor fish and seabass with levels of 7.0 to $3.9 \mathrm{pg} \mathrm{TEQ}_{\mathrm{who}} / \mathrm{g}$ fresh weight. The least contaminated fish are catshark, anglerfish, saithe and cod with less than $0.15 \mathrm{pg} \mathrm{TEQ}_{\mathrm{who}} / \mathrm{g}$ fresh weight.

These results are consistent with the DGAL ${ }^{109}$ monitoring plans and with the English data ${ }^{112}$.
The results for eel are subject to reservation in view of the composition and origin of the sample, as explained earlier concerning the fatty acid composition. We note that this sample of eel exceeds the regulatory limit fixed for PCDD/Fs and DL-PCBs ( $12 \mathrm{pg} \mathrm{TEQ}_{\mathrm{wHO}} / \mathrm{g}$ fresh weight), as does the composite sample of sardine (limit fixed at $8 \mathrm{pg} \mathrm{TEQ}_{\mathrm{who}} / \mathrm{g}$ fresh weight). The other samples all have levels less than the limits ( 4 pg TEQ ${ }_{\text {who }} / \mathrm{g}$ fresh weight for PCDD/Fs and $8 \mathrm{pg} \mathrm{TEQ}_{\text {who }} / \mathrm{g}$ fresh weight for the total of the PCDD/Fs and DL-PCBs) ${ }^{113}$.
i-PCB : Eel and sardine are also the fish most heavy contaminated by "indicator polychlorobiphenyls" (i-PCB), with respectively 2,26 and $117 \mathrm{ng} / \mathrm{g}$ fresh weight. The i-PCBs being representative of the PCBs contamination, we find emperor, seabass and seabream in which the i-PCBs levels exceed $30 \mathrm{ng} / \mathrm{g}$ fresh weight. The least contaminated fish are saithe and cod with values of 1.1 and $1.2 \mathrm{ng} / \mathrm{g}$ fresh weight respectively. With the exception of the eel, these results are consistent with the DGAL ${ }^{109}$ monitoring plans. We recall that at present there is no regulation concerning the i-PCBs levels in fish and seafood products.

PBDE : The fish the most contaminated by polybromodiphenylethers (PBDE 28, 47, 99, 100, 153, 154, 183) is again eel with an average of $26.6 \mathrm{ng} / \mathrm{g}$ fresh weight. This level is exceptional in that the other fish have contamination levels of less than $3 \mathrm{ng} / \mathrm{g}$. The PBDEs level increases with the fat content: mackerel, anchovy, seabass, sardine and salmon all have moderately heavy high contaminations between 2 and $3 \mathrm{ng} / \mathrm{g}$ fresh weight. The least contaminated fish is catshark with $0.3 \mathrm{ng} / \mathrm{g}$ fresh weight and less than $1 \%$ of lipids. The PBDEs contamination does not depend on whether the species is a predator or not. These results are consistent with the JECFA data in 2005 on fish and seafood products ${ }^{114}$. As in the case of i-PCBs, to date there is no regulation concerning PBDEs levels in fish and seafood.

[^21]
### 3.3.2 Molluscs and crustaceans

When the contamination is expressed relative to grams of fat, molluscs and crustaceans generally have lower levels of persistent organic pollutants than fish (results not presented), but when the contamination is expressed relative to grams of fresh weight this difference (apart from eel) is less marked (Table 26).

PCDD/F and DL-PCB : The species most contaminated by PCDD/Fs and DL-PCBs are swimcrab with 18.6 $\mathrm{pg} \mathrm{TEQ}_{\text {wHo }} / \mathrm{g}$ fresh weight and crab with $6.5 \mathrm{pg} \mathrm{TEQ}_{\text {who }} / \mathrm{g}$ fresh weight. This can be partly explained by the very heavy contamination of the crab and swimcrab samples from Le Havre. Moreover, the composite swimcrab sample exceeds the regulatory limits for PCDD/Fs ( $4 \mathrm{pg} \mathrm{TEQ}{ }_{\text {who }} / \mathrm{g}$ fresh weight) and for PCDD/Fs and DL-PCBs (8 pg TEQ ${ }_{\text {who }} / \mathrm{g}$ fresh weight $)^{113}$. These species are followed by the spider crab with 5.6 pg $\mathrm{TEQ}_{\text {who }} / \mathrm{g}$ fresh weight. The least contaminated species are shrimp and periwinkle with $0.1 \mathrm{pg} \mathrm{TEQ}_{\mathrm{wHo}} / \mathrm{g}$ fresh weight.
i-PCB : Swimcrab, crab and spider crab also display the highest levels of i-PCB, with respectively 187,58 and $20 \mathrm{ng} / \mathrm{g}$ gross fresh weight. Shrimp and cockle are the least contaminated with 0.4 and $0.7 \mathrm{ng} / \mathrm{g}$ fresh weight respectively.

PBDE : Spider crab displays the highest level of PBDEs (28, 47, 99, 100, 153, 154, 183) with $3.0 \mathrm{ng} / \mathrm{g}$ fresh weight. Octopus and calico scallop are the species the least contaminated with PBDEs, with an average level below $0.2 \mathrm{ng} / \mathrm{g}$ fresh weight.

### 3.3.3 Other seafood

PCDD/F and DL-PCB : The products with the highest concentrations of PCDD/Fs and DL-PCBs are canned sardine with $3.9 \mathrm{pg} \mathrm{TEQ}_{\mathrm{who}} / \mathrm{g}$ fresh weight (Table 27). These data are quite consistent with the English data in $2006{ }^{112}$. No product exceeds the regulatory values.
i-PCB : Canned sardine also has the highest level of i-PCBs ( $35.5 \mathrm{ng} / \mathrm{g}$ fresh weight) along with smoked mackerel and smoked salmon ( 13.9 and $12.8 \mathrm{ng} / \mathrm{g}$ fresh weight respectively).

PBDE : Canned pilchard, smoked mackerel and smoked salmon have the highest levels of PBDEs (28, 47, $99,100,153,154,183$ ) with $3.2,2.8$ and $2.7 \mathrm{ng} / \mathrm{g}$ fresh weight respectively. These same products are relatively rich in total lipids (>10\%).

The products the least contaminated by POPs, all pollutants included, are paella and smoked haddock.
Table 25: Mean contamination by persistent organic pollutants of fresh and frozen fish

| Fish | Nb samples ${ }^{\text {a }}$ | $\begin{aligned} & \text { Lipids } \\ & (\mathrm{g} / 100 \mathrm{~g}) \end{aligned}$ | $\begin{gathered} \mathrm{PCDD} / \mathrm{F} \\ \left(\mathrm{pg} \mathrm{TEQ}_{\mathrm{oms} /} / \mathrm{g} \mathrm{FW}\right) \end{gathered}$ | $\begin{gathered} \text { PCB-DL } \\ \text { ( } \mathrm{pg} \text { TEQ } \mathrm{O}_{\mathrm{oms}} / \mathrm{g} \mathrm{FW} \text { ) } \end{gathered}$ | Total PCDD/F and PCB-DL ( pg TEQoms/g FW) | iPCB ( $\mathrm{ng} / \mathrm{g} \mathrm{FW}$ ) | $\begin{gathered} \text { PBDE } \\ (\mathrm{ng} / \mathrm{g} \text { FW) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchovy | 1 | 10.8 | 0.10 | 0.67 | 0.77 | 8.90 | 2.24 |
| Angler fish* | 4 | 0.33 | 0.03 | 0.08 | 0.11 | 1.67 | 0.46 |
| Atshark* | 4 | 0.88 | 0.03 | 0.08 | 0.10 | 2.38 | 0.27 |
| Cod | 4 | 0.52 | 0.03 | 0.11 | 0.14 | 1.19 | 0.54 |
| Common dab | 4 | 1.02 | 0.21 | 0.34 | 0.55 | 2.61 | 0.59 |
| Eel* | 1 | 22.1 | 1.50 | 86.8 | 88.3 | 2257 | 26.6 |
| Emperor* | 3 | 6.42 | 1.44 | 5.58 | 7.02 | 56.4 | 1.21 |
| Goatfish | 3 | 4.25 | 0.54 | 2.07 | 2.61 | 18.8 | 0.74 |
| Grenadier / hoki* | 4 | 0.59 | 0.08 | 0.09 | 0.17 | 2.83 | 0.52 |
| Gumard | 1 | 1.15 | 0.49 | 1.11 | 1.60 | 13.3 | 0.51 |
| Haddock | 2 | 0.37 | 0.07 | 0.21 | 0.28 | 2.74 | 0.64 |
| Hake | 4 | 0.96 | 0.04 | 0.26 | 0.30 | 3.36 | 0.49 |
| Halibut* | 4 | 12.5 | 0.89 | 1.37 | 2.27 | 15.0 | 1.59 |
| John dory | 2 | 0.91 | 0.08 | 0.41 | 0.50 | 5.99 | 0.51 |
| Ling* | 4 | 0.44 | 0.04 | 0.11 | 0.15 | 1.75 | 0.49 |
| Mackerel | 4 | 7.93 | 0.60 | 2.20 | 2.80 | 34.5 | 2.71 |
| Plaice* | 2 | 0.52 | 0.24 | 0.53 | 0.77 | 6.47 | 0.63 |
| Pollack | 3 | 0.30 | 0.02 | 0.23 | 0.25 | 3.26 | 0.41 |
| Pout | 1 | 0.43 | 0.05 | 0.18 | 0.23 | 1.95 | 0.42 |
| Ray* | 4 | 1.17 | 0.09 | 0.13 | 0.22 | 1.52 | 0.43 |
| Saithe/coalfish | 4 | 1.43 | 0.02 | 0.10 | 0.12 | 1.08 | 0.75 |
| Salmon | 4 | 13.5 | 0.50 | 1.32 | 1.82 | 14.5 | 2.55 |
| Sardine | 4 | 5.64 | 1.80 | 8.77 | 10.6 | 117 | 2.10 |
| Scorpionfish | 1 | 3.39 | 0.47 | 1.74 | 2.20 | 16.0 | 0.60 |
| Seabas** | 4 | 3.70 | 0.64 | 3.22 | 3.86 | 37.8 | 2.39 |
| Sea bream* | 4 | 5.49 | 0.38 | 2.20 | 2.58 | 26.9 | 1.10 |
| Sole | 4 | 0.50 | 0.05 | 0.15 | 0.21 | 4.91 | 0.39 |
| Swordfish* | 4 | 13.8 | 0.09 | 0.43 | 0.52 | 4.23 | 0.85 |
| Tuna* | 4 | 1.02 | 0.04 | 0.35 | 0.39 | 3.88 | 0.56 |
| Whiting | 4 | 0.42 | 0.05 | 0.24 | 0.29 | 4.26 | 0.54 |
| FW: fresh weight. a: Nb composite samples. Each sample is composed by 5 subsamples of the same species, representative of the provisioning methods in each zone (pot * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 |  |  |  |  |  |  |  |

Table 26: Mean contamination by persistent organic pollutants of molluscs and crustaceans

| Mollusc, crustacean | $\begin{aligned} & \text { © } \\ & \frac{0}{O} \\ & E \\ & \tilde{W} \\ & \hat{Z} \end{aligned}$ |  |  |  |  | 5 <br> 5 <br> 0 <br> 0 <br> 5 <br> 5 <br> 0 <br> 0 <br> 1 | PBDE ( $\mathrm{ng} / \mathrm{g}$ FW) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calico scallop | 1 | 1.37 | 0.20 | 0.15 | 0.34 | 3.15 | 0.20 |
| Cockle | 2 | 0.57 | 0.07 | 0.11 | 0.18 | 0.73 | 0.20 |
| Common periwinkle | 3 | 2.46 | 0.06 | 0.09 | 0.15 | 1.04 | 0.60 |
| Crab | 3 | 7.76 | 2.62 | 3.90 | 6.52 | 58.1 | 0.77 |
| Cuttle fish | 2 | 1.63 | 0.07 | 0.10 | 0.16 | 2.15 | 0.23 |
| Great scallop | 4 | 1.31 | 0.20 | 0.19 | 0.39 | 4.98 | 0.30 |
| Lobster | 1 | 2.03 | 0.72 | 0.81 | 1.52 | 4.38 | 0.42 |
| Mussel | 4 | 1.48 | 0.23 | 0.33 | 0.56 | 3.95 | 0.45 |
| Octopus | 1 | 0.66 | 0.06 | 0.20 | 0.25 | 1.97 | 0.20 |
| Oyster | 4 | 0.93 | 0.27 | 0.32 | 0.60 | 2.70 | 0.25 |
| Scampi | 3 | 0.86 | 0.27 | 0.20 | 0.47 | 1.82 | 0.28 |
| Sea urchin | 1 | 1.09 | 0.04 | 0.24 | 0.28 | 1.34 | 0.25 |
| Shrimp | 4 | 1.22 | 0.05 | 0.06 | 0.11 | 0.44 | 0.32 |
| Spider crab | 1 | 4.94 | 2.36 | 3.22 | 5.58 | 19.5 | 3.01 |
| Squid | 4 | 1.87 | 0.33 | 0.59 | 0.91 | 6.59 | 0.69 |
| Swimcrab | 2 | 4.66 | 4.79 | 13.8 | 18.6 | 187 | 1.03 |
| Whelk | 3 | 1.45 | 0.49 | 0.20 | 0.68 | 1.74 | 0.38 |

Table 27: Mean contamination by persistents organic pollutants of other seafood

| Other seafood | $\begin{aligned} & \text { そ } \\ & \text { 를 } \\ & \text { N } \\ & \text { C } \end{aligned}$ |  |  |  |  | K 0 0 0 0 0 0 0 0 | PBDE (ng/g FW) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Canned food |  |  |  |  |  |  |  |
| Anchovy | 2 | 8.19 | 0.03 | 0.14 | 0.17 | 1.21 | 1.02 |
| Crab | 1 | 13.9 | 0.05 | 0.05 | 0.10 | 0.22 | 0.74 |
| Mackerel | 1 | 10.4 | 0.21 | 0.65 | 0.86 | 6.06 | 1.46 |
| Pilchard | 1 | 13.0 | 0.85 | 0.93 | 1.78 | 9.32 | 3.24 |
| Sardine | 1 | 7.84 | 0.77 | 3.12 | 3.87 | 35.5 | 1.36 |
| Tuna* | 5 | 0.88 | 0.02 | 0.07 | 0.09 | 1.47 | 0.59 |
| Smoked fish |  |  |  |  |  |  |  |
| Haddock | 1 | 0.35 | 0.032 | 0.038 | 0.070 | 0.315 | 0.275 |
| Herring | 1 | 10.3 | 0.346 | 0.434 | 0.779 | 4.963 | 0.958 |
| Mackerel | 1 | 17.1 | 0.331 | 1.014 | 1.345 | 13.89 | 2.828 |
| Salmon | 1 | 10.3 | 0.303 | 1.057 | 1.360 | 12.84 | 2.733 |
| Seafood-based dish |  |  |  |  |  |  |  |
| Fish soup | 2 | 0.98 | 0.036 | 0.095 | 0.131 | 0.920 | 0.240 |
| Paella | 1 | 28.0 | 0.027 | 0.019 | 0.046 | 0.173 | 0.199 |
| Surimi | 1 | 4.08 | 0.013 | 0.023 | 0.036 | 1.260 | 0.628 |
| Tarama, terrine ou mousse | 1 | 4.43 | 0.026 | 0.062 | 0.089 | 1.159 | 1.060 |

FW: fresh weight. a: Nb composite samples. Each sample is composed by 5 subsamples of the same species, representative of the provisioning methods in each zone (port, market, supermarket...).

* Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005


### 3.3.4 Regional variations

Generally speaking, the contamination by persistent organic pollutants of our fish and seafood samples displays a north-south gradient. The Le Havre samples are the most contaminated, for all the pollutants considered, and the Toulon samples are the least contaminated (Table 28), although these differences are not statistically significant (on all the products and on the 19 common fish sampled in the four zones).

In Le Havre, the average PCDD/Fs and DL-PCBs contamination is $1.9 \mathrm{pg} \mathrm{TEQ}_{w H O} / \mathrm{g}$ fresh weight for fish and $5.1 \mathrm{pg} \mathrm{TEQ}_{\mathrm{wHO}} / \mathrm{g}$ fresh weight for molluscs and crustaceans. The average i-PCB contamination reaches $20.5 \mathrm{ng} / \mathrm{g}$ fresh weight for fish and $55.0 \mathrm{ng} / \mathrm{g}$ for molluscs and crustaceans. The average PBDEs (28, 47, $99,100,153,154,183$ ) contamination is $1.3 \mathrm{ng} / \mathrm{g}$ of fresh weight for fish and $0.6 \mathrm{ng} / \mathrm{g}$ of fresh weight for molluscs and crustaceans, the highest average being found in the Lorient samples (just slightly higher at $0.7 \mathrm{ng} / \mathrm{g}$ fresh weight). However, we note that these averages are not calculated for the same species in the four regions, but for species that in each region cover about $90 \%$ of the fish and seafood consumption of heavy consumers see 1.3.

The samples in Toulon are generally the least contaminated with POPs with average PCDD/Fs and DL-PCBs levels of $1.1 \mathrm{pg} \mathrm{TEQ}_{\text {who }} / \mathrm{g}$ in fish and $0.39 \mathrm{pg}_{\mathrm{TEQ}_{\text {wHo }} / \mathrm{g}}$ in molluscs and crustaceans. The average i-PCBs contamination is $12.2 \mathrm{ng} / \mathrm{g}$ for fish and $2.1 \mathrm{ng} / \mathrm{g}$ for molluscs and crustaceans. Finally, the average PBDEs contaminations of the Toulon samples are again the lowest: $0.8 \mathrm{ng} / \mathrm{g}$ fresh weight for fish and $0.3 \mathrm{ng} / \mathrm{g}$ fresh weight for molluscs and crustaceans.

When the contamination is expressed per gram of fat rather than fresh weight, the contamination gradient still exists for PBDEs contamination of fish and PCDD/Fs and DL-PCBs contamination of molluscs and crustaceans (results not presented). We also note that the PBDEs contamination of molluscs and crustaceans, when expressed per gram of fat, is relatively homogeneous across the study zones. The same is true for i-PCBs contamination of fish, molluscs and crustaceans (with the exception of samples from Le Havre). The high PCDD/Fs, DL-PCBs and i-PCBs levels found in molluscs and crustaceans from Le Havre are due to the heavy contamination of the crab and swimcrab samples.

Table 28: Mean contamination by persistent organic pollutants of fish (excluding eel), molluscs and crustaceans per site

|  |  | $\begin{aligned} & \text { § } \\ & \frac{0}{O} \\ & \text { B } \\ & 0 \\ & \mathbf{2} \end{aligned}$ |  | 6 <br> 8 <br> 8 <br> -8 <br> -6 <br> 0 <br> 0 <br> 0 |  |  |  | $\begin{aligned} & 3 \\ & \underline{y} \\ & 0 \\ & 6 \\ & 5 \\ & 0 \\ & 0 \\ & 2 \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Fish | 22 | Mean | 3.74 | 0.363 | 1.565 | 1.929 | 20.49 | 1.311 |
|  |  |  | SD | 4.78 | 0.625 | 3.580 | 4.176 | 48.94 | 1.455 |
|  | Mollusc, Crustacean | 10 | Mean | 2.70 | 1.440 | 3.707 | 5.147 | 55.01 | 0.616 |
|  |  |  | SD | 2.13 | 2.457 | 7.852 | 0.28 | 114.9 | 0.507 |
|  | Fish | 27 | Mean | 3.24 | 0.341 | 1.277 | 1.618 | 14.60 | 0.852 |
|  |  |  | SD | 4.85 | 0.649 | 2.365 | 2.979 | 26.76 | 0.750 |
|  | Mollusc, Crustacean | 11 | Mean | 2.56 | 0.731 | 0.882 | 1.614 | 5.965 | 0.695 |
|  |  |  | SD | 2.62 | 1.012 | 1.226 | 2.222 | 7.795 | 0.832 |
|  | Fish | 23 | Mean | 3.19 | 0.336 | 1.196 | 1.532 | 13.78 | 0.881 |
|  |  |  | SD | 4.21 | 0.473 | 1.835 | 2.283 | 20.58 | 0.676 |
|  | Mollusc, Crustacean | 12 | Mean | 1.78 | 0.314 | 0.225 | 0.539 | 3.008 | 0.343 |
|  |  |  | SD | 1.54 | 0.399 | 0.221 | 0.571 | 3.259 | 0.144 |
| $\begin{aligned} & \text { ㄷ } \\ & \text { of } \\ & \text { ㅇ } \end{aligned}$ | Fish | 23 | Mean | 3.91 | 0.246 | 0.884 | 1.130 | 12.22 | 0.809 |
|  |  |  | SD | 4.73 | 0.291 | 0.992 | 1.243 | 14.57 | 0.626 |
|  | Mollusc, Crustacean | 10 | Mean | 1.22 | 0.161 | 0.226 | 0.387 | 2.061 | 0.319 |
|  |  |  | SD | 0.69 | 0.209 | 0.228 | 0.432 | 1.648 | 0.102 |

FW: fresh weight. a: Nb composite samples. Each sample is composed by 5 subsamples of the same species, representative of the provisioning methods in each zone (port, market, supermarket...).

# FOURTH PART 

## Nutritionalintakes and

 exposure to contaminants
### 4.1 Fatty acid intakes

### 4.1.1 Food exposure

Fatty acid intakes through fish and seafood consumption for the four study zones are presented in Tables 29 to 33.

Note that these results correspond to fatty acid intakes only through fish and seafood, not intakes through the total diet. However, as mentioned in the introductory section of this report, leaving aside consumption of dietary supplements, marine products are the main source of long-chain polyunsaturated omega 3 fatty acids given that the conversion of the precursor ALA is very low (less than $1 \%$, see "Methodology and General Presentation").

For the four study zones, the intakes of long-chain polyunsaturated omega 3 are lower but of the same order of magnitude as the estimated intakes of the Inuit and Japanese populations, both high seafood consumers ${ }^{151116}$. Average EPA intakes are 419 to $517 \mathrm{mg} /$ day for adult males and 403 to $509 \mathrm{mg} / \mathrm{day}$ for adult females; DHA intakes are 739 to $960 \mathrm{mg} / \mathrm{day}$ for men and 713 to $885 \mathrm{mg} /$ day for women. The variability between individuals is also consistent with the results found in the literature.

Compared to intakes of the general French population through seafood ${ }^{117}$, the PUFA intakes of the subjects of this study are 4.1 times higher in adult males and 4.2 times higher in women. More than half of these PUFAs are LC-PUFAs of the omega 3 family, EPA, DPA and DHA.

The RDAs of long-chain PUFAs, in particular DHA, are well covered ( $786 \pm 612 \mathrm{mg}$ DHA/day on average for an RDA of 100 to $120 \mathrm{mg} /$ day), regardless of the age and sex and notably in adult females and women of child-bearing age.

As regards women of child-bearing age ( 18 to 44 years), in the four study zones fish and seafood consumption alone provides average intakes largely exceeding the RDA of LC-PUFAs and DHA for adult females and pregnant women.

Generally speaking, in all the zones, and whatever the age group and sex considered, the average EPA+DHA intakes exceed the $1 \mathrm{~g} /$ day recommended by the American Heart Association, yet they generally remain below the maximum limit of EPA+DHA intake of $2 \mathrm{~g} /$ day according to the AFSSA in 2003 ${ }^{10}$; $14 \%$ of our subjects exceed this recommendation - through fish and seafood consumption alone. However, given the rarity of available data, this limit is not considered to be an intake beyond which health risk can appear, but rather an intake beyond which there is no proven nutritional benefit.

The statistical analyses reveal that the subjects in Le Havre, regardless of age and sex, have an estimated EPA intake through their fish and seafood consumption higher than subjects in Lorient and La Rochelle, and a DHA intake higher than subjects in La Rochelle and Toulon (Table 33).

Moreover, within a given zone there are no clear disparities in terms of n-3 LC-PUFA intakes between the different age groups and sexes (results not presented), apart from Lorient where men aged 18 to 64 have EPA, DHA, PUFA and omega 3 intakes significantly higher than women in the same age group ( $p<0.05$ ).

[^22]Elderly subjects ( $\geq 65$ years) in Le Havre have relatively higher intakes than those in other zones, with average EPA and DHA levels of 693 and $1,164 \mathrm{mg} /$ day respectively compared to a maximum of 416 mg EPA and 770 mg DHA per day in the other zones. However these regional differences are not statistically significant. This particularity of elderly people in Le Havre can no doubt be explained by their very high consumption of herring (excluding smoked herring) of $129.3 \mathrm{~g} / \mathrm{wee}$. According to the literature this fish is one of the most oily ( 8.5 to 12.3 g of lipids for $100 \mathrm{~g}{ }^{104} ; 17.8 \mathrm{~g}$ of lipids for 100 g including 2.8 g of n-3 LC-PUFA ${ }^{105}$ ).

According to the zone and the group of individuals considered, consumption of fish and seafood products provides $3.3 \%$ to $5.8 \%$ of the RDA of the omega 3 precursor (ALA), $6.4 \%$ to $10 \%$ of the RDA of saturated fatty acids and $2.6 \%$ to $4.2 \%$ of the RDA of monounsaturated fatty acids.

The Appendix 5 shows that the main contributors to omega 3 exposure are salmon ( $27 \%$ ), mackerel ( $12 \%$ ), sardine ( $10 \%$ ), anchovy and herring (about 5\%). The consumption of salmon contribute on average to $33 \%$ of the recommended daily intake of EPA and DHA, mackerel to $28 \%$ and sardine to 24\% (Appendix 6).

Indeed mackerel, sardine and salmon are major contributors in all four study zones, providing respectively $7 \%$ to $16 \%, 6 \%$ to $17 \%$ and $24 \%$ to $31 \%$ of the intake. Herring, an other oily fish, account for at least $5 \%$ of the intake only in Le Havre and La Rochelle, while anchovy contributes to omega 3 intake in Lorient and Toulon.
Table 29: Dietary intakes of fatty acids from fish and other seafood - Le Havre (mg/d)

Table 30: Dietary intakes of fatty acids from fish and other seafood - Lorient (mg/d)

|  | $\begin{gathered} \text { Adult } 917060-64 \mathrm{y} 2 \\ \mathrm{n}=52 \end{gathered}$ |  |  | $\begin{gathered} 57772 \text { women }(18-64 y) \\ n=158 \end{gathered}$ |  |  | Older subjects ( 65 y and more)$\mathrm{n}=37$ |  |  | Women of childbearing age (18-44 y)$\mathrm{n}=76$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Me87 | 12 | 725 |
| Lip22 742d) | 6.76 | $2=2$ | 14.51 | 123 | 1.62 | 12.10 | c. 09 | 2.l\| 6 | (A) | 5.28 | 7.52 | 1295 |
| F283108id |  |  |  |  |  |  |  |  |  |  |  |  |
| C162 | 3 | 1 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| C14:0 | 158 | 96 | 369 | 123 | 104 | 271 | 131 | 92 | 286 | 127 | 120 | 342 |
| C14:1 n -5 | 2 | 0 | 5 | 1 | 1 | 2 | 1 | 1 | 2 | 1 | 1 | 1 |
| C15:0 | 7 | 3 | 16 | 5 | 3 | 12 | 5 | 3 | 11 | 1 | 0 | 46 |
| 016:0 | 2237 | 255 | 2,495 | 870 | 594 | 1,992 | 87 | 514 | 1,630 | 857 | 661 | 2,020 |
| C16:1 n -7 | 197 | 99 | 417 | 157 | 115 | 340 | 160 | 99 | 288 | 146 | 123 | 332 |
| C18:0 | 393 | 66 | 914 | 290 | 198 | 618 | 265 | 136 | 48 | 408 | 724 | 679 |
| 11611 SFAns | 77 | 9 | 745 | 3 | 133 | 1/3 | 52 | 90 | 19\% | 05 | 74 | 34 |
| C1622 174.9 | 1,276 | 1,172 | 2,626 | 1,031 | 849 | 2,753 | 919 | 688 | 1,899 | 1,081 | 943 | 3,015 |
|  | 41 | 2 | 100 | 26 | 23 | 74 | 24 | 19 | 61 | 27 | 22 | 78 |
| C18:2n-6 (LA) | 315 | 298 | 715 | 261 | 232 | 649 | 19 m | enc | 183 | 280 | ) Ad | ult |
| C18:3n-6 | 6 | 5 | 14 | 5 | 5 | 13 | 4 | 3 | 9 | 4 | 5 | 12 |
| C18:3n-3 (ALA) | 87 | 34 | 223 | 64 | 61 | 155 | 50 | 32 | 99 | 73 | 76 | 178 |
| C18:4 n -3 | 263 | 66 | 722 | 187 | 131 | 469 | 17 | 116 | 345 | 180 | 130 | 489 |
| 10010 | 30 | 1 a | ns | DP | 95 | 72 | 31 | ds | (9) | 17 | 28 | 78 |
| 43n13n-6 | 5 | 0 | 5. | 3 | 0 | 30 | 2 | 3 | 9 | 1 | - | 89 |
| 510:4n-3914 | 183 | 34 | -11 | 121 | 123 | att | yac | 55 | 202 | :00 | 039 | 278 |
| C20:5 n-3 (EPA) | 517 | 247 | 1,155 | 403 | 303 | 889 | 406 | 226 | 743 | 389 | 347 | 862 |
| C22:5 n-3 (DPA) | 148 | 53 | 512 | 114 | 173 | 274 | 431 | 1 | 189 | 11/ | 138 | 561 |
| C22:6 n-3 (DHA) | 960 | 421 | 2,235 | 733 | 559 | 1,779 | 770 | 458 | 1285 | 24 | 537 | C, $8=1$ |
| tra | 1,471 | 112 | 4,933 | 1341 | 626 | 3,311 | 4.839 | 78= | 10sn | 1,332 | 1,003 | 3,042 |
| MUFA | 1,602 | 1,299 | 2,932 | 1,289 | 967 | 3,462 | 1,176 | 799 | 2,202 | 1,310 | 1,078 | 3,759 |
| PUFA | 2,514 | 1,260 | 5,853 | 1,927 | 1,440 | 4,269 | 1,840 | 1,073 | 3,290 | 1,939 | 1,701 | 4,506 |
| omega 3 | 1,975 | 822 | 4,587 | 1,502 | 1,164 | 3,439 | 1,506 | 872 | 2,732 | 1,488 | 1,372 | 3,411 |
| omega 6 | 508 | 339 | 1,002 | 389 | 319 | 983 | 313 | 214 | 739 | 410 | 368 | 120= |

Table 31: Dietary intakes of fatty acids from fish and other seafood - La Rochelle (mg/d)

Table 32: Dietary intakes of fatty acids from fish and other seafood - Toulon (mg/d)

|  | $\begin{aligned} & \text { Adult men ( } 18-64 \mathrm{y}) \\ & \mathrm{n}=52 \end{aligned}$ |  |  | Adult women (18-64 y)$\mathrm{n}=158$ |  |  | Older subjects ( 65 y and more)$n=37$ |  |  | Women of childbearing age (18-44y)$n=76$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 |
| Lipids (g/d) | 6.49 | 4.32 | 14.69 | 6.28 | 4.57 | 15.43 | 5.04 | 2.49 | 924 | 6.40 | 5.00 | 15.61 |
| Fatty add |  |  |  |  |  |  |  |  |  |  |  |  |
| C12:0 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 1 | 2 |
| C14:0 | 146 | 117 | 403 | 152 | 137 | 427 | 127 | 85 | 311 | 150 | 129 | 427 |
| C14:1 n -5 | 3 | 3 | 9 | 4 | 4 | 10 | 3 | 3 | 9 | 4 | 4 | 10 |
| C15:0 | 7 | 7 | 22 | 7 | 6 | 20 | 5 | 4 | 12 | 7 | 6 | 17 |
| C16:0 | 972 | 644 | 2,208 | 927 | 664 | 2,316 | 814 | 394 | 1,505 | 918 | 682 | 2,316 |
| C16:1 n-7 | 189 | 137 | 490 | 191 | 160 | 498 | 174 | 106 | 374 | 181 | 157 | 473 |
| C18:0 | 375 | 248 | 810 | 319 | 230 | 742 | 254 | 129 | 456 | 331 | 267 | 818 |
| C18:1 trans | 12 | 12 | 35 | 15 | 14 | 45 | 10 | 9 | 29 | 14 | 13 | 43 |
| C18:1 1 is $\mathrm{n}-9$ | 1,349 | 934 | 3,162 | 1,376 | 1,386 | 3,686 | 939 | 657 | 1,980 | 1,511 | 1,707 | 4,376 |
| C18:1 1 is $\mathrm{n}-7$ | 51 | 54 | 152 | 42 | 43 | 114 | 26 | 23 | 75 | 46 | 47 | 132 |
| C18:2n-6 (LA) | 391 | 266 | 851 | 368 | 384 | 959 | 244 | 188 | 515 | 402 | 470 | 1,228 |
| C18:3n-6 | 6 | 7 | 22 | 7 | 9 | 23 | 4 | 4 | 14 | 8 | 10 | 25 |
| C18:3 n -3 (ALA) | 90 | 67 | 212 | 83 | 63 | 202 | 55 | 30 | 98 | 88 | 72 | 208 |
| C18:4n-3 | 207 | 175 | 565 | 223 | 205 | 634 | 185 | 138 | 400 | 218 | 200 | 604 |
| C20:0 | 25 | 36 | 78 | 22 | 31 | 78 | 20 | 16 | 45 | 20 | 32 | 73 |
| C20:2n-6 | 2 | 4 | 12 | 2 | 5 | 10 | 1 | 1 | 3 | 3 | 7 | 11 |
| C20:4n-6 (AA) | 144 | 123 | 398 | 127 | 101 | 313 | 102 | 48 | 172 | 126 | 113 | 332 |
| C20:5 n-3 (EPA) | 467 | 354 | 1,374 | 433 | 326 | 1,237 | 388 | 194 | 650 | 407 | 327 | 1,133 |
| C22:5 n-3 (DPA) | 145 | 104 | 375 | 136 | 102 | 313 | 135 | 86 | 240 | 127 | 104 | 319 |
| C22:6 n-3 (DHA) | 750 | 514 | 1,564 | 713 | 517 | 1,648 | 686 | 336 | 1,074 | 678 | 541 | 1,648 |
| SFA | 1,674 | 1,129 | 4,076 | 1,604 | 1,97 | 4,046 | 1,353 | 674 | 2,345 | 1,599 | 1,225 | 4,043 |
| MUFA | 1,645 | 1,123 | 3,684 | 1,669 | 1,524 | 4,477 | 1,192 | 773 | 2,470 | 1,796 | 1,831 | 4,747 |
| PUFA | 2,249 | 1,565 | 4,778 | 2,149 | 1,569 | 5,118 | 1,819 | 882 | 3,147 | 2,127 | 1,685 | 5,396 |
| omega 3 | 1,659 | 1,184 | 3,549 | 1,588 | 1,166 | 3,827 | 1,449 | 689 | 2,544 | 1,517 | 1,196 | 3,820 |
| omega 6 | 544 | 378 | 1,193 | 504 | 454 | 1,216 | 351 | 224 | 734 | 538 | 549 | 1,576 |

Table 33: Dietary intakes of fatty acids from fish and other seafood per site (mg/d, Mean $\pm$ SD)

|  | Le Havre $n=249$ | Lorient $n=247$ | La Rochelle $n=248$ | Toulon $\mathrm{n}=252$ | All subjects $\mathrm{n}=996$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| EPA | $516 \pm 454{ }^{\text {a }}$ | $428 \pm 304^{\text {b }}$ | $428 \pm 316^{\text {b }}$ | $437 \pm 324$ a.b | $452 \pm 356$ |
| DPA | $141 \pm 142^{\text {a }}$ | $120 \pm 153{ }^{\text {a }}$ | $111 \pm 124^{\text {a }}$ | $138 \pm 101^{\text {a }}$ | $127 \pm 132$ |
| DHA | $896 \pm 800{ }^{\text {a }}$ | $786 \pm 568{ }^{\text {a. b }}$ | $743 \pm 522^{\text {b }}$ | $720 \pm 502{ }^{\text {b }}$ | $786 \pm 612$ |
| Omega 3* | $1,814 \pm 1.596{ }^{\text {a }}$ | $1,602 \pm 1,177{ }^{\text {a }}$ | $1,521 \pm 1,096$ a | $1,594 \pm 1,136$ a | $1,633 \pm 1,270$ |
| PUFA | $2,248 \pm 1.937{ }^{\text {a }}$ | $2,037 \pm 1,452^{\text {a }}$ | 1,909 $\pm 1,293{ }^{\text {a }}$ | $2,145 \pm 1,522{ }^{\text {a }}$ | $2,085 \pm 1,572$ |

* The intake of Omega 3 correspond to the intakes of ALA, C18:4 n-3, EPA, DPA and DHA

Values in the same raw with different superscript letters are significantly different, $\mathrm{p}<0.05$ (Tukey's test)

### 4.1.2 Biomarker of exposure

The results of the erythrocyte fatty acid profile of the 391 subjects in the study are presented in Tables 34 to 38.

Unlike the food exposure results obtained by crossing food composition and individual consumption data, the results of the direct approach yield the fatty acid biological level resulting from the total diet.

The results obtained suggest that adult males have the most triglycerides in the blood, but with high variability ( $0.9 \mathrm{~g} / \mathrm{L} \pm 0.6$ to $1.6 \mathrm{~g} / \mathrm{L} \pm 2.3$ depending on the zone) compared to women and to males in other age groups, the norm being 0.5 to $2 \mathrm{~g} / \mathrm{L}^{118}$. On the other hand, elderly subjects ( $\geq 65$ years) have the highest total cholesterol ( 2.15 to $2.48 \mathrm{~g} / \mathrm{L}$ ), HDL-cholesterol ( 0.62 to $0.69 \mathrm{~g} / \mathrm{L}$ ) and LDL-cholesterol ( 1.35 to $1.66 \mathrm{~g} / \mathrm{L}$ ). $39 \%$ of the subjects exceeds the norm of the total cholesterol level, fixed at 2.0 to $2.6 \mathrm{~g} / \mathrm{L}$ depending on age. Generally, only women of child-bearing age ( 18 to 44 years) have average total cholesterol levels corresponding to the norm for their age.

The lipidic fraction of the erythrocyte membrane is constituted by $50 \%$ of SFA on average in adult males and $47 \%$ in women and elderly people. MUFAs represent about $18 \%$ of total lipids, and PUFAs $27 \%$ for omega 6 and $7 \%$ for omega 3. The EPA, DPA and DHA account for almost all the omega 3 in the erythrocyte membrane, the precursor ALA representing only $0.2 \%$ of total lipids on average. The omega 6 precursor (LA) constitutes only about $11 \%$ of the membrane lipidic fraction.

Table 34: Fatty acid composition of red blood cells - Le Havre (\% total FA)

|  | Adult men$\begin{gathered} (18-64 y) \\ n=18 \end{gathered}$ |  |  | $\begin{aligned} & \text { Adult women } \\ & \qquad \begin{array}{c} (18-64 y) \\ n=60 \end{array} \end{aligned}$ |  |  | $\begin{aligned} & \text { Older subjects } \\ & \text { (65y and more) } \\ & n=7 \end{aligned}$ |  |  | Women of childbearing age$\begin{gathered} (18-44 y) \\ n=29 \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 |
| Triglycerides (g/L) | 0.90 | 0.63 | 1.86 | 0.86 | 0.43 | 1.85 | 0.81 | 0.20 | 0.99 | 0.78 | 0.32 | 1.39 |
| Total cholesterol (g/L) | 1.99 | 0.37 | 2.55 | 2.00 | 0.39 | 2.62 | 2.15 | 0.44 | 2.61 | 1.82 | 0.33 | 2.44 |
| HDL (g/L) | 0.57 | 0.10 | 0.73 | 0.66 | 0.20 | 0.91 | 0.64 | 0.13 | 0.81 | 0.61 | 0.24 | 0.83 |
| Cholesterol/HDL | 3.61 | 0.95 | 5.01 | 3.25 | 0.86 | 4.71 | 3.46 | 0.86 | 4.79 | 3.29 | 0.91 | 4.76 |
| LDL (g/L) | 1.24 | 0.32 | 1.67 | 1.18 | 0.33 | 1.71 | 1.35 | 0.39 | 1.75 | 1.09 | 0.30 | 1.66 |
| Fatty acids (\% total lipids) |  |  |  |  |  |  |  |  |  |  |  |  |
| C12:0 | 0.28 | 0.13 | 0.49 | 0.32 | 0.19 | 0.59 | 0.30 | 0.18 | 0.59 | 0.28 | 0.13 | 0.50 |
| C14:0 | 0.84 | 0.24 | 1.27 | 0.77 | 0.25 | 1.27 | 0.90 | 0.35 | 1.37 | 0.78 | 0.24 | 1.25 |
| C14:1 n-5 | 0.12 | 0.05 | 0.21 | 0.12 | 0.05 | 0.19 | 0.17 | 0.12 | 0.36 | 0.12 | 0.05 | 0.22 |
| C15:0 | 0.25 | 0.11 | 0.44 | 0.26 | 0.06 | 0.37 | 0.32 | 0.10 | 0.46 | 0.27 | 0.06 | 0.38 |
| C16:0 | 20.58 | 1.36 | 22.67 | 20.29 | 1.04 | 21.99 | 20.86 | 0.97 | 21.91 | 20.55 | 1.14 | 22.27 |
| C16:1 n-9 | 0.36 | 0.14 | 0.61 | 0.33 | 0.18 | 0.77 | 0.38 | 0.25 | 0.77 | 0.36 | 0.18 | 0.77 |
| C16:1 n-7 | 1.25 | 0.53 | 2.04 | 1.15 | 0.46 | 1.87 | 1.48 | 0.54 | 2.14 | 1.19 | 0.44 | 1.85 |
| C18:0 | 23.69 | 6.85 | 36.29 | 24.02 | 4.72 | 30.20 | 23.09 | 4.96 | 29.41 | 23.59 | 4.75 | 31.05 |
| C18:1 n-9t | 0.40 | 0.20 | 0.72 | 0.30 | 0.18 | 0.62 | 0.37 | 0.16 | 0.56 | 0.30 | 0.17 | 0.62 |
| C18:1 n-9 | 15.09 | 3.13 | 19.19 | 14.56 | 2.18 | 18.49 | 15.83 | 2.09 | 18.17 | 14.74 | 2.27 | 18.34 |
| C18:1 n-7 | 1.17 | 0.21 | 1.43 | 1.20 | 0.16 | 1.55 | 1.29 | 0.17 | 1.50 | 1.22 | 0.15 | 1.52 |
| C18:2 n-6 (LA) | 14.22 | 5.50 | 21.22 | 12.90 | 4.53 | 20.62 | 12.99 | 3.84 | 17.65 | 13.42 | 4.84 | 20.54 |
| C18:3 n-6 | 0.25 | 0.12 | 0.39 | 0.19 | 0.10 | 0.34 | 0.26 | 0.09 | 0.34 | 0.20 | 0.10 | 0.35 |
| C18:3 n-3 (ALA) | 0.25 | 0.11 | 0.41 | 0.25 | 0.11 | 0.45 | 0.26 | 0.10 | 0.41 | 0.23 | 0.08 | 0.35 |
| C20:0 | 0.19 | 0.07 | 0.33 | 0.18 | 0.08 | 0.31 | 0.23 | 0.07 | 0.31 | 0.19 | 0.08 | 0.32 |
| C20:2 n-6 | 0.31 | 0.06 | 0.41 | 0.34 | 0.05 | 0.40 | 0.34 | 0.05 | 0.40 | 0.34 | 0.06 | 0.42 |
| C20:3 n-6 | 1.28 | 0.25 | 1.60 | 1.42 | 0.32 | 2.02 | 1.39 | 0.21 | 1.60 | 1.51 | 0.33 | 2.08 |
| C20:4 n-6 (AA) | 11.00 | 1.61 | 13.91 | 12.13 | 1.98 | 14.69 | 10.58 | 2.47 | 13.92 | 12.04 | 2.17 | 15.91 |
| C20:5 n-3 (EPA) | 0.87 | 0.53 | 1.49 | 0.79 | 0.41 | 1.71 | 1.12 | 0.53 | 1.95 | 0.58 | 0.21 | 1.05 |
| C22:4 n-6 | 1.87 | 0.63 | 2.78 | 2.02 | 0.69 | 3.06 | 1.58 | 0.56 | 2.25 | 2.12 | 0.69 | 3.35 |
| C22:5 n-6 | 0.39 | 0.09 | 0.50 | 0.43 | 0.16 | 0.75 | 0.37 | 0.10 | 0.51 | 0.49 | 0.16 | 0.75 |
| C22:5 n-3 (DPA) | 1.70 | 0.40 | 2.37 | 1.74 | 0.38 | 2.33 | 1.61 | 0.55 | 2.26 | 1.57 | 0.31 | 2.06 |
| C22:6 n-3 (DHA) | 3.64 | 1.05 | 5.25 | 4.29 | 1.17 | 6.66 | 4.27 | 1.24 | 5.60 | 3.90 | 1.02 | 5.30 |
| SFA | 45.83 | 7.56 | 59.54 | 45.85 | 4.48 | 53.07 | 45.71 | 5.53 | 52.61 | 45.67 | 4.49 | 53.14 |
| MUFA | 17.99 | 3.74 | 23.36 | 17.36 | 2.73 | 22.38 | 19.14 | 2.67 | 21.78 | 17.63 | 2.77 | 22.34 |
| omega 6 | 29.32 | 4.79 | 35.50 | 29.43 | 3.51 | 34.52 | 27.51 | 3.21 | 31.94 | 30.13 | 3.40 | 34.46 |
| omega 3 | 6.46 | 1.53 | 8.83 | 7.06 | 1.67 | 10.07 | 7.27 | 1.97 | 9.30 | 6.27 | 1.23 | 7.82 |
| omega 6/omega 3 | 4.82 | 1.59 | 7.90 | 4.44 | 1.30 | 6.74 | 4.00 | 1.03 | 5.32 | 5.04 | 1.36 | 7.05 |
| LA/ALA | 70.20 | 59.14 | 214.40 | 58.19 | 23.71 | 99.77 | 52.89 | 12.38 | 70.52 | 63.33 | 28.05 | 119.21 |

Table 35: Fatty acid composition of red blood cells - Lorient (\% total FA)

|  | Adult men$\begin{gathered} (18-64 y) \\ n=21 \end{gathered}$ |  |  | Adult women$\begin{gathered} (18-64 y) \\ n=84 \end{gathered}$ |  |  | $\begin{aligned} & \text { Older subjects } \\ & \text { (65y and more) } \\ & n=10 \end{aligned}$ |  |  | Women of childbearing age$\begin{gathered} (18-44 y) \\ n=39 \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 |
| Triglycerides ( $\mathrm{g} / \mathrm{L}$ ) | 1.28 | 0.67 | 2.62 | 0.90 | 0.49 | 1.77 | 1.01 | 0.47 | 1.75 | 0.85 | 0.42 | 1.70 |
| Total cholesterol (g/L) | 2.10 | 0.46 | 2.81 | 2.21 | 0.41 | 2.75 | 2.48 | 0.44 | 3.17 | 2.11 | 0.42 | 2.76 |
| HDL (g/L) | 0.50 | 0.10 | 0.64 | 0.61 | 0.13 | 0.84 | 0.66 | 0.20 | 0.89 | 0.59 | 0.13 | 0.79 |
| Cholesterol/HDL | 4.37 | 1.19 | 6.20 | 3.73 | 0.92 | 5.32 | 4.07 | 1.21 | 5.91 | 3.69 | 0.98 | 5.13 |
| LDL (g/L) | 1.39 | 0.41 | 1.95 | 1.44 | 0.35 | 2.00 | 1.66 | 0.39 | 2.21 | 1.37 | 0.36 | 1.88 |
| Fatty acids (\% total lipids) |  |  |  |  |  |  |  |  |  |  |  |  |
| C12:0 | 0.87 | 0.49 | 1.24 | 0.59 | 0.34 | 1.11 | 0.82 | 0.46 | 1.58 | 0.60 | 0.38 | 1.13 |
| C14:0 | 1.16 | 0.50 | 1.80 | 0.97 | 0.40 | 1.67 | 1.04 | 0.43 | 1.46 | 0.96 | 0.44 | 1.72 |
| C14:1 n-5 | 0.39 | 0.16 | 0.65 | 0.34 | 0.18 | 0.64 | 0.32 | 0.15 | 0.54 | 0.33 | 0.18 | 0.62 |
| C15:0 | 0.80 | 0.33 | 1.38 | 0.67 | 0.30 | 1.18 | 0.66 | 0.26 | 0.98 | 0.65 | 0.32 | 1.26 |
| C16:0 | 21.79 | 2.65 | 25.72 | 20.40 | 2.26 | 24.76 | 20.39 | 1.70 | 23.23 | 20.58 | 2.33 | 24.07 |
| C16:1 n-9 | 0.39 | 0.16 | 0.67 | 0.38 | 0.19 | 0.82 | 0.45 | 0.26 | 0.81 | 0.41 | 0.20 | 0.82 |
| C16:1 n-7 | 1.33 | 0.46 | 2.04 | 1.34 | 0.60 | 2.37 | 1.37 | 0.51 | 2.12 | 1.31 | 0.57 | 2.23 |
| C18:0 | 25.44 | 5.33 | 32.03 | 22.81 | 5.81 | 32.55 | 21.76 | 4.72 | 27.49 | 23.26 | 5.88 | 32.76 |
| C18:1 n-9t | 0.40 | 0.41 | 0.97 | 0.42 | 0.42 | 0.70 | 0.32 | 0.19 | 0.60 | 0.37 | 0.19 | 0.68 |
| C18:1 n-9 | 14.62 | 1.57 | 16.60 | 15.61 | 3.03 | 22.71 | 16.22 | 2.77 | 20.75 | 15.69 | 2.71 | 20.06 |
| C18:1 n -7 | 1.48 | 0.24 | 1.84 | 1.46 | 0.23 | 1.84 | 1.48 | 0.25 | 1.87 | 1.43 | 0.22 | 1.79 |
| C18:2 n-6 (LA) | 10.82 | 2.20 | 13.63 | 13.36 | 4.17 | 21.11 | 12.83 | 4.17 | 20.00 | 13.27 | 3.84 | 20.14 |
| C18:3 n-6 | 0.21 | 0.10 | 0.39 | 0.21 | 0.09 | 0.36 | 0.20 | 0.09 | 0.35 | 0.21 | 0.09 | 0.37 |
| C18:3 n-3 (ALA) | 0.27 | 0.15 | 0.53 | 0.25 | 0.14 | 0.53 | 0.25 | 0.12 | 0.40 | 0.23 | 0.14 | 0.50 |
| C20:0 | 0.20 | 0.10 | 0.38 | 0.26 | 0.12 | 0.45 | 0.25 | 0.12 | 0.40 | 0.26 | 0.14 | 0.45 |
| C20:2 n-6 | 0.61 | 0.29 | 1.01 | 0.48 | 0.23 | 0.95 | 0.41 | 0.23 | 0.72 | 0.50 | 0.22 | 0.94 |
| C20:3 n-6 | 1.26 | 0.49 | 2.00 | 1.28 | 0.45 | 2.18 | 1.34 | 0.34 | 1.82 | 1.34 | 0.49 | 2.18 |
| C20:4 n-6 (AA) | 10.20 | 3.37 | 14.56 | 10.73 | 2.78 | 14.52 | 11.00 | 3.02 | 14.58 | 10.68 | 3.01 | 14.27 |
| C20:5 n-3 (EPA) | 0.64 | 0.33 | 1.12 | 0.75 | 0.34 | 1.27 | 0.82 | 0.41 | 1.41 | 0.59 | 0.29 | 1.09 |
| C22:4 n-6 | 1.87 | 0.76 | 3.16 | 1.76 | 0.65 | 2.87 | 1.67 | 0.63 | 2.45 | 1.86 | 0.70 | 3.02 |
| C22:5 n-6 | 0.66 | 0.28 | 1.07 | 0.54 | 0.24 | 0.90 | 0.43 | 0.26 | 0.83 | 0.56 | 0.22 | 0.91 |
| C22:5 n-3 (DPA) | 1.58 | 0.76 | 2.60 | 1.60 | 0.56 | 2.45 | 2.03 | 0.97 | 3.36 | 1.49 | 0.60 | 2.45 |
| C22:6 n-3 (DHA) | 3.01 | 1.79 | 6.59 | 3.78 | 1.43 | 6.08 | 3.93 | 1.40 | 5.74 | 3.41 | 1.30 | 5.26 |
| SFA | 50.26 | 7.69 | 62.66 | 45.70 | 7.01 | 55.93 | 44.92 | 5.87 | 53.59 | 46.31 | 7.43 | 57.11 |
| MUFA | 18.21 | 1.65 | 20.89 | 19.13 | 3.51 | 27.47 | 19.85 | 3.51 | 25.56 | 19.16 | 3.08 | 22.89 |
| omega 6 | 25.63 | 5.27 | 31.52 | 28.36 | 4.81 | 36.24 | 27.88 | 3.92 | 32.92 | 28.43 | 5.20 | 36.30 |
| omega 3 | 5.49 | 2.55 | 10.10 | 6.38 | 2.01 | 9.31 | 7.03 | 1.97 | 9.22 | 5.73 | 1.87 | 8.34 |
| omega 6/omega 3 | 5.47 | 2.01 | 8.70 | 4.91 | 1.78 | 8.78 | 4.40 | 1.79 | 7.24 | 5.49 | 1.90 | 8.93 |
| LA/ALA | 52.99 | 29.23 | 103.25 | 68.64 | 48.30 | 130.85 | 63.72 | 34.29 | 121.65 | 76.54 | 61.03 | 132.87 |

Table 36: Fatty acid composition of red blood cells - La Rochelle (\% total FA)

|  | Adult men$\begin{gathered} (18-64 y) \\ n=37 \end{gathered}$ |  |  | $\begin{aligned} & \text { Adult women } \\ & \begin{array}{c} (18-64 y) \\ n=46 \end{array} \end{aligned}$ |  |  | $\begin{aligned} & \text { Older subjects } \\ & \text { (65y and more) } \\ & n=14 \end{aligned}$ |  |  | Women of childbearing age$\begin{gathered} (18-44 y) \\ n=28 \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 |
| Triglycerides (g/L) | 1.62 | 2.30 | 3.21 | 0.90 | 0.37 | 1.60 | 1.18 | 0.55 | 2.18 | 0.86 | 0.36 | 1.26 |
| Total cholesterol (g/L) | 2.14 | 0.32 | 2.79 | 2.09 | 0.38 | 2.71 | 2.29 | 0.35 | 2.80 | 1.99 | 0.35 | 2.57 |
| HDL (g/L) | 0.54 | 0.13 | 0.73 | 0.68 | 0.16 | 0.96 | 0.69 | 0.19 | 1.03 | 0.67 | 0.17 | 1.02 |
| Cholesterol/HDL | 4.20 | 1.28 | 6.71 | 3.28 | 1.02 | 4.75 | 3.48 | 0.81 | 4.43 | 3.15 | 0.88 | 4.56 |
| LDL ( $\mathrm{g} / \mathrm{L}$ ) | 1.34 | 0.29 | 1.95 | 1.25 | 0.39 | 1.81 | 1.36 | 0.27 | 1.67 | 1.17 | 0.39 | 1.71 |
| Fatty acids (\% total lipids) |  |  |  |  |  |  |  |  |  |  |  |  |
| C12:0 | 0.57 | 0.39 | 1.20 | 0.55 | 0.36 | 1.17 | 0.57 | 0.37 | 1.01 | 0.57 | 0.36 | 1.20 |
| C14:0 | 0.84 | 0.25 | 1.23 | 0.84 | 0.30 | 1.30 | 0.85 | 0.40 | 1.49 | 0.85 | 0.30 | 1.29 |
| C14:1 n-5 | 0.15 | 0.11 | 0.41 | 0.18 | 0.12 | 0.42 | 0.19 | 0.19 | 0.58 | 0.20 | 0.14 | 0.43 |
| C15:0 | 0.26 | 0.17 | 0.78 | 0.22 | 0.09 | 0.30 | 0.23 | 0.10 | 0.41 | 0.23 | 0.10 | 0.30 |
| C16:0 | 20.20 | 1.40 | 22.52 | 20.59 | 2.13 | 24.37 | 20.39 | 1.65 | 22.76 | 20.60 | 2.05 | 24.25 |
| C16:1 n-9 | 0.23 | 0.13 | 0.35 | 0.24 | 0.10 | 0.38 | 0.25 | 0.15 | 0.52 | 0.23 | 0.11 | 0.39 |
| C16:1 n-7 | 0.64 | 0.27 | 1.17 | 0.71 | 0.25 | 1.14 | 0.89 | 0.49 | 1.91 | 0.67 | 0.26 | 0.98 |
| C18:0 | 28.16 | 8.07 | 42.53 | 28.29 | 7.86 | 39.35 | 25.54 | 7.17 | 34.18 | 28.55 | 8.26 | 39.31 |
| C18:1 n-9t | 0.36 | 0.14 | 0.67 | 0.34 | 0.10 | 0.48 | 0.34 | 0.12 | 0.47 | 0.35 | 0.11 | 0.48 |
| C18:1 n-9 | 12.93 | 2.33 | 16.36 | 12.89 | 2.70 | 16.36 | 15.13 | 4.43 | 24.35 | 12.54 | 2.77 | 16.33 |
| C18:1 n-7 | 1.07 | 0.19 | 1.40 | 1.12 | 0.19 | 1.41 | 1.24 | 0.24 | 1.71 | 1.11 | 0.20 | 1.39 |
| C18:2 n-6 (LA) | 8.65 | 2.27 | 12.21 | 8.62 | 1.97 | 11.15 | 9.82 | 2.96 | 14.93 | 8.64 | 1.96 | 11.14 |
| C18:3 n-6 | 0.16 | 0.07 | 0.26 | 0.18 | 0.10 | 0.36 | 0.14 | 0.09 | 0.28 | 0.18 | 0.11 | 0.36 |
| C18:3 n-3 (ALA) | 0.14 | 0.05 | 0.21 | 0.18 | 0.07 | 0.26 | 0.19 | 0.13 | 0.42 | 0.17 | 0.06 | 0.26 |
| C20:0 | 0.19 | 0.09 | 0.36 | 0.19 | 0.08 | 0.33 | 0.22 | 0.09 | 0.38 | 0.18 | 0.08 | 0.30 |
| C20:2 n-6 | 0.38 | 0.16 | 0.65 | 0.36 | 0.18 | 0.63 | 0.33 | 0.14 | 0.47 | 0.35 | 0.19 | 0.63 |
| C20:3 n-6 | 1.61 | 0.48 | 2.32 | 1.37 | 0.40 | 1.92 | 1.33 | 0.43 | 2.07 | 1.39 | 0.43 | 1.90 |
| C20:4 n-6 (AA) | 12.98 | 2.86 | 17.56 | 12.77 | 3.39 | 17.39 | 12.15 | 3.77 | 16.27 | 12.84 | 3.44 | 17.39 |
| C20:5 n-3 (EPA) | 0.80 | 0.36 | 1.43 | 0.83 | 0.31 | 1.37 | 0.80 | 0.37 | 1.35 | 0.77 | 0.23 | 1.14 |
| C22:4 n-6 | 2.17 | 0.62 | 3.46 | 2.08 | 0.75 | 3.46 | 1.86 | 0.70 | 2.53 | 2.25 | 0.76 | 3.57 |
| C22:5 n-6 | 0.49 | 0.16 | 0.78 | 0.44 | 0.15 | 0.69 | 0.36 | 0.13 | 0.48 | 0.49 | 0.14 | 0.73 |
| C22:5 n-3 (DPA) | 2.14 | 0.60 | 2.98 | 1.89 | 0.59 | 2.92 | 1.97 | 0.65 | 2.82 | 1.83 | 0.50 | 2.67 |
| C22:6 n-3 (DHA) | 4.88 | 1.44 | 7.35 | 5.12 | 1.53 | 7.46 | 5.23 | 1.91 | 7.43 | 5.02 | 1.49 | 7.17 |
| SFA | 50.22 | 9.17 | 66.60 | 50.69 | 8.87 | 63.59 | 47.80 | 7.45 | 56.11 | 50.97 | 9.32 | 64.17 |
| MUFA | 15.02 | 2.62 | 18.71 | 15.14 | 2.97 | 18.99 | 17.69 | 5.31 | 28.92 | 14.75 | 3.00 | 18.95 |
| omega 6 | 26.43 | 5.45 | 33.68 | 25.81 | 5.77 | 33.74 | 25.98 | 4.85 | 32.34 | 26.14 | 6.05 | 34.26 |
| omega 3 | 7.96 | 2.26 | 11.67 | 8.02 | 2.23 | 11.54 | 8.18 | 2.60 | 11.00 | 7.79 | 2.12 | 10.99 |
| omega 6/omega 3 | 3.47 | 0.78 | 4.62 | 3.35 | 0.74 | 4.02 | 3.46 | 1.14 | 5.89 | 3.45 | 0.63 | 4.00 |
| LA/ALA | 68.50 | 34.57 | 131.20 | 53.17 | 22.55 | 95.93 | 64.21 | 30.39 | 119.10 | 59.08 | 26.43 | 103.42 |

Table 37: Fatty acid composition of red blood cells - Toulon (\% total FA)

|  | Adult men$\begin{gathered} (18-64 y) \\ n=17 \end{gathered}$ |  |  | Adult women$\begin{gathered} (18-64 y) \\ n=69 \end{gathered}$ |  |  | $\begin{aligned} & \text { Older subjects } \\ & \text { (65y and more) } \\ & n=9 \end{aligned}$ |  |  | Women of childbearing age$\begin{gathered} (18-44 y) \\ n=41 \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 |
| Triglycerides (g/L) | 1.26 | 0.59 | 2.06 | 0.88 | 0.60 | 1.88 | 1.00 | 0.35 | 1.56 | 0.90 | 0.66 | 2.04 |
| Total cholesterol (g/L) | 2.12 | 0.35 | 2.73 | 2.03 | 0.38 | 2.62 | 2.37 | 0.54 | 3.22 | 1.90 | 0.35 | 2.44 |
| HDL (g/L) | 0.42 | 0.10 | 0.56 | 0.54 | 0.16 | 0.85 | 0.62 | 0.17 | 0.84 | 0.48 | 0.13 | 0.77 |
| Cholesterol/HDL | 5.40 | 1.79 | 9.14 | 4.01 | 1.13 | 6.15 | 3.91 | 0.72 | 4.88 | 4.15 | 1.21 | 6.20 |
| LDL ( $\mathrm{g} / \mathrm{L}$ ) | 1.45 | 0.32 | 1.97 | 1.32 | 0.32 | 1.84 | 1.55 | 0.37 | 2.09 | 1.24 | 0.30 | 1.75 |
| Fatty acids (\% total lipids) |  |  |  |  |  |  |  |  |  |  |  |  |
| C12:0 | 0.77 | 0.40 | 1.37 | 0.69 | 0.34 | 1.26 | 0.70 | 0.28 | 1.06 | 0.68 | 0.37 | 1.28 |
| C14:0 | 1.38 | 0.54 | 1.96 | 1.21 | 0.49 | 1.92 | 1.42 | 0.43 | 1.88 | 1.19 | 0.54 | 1.94 |
| C14:1 n-5 | 0.34 | 0.15 | 0.58 | 0.34 | 0.17 | 0.59 | 0.34 | 0.08 | 0.43 | 0.33 | 0.16 | 0.58 |
| C15:0 | 1.00 | 0.34 | 1.52 | 0.80 | 0.35 | 1.47 | 0.77 | 0.39 | 1.37 | 0.77 | 0.37 | 1.52 |
| C16:0 | 22.94 | 3.55 | 27.87 | 21.32 | 3.49 | 27.06 | 22.11 | 2.60 | 26.18 | 21.24 | 3.45 | 26.73 |
| C16:1 n-9 | 0.68 | 0.21 | 0.88 | 0.56 | 0.23 | 0.87 | 0.43 | 0.22 | 0.77 | 0.51 | 0.24 | 0.87 |
| C16:1 n -7 | 1.19 | 0.60 | 1.96 | 1.26 | 0.64 | 2.22 | 1.24 | 0.73 | 2.49 | 1.29 | 0.69 | 2.25 |
| C18:0 | 26.37 | 4.38 | 32.23 | 23.76 | 5.44 | 33.98 | 25.84 | 8.00 | 39.03 | 23.26 | 5.53 | 33.10 |
| C18:1 n-9t | 0.27 | 0.13 | 0.49 | 0.24 | 0.13 | 0.43 | 0.22 | 0.14 | 0.44 | 0.23 | 0.12 | 0.42 |
| C18:1 n-9 | 14.41 | 2.45 | 17.92 | 15.05 | 2.76 | 19.56 | 15.09 | 3.36 | 19.87 | 15.25 | 2.85 | 18.86 |
| C18:1 n -7 | 1.44 | 0.24 | 1.81 | 1.42 | 0.31 | 1.93 | 1.43 | 0.27 | 1.81 | 1.42 | 0.30 | 1.91 |
| C18:2 n-6 (LA) | 10.30 | 2.06 | 13.80 | 11.74 | 3.33 | 17.70 | 11.69 | 2.83 | 15.82 | 12.07 | 2.98 | 17.83 |
| C18:3 n-6 | 0.17 | 0.09 | 0.32 | 0.14 | 0.09 | 0.31 | 0.17 | 0.09 | 0.30 | 0.14 | 0.09 | 0.30 |
| C18:3 n-3 (ALA) | 0.20 | 0.13 | 0.47 | 0.26 | 0.13 | 0.47 | 0.30 | 0.09 | 0.40 | 0.24 | 0.10 | 0.43 |
| C20:0 | 0.28 | 0.11 | 0.43 | 0.27 | 0.11 | 0.43 | 0.28 | 0.10 | 0.43 | 0.25 | 0.11 | 0.42 |
| C20:2 n-6 | 0.42 | 0.14 | 0.66 | 0.38 | 0.20 | 0.74 | 0.39 | 0.15 | 0.58 | 0.38 | 0.19 | 0.74 |
| C20:3 n-6 | 1.15 | 0.44 | 1.89 | 1.21 | 0.34 | 1.75 | 1.15 | 0.42 | 1.75 | 1.25 | 0.34 | 1.78 |
| C20:4 n-6 (AA) | 9.80 | 3.99 | 15.79 | 11.13 | 4.14 | 16.04 | 9.66 | 4.17 | 15.25 | 11.36 | 4.10 | 15.79 |
| C20:5 n-3 (EPA) | 0.52 | 0.26 | 0.83 | 0.62 | 0.31 | 1.05 | 0.54 | 0.29 | 0.97 | 0.52 | 0.22 | 0.85 |
| C22:4 n-6 | 1.83 | 0.89 | 3.05 | 2.17 | 1.06 | 3.70 | 1.86 | 1.28 | 3.72 | 2.39 | 1.16 | 4.12 |
| C22:5 n-6 | 0.37 | 0.20 | 0.68 | 0.41 | 0.26 | 0.89 | 0.27 | 0.17 | 0.54 | 0.44 | 0.29 | 0.95 |
| C22:5 n-3 (DPA) | 1.38 | 0.74 | 2.66 | 1.55 | 0.66 | 2.64 | 1.38 | 0.47 | 1.91 | 1.55 | 0.65 | 2.58 |
| C22:6 n-3 (DHA) | 2.79 | 2.00 | 5.86 | 3.49 | 1.80 | 6.79 | 2.72 | 1.13 | 4.19 | 3.25 | 1.85 | 6.87 |
| SFA | 52.73 | 8.16 | 63.30 | 48.03 | 8.62 | 64.49 | 51.13 | 8.86 | 65.98 | 47.39 | 8.36 | 63.23 |
| MUFA | 18.06 | 2.40 | 21.98 | 18.62 | 2.89 | 23.27 | 18.53 | 4.07 | 24.65 | 18.81 | 3.12 | 23.12 |
| omega 6 | 24.04 | 6.02 | 30.64 | 27.19 | 6.53 | 36.23 | 25.20 | 6.03 | 31.59 | 28.03 | 6.43 | 36.75 |
| omega 3 | 4.90 | 2.82 | 9.45 | 5.92 | 2.48 | 10.51 | 4.93 | 1.67 | 7.10 | 5.55 | 2.49 | 10.39 |
| omega 6/omega 3 | 6.07 | 2.59 | 10.16 | 5.17 | 1.79 | 8.35 | 5.37 | 1.27 | 7.12 | 5.75 | 1.91 | 8.47 |
| LA/ALA | 67.65 | 35.22 | 117.67 | 56.48 | 30.23 | 99.20 | 42.50 | 16.93 | 70.39 | 61.29 | 29.98 | 99.42 |

Table 38: Composition in EPA, DPA, DHA and omega 3 of the red blood cells of the subjects of all areas regardless of the age and sex (\% total lipids, Mean $\pm$ SD)

|  | Le Havre $n=84$ | Lorient $n=115$ | La Rochelle n=97 | Toulon $n=95$ | All subjects $n=391$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| EPA | $0.83 \pm 0.45{ }^{\text {a }}$ | $0.74 \pm 0.35$ a | $0.81 \pm 0.34$ a | $0.59 \pm 0.30{ }^{\text {b }}$ | $0.74 \pm 0.37$ |
| DPA | $1.72 \pm 0.40$ a | $1.63 \pm 0.65$ a | $2.00 \pm 0.61{ }^{\text {b }}$ | $1.51 \pm 0.66$ a | $1.71 \pm 0.62$ |
| DHA | $4.16 \pm 1.17^{\text {a }}$ | $3.65 \pm 1.52^{\text {a }, ~} \mathrm{c}$ | $5.04 \pm 1.55{ }^{\text {b }}$ | $3.29 \pm 1.80{ }^{\text {c }}$ | $4.02 \pm 1.66$ |
| Omega 3* | $6.96 \pm 1.67{ }^{\text {a }}$ | $6.28 \pm 2.14^{\text {a,c }}$ | $8.02 \pm 2.27{ }^{\text {b }}$ | $5.65 \pm 2.50^{\text {c }}$ | $6.70 \pm 2.34$ |
| * The composition in Omega 3 corresponds to ALA, EPA, DPA and DHA <br> Values in the same raw with different superscript letters are significantly different, $\mathrm{p}<0.05$ (Tukey's test) |  |  |  |  |  |

Table 38 shows that heavy consumers in Toulon have significantly less EPA in their erythrocyte membrane than subjects in the other zones ( $p<0.05$ ). The people in La Rochelle have erythrocyte DPA, DHA and omega 3 levels significantly higher than those in the other zones, although the differences are relatively small.

### 4.2 Exposure to trace elements

### 4.2.1 Food exposure

As for fatty acids, it is important to note that the results presented in this section correspond to the exposure to the trace elements only through consumption of fish and seafood, not through the total diet. Nevertheless, as stated in the first part, food (and fish and seafood in particular) remain the main contributor of exposure to arsenic, organic tin and mercury, particularly methylmercury, its most toxic form. The intake results are presented in Tables 39 to 43 . The contributions of food to the exposure to the different contaminants and to the TRV are presented in the Appendix 5 and 6.

Some of the contamination data being censured (levels below the limit of detection), these have been taken to be equal to $1 / 2$ LOD, particularly $\mathrm{As}(\mathrm{V})$ and octyltins for which almost all the data were censured.

Arsenic: The average exposure to total arsenic (AsT) of $84.0 \pm 64.2 \mu \mathrm{~g} / \mathrm{kg}$ bw/week is very much higher than the average intake of French people estimated in 2004 to be $6.2 \mu \mathrm{~g} / \mathrm{kg}$ bw/week (see Introduction). The proportion of inorganic arsenic in the total arsenic of $0.8 \%$ is consistent with the figures of $0.4 \%$ to $5 \%$ usually found in the literature ${ }^{63}$. The average exposure to inorganic arsenic is between 0.40 and $0.72 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} /$ week for men aged 18 to 64 years and between 0.52 and $0.85 \mu \mathrm{~g} / \mathrm{kg}$ bw/week in women of the same age group and the P95 reaches $1.82 \mu \mathrm{~g} / \mathrm{kg}$ bw/week in these same women who are more
exposed than the other groups, although these values remain well below the PTWI of $15 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} / \mathrm{week}^{38}$ ( $<15 \%$ of the PTWI) ${ }^{119}$ established by the JECFA, an intake that none of our subjects exceeds. However, inorganic arsenic may be absorbed from sources other than seafood, in particular from drinking water. The Appendix 5 show that the products contributing most to the exposure of our population to toxic inorganic arsenic are great scallop ( $8.6 \%$ ), oyster ( $7.0 \%$ ), cod ( $6.3 \%$ ) and ray ( $5.1 \%$ ). Some differences are noticed between the different regions. While ray and cod appear to be major contributors to As and As inorg exposure in all four zones, the great scallop is a majority contributor to As ${ }_{\text {inorg }}$ exposure only in Le Havre ( $15 \%$ ) and in Toulon (14\%), and the sea urchin is a majority contributor only in Toulon (12\%). Nevertheless the fish and seafood consumption only provide $4.2 \%$ of the TRV (Appendix 6).

Mercury : The data reveal that mercury is almost exclusively absorbed in the form of methylmercury with an average exposure ranging from 0.88 to $1.50 \mu \mathrm{MeHg} / \mathrm{kg}$ bw/week for adult males and from 1.17 to $1.69 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{kg}$ bw/week for adult females. The exposure is of the same order of magnitude for elderly subjects ( 1.26 to $1.79 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{kg}$ bw/week) and women of child-bearing age ( 1.07 to 1.60 $\mu \mathrm{g} \mathrm{MeHg} / \mathrm{kg} \mathrm{bw} /$ week). We should underline that these average exposures are close to or even above the JECFA's PTWI of $1.6 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{kg}$ bw/week. More than a third of the subjects $(35 \%)$ have an intake exceeding the PTWI. Among these $29 \%$ are in Lorient, $28 \%$ in La Rochelle, $28 \%$ in Toulon and $14 \%$ in Le Havre. A third ( $32 \%$ ) are women of child-bearing age, considered to be the most sensitive population in view of the effects of high exposure on the foetus. The 95th percentile exposure of women of childbearing age is $3.09 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{kg}$ bw/week in La Rochelle and $4.26 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{kg}$ bw/week in Toulon, or 1.9 and 2.7 times the PTWI respectively, and 2 to 3 times the P95 level of such women in the INCA survey ${ }^{32}$.

The products contributing most to MeHg exposure, in all the subjects combined, are tuna (19\%), cod ( $7 \%$ ), ling and sole ( $6 \%$ each), with little difference from one zone to another. Fish generally accounts for $86 \%$ of the MeHg exposure, molluscs and crustaceans for $13 \%$ and other seafood for less than $2 \%$ (Appendix 5). This consumption contribute to $92 \%$ of the TRV with the same major contributors (Appendix 6).

Lead : The average exposures of high seafood consumers, even the highest percentiles, are well below the PTWI ( $25 \mu \mathrm{~g} / \mathrm{kg}$ bw/week). The average exposure ranges from 0.27 to $0.49 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} /$ week according to the group of individuals and the zone considered, or $1 \%$ to $2 \%$ of the PTWI. The highest P95 level is found in women of child-bearing age in La Rochelle ( $1.14 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} / \mathrm{week}$ ). The main contributor to lead exposure is sardine ( $17 \%$ ), but we also note large contributions from hake in Lorient ( $28 \%$ ), great scallop in Le Havre ( $22 \%$ ), mussels in La Rochelle ( $16 \%$ ) and sea urchin in Toulon ( $14 \%$ ). We should remember that there exist contributors to lead exposure other than fish and seafood.

Cadmium : The highest average exposure, in Le Havre subjects ( 3.50 to $5.00 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} / \mathrm{week}$ ), is less than the JECFA's PTWI of $7 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} / \mathrm{week}$. However, in the French population fish and seafood represent only $8 \%$ to $25 \%$ of the cadmium intake via food. The exposure of our heavy consumers should be interpreted with caution since it takes into account only their fish and seafood consumptions; their average total diet exposure could be higher. The PTWI is exceeded by $8.5 \%$ of our subjects through their fish and seafood consumption alone. The main contributors to this exposure are shrimp (16\%), crab ( $15 \%$ ), anchovy ( $10 \%$ ), periwinkle ( $7 \%$ ), great scallop ( $9 \%$ ) and sardine ( $5 \%$ ). However these contributions are not the same in all zones: the main contributors in Le Havre are shrimp ( $60 \%$ ), great scallop (14\%) and catshark ( $10 \%$ ); in Lorient they are crab ( $53 \%$ ) and saithe ( $15 \%$ ); in La Rochelle they are molluscs: periwinkle ( $21 \%$ ), calico scallop ( $19 \%$ ) and oyster ( $11 \%$ ); in Toulon they are anchovy ( $23 \%$ ) and great scallop (20\%). For the most exposed individuals in each zone (P90) we find the same main contributors (results not presented). These consumptions represent $35 \%$ of the TRV. Shrimps and crab are the major contributors (Appendix 6).

Organic tin : The average exposure to all the organostannic compounds does not exceed $0.14 \mu \mathrm{gn} / \mathrm{kg}$ $\mathrm{bw} /$ week, or $0.34 \mu \mathrm{gn} / \mathrm{kg}$ bw/week at P95, regardless of the zone, age group and sex, which represent respectively $19 \%$ and $47 \%$ of the PTWI fixed at $0.72 \mu \mathrm{~g} \mathrm{Sn} / \mathrm{kg}$ bw/week ${ }^{68}$ for TBT, DBT, TPT and DOT alone.

Generally speaking, regardless of the zone considered and whatever the trace element, women and elderly subjects ( $\geq 65$ years) are more exposed than men aged 18 to 64 years. More particularly, adult males are significantly less exposed to inorganic arsenic than women of the same age group ( $p<0.05$ ), all zones included.

Table 39: Food exposure of the high fish and seafood consumers to trace elements - Le Havre ( $\mu \mathrm{g} / \mathrm{kg}$ bw/week)

|  | Adult men$\begin{gathered} (18-64 y) \\ n=44 \end{gathered}$ |  |  | Adult women$\begin{array}{r} (18-64 y) \\ n=179 \end{array}$ |  |  | Older subjects (65 y and more)$n=26$ |  |  | Women of childbearing age$\begin{gathered} (18-44 y) \\ n=98 \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 |
| $\mathrm{As}_{\text {T }}$ | 59.49 | 43.31 | 144.65 | 89.79 | 70.56 | 242.83 | 87.97 | 56.03 | 200.18 | 78.93 | 63.13 | 212.64 |
| AsB | 57.02 | 42.09 | 135.96 | 85.20 | 66.08 | 230.09 | 83.58 | 53.01 | 185.94 | 75.45 | 59.37 | 192.00 |
| MMA | 0.22 | 0.19 | 0.55 | 0.29 | 0.36 | 0.77 | 0.29 | 0.29 | 0.94 | 0.31 | 0.45 | 1.04 |
| DMA | 0.34 | 0.26 | 0.81 | 0.44 | 0.36 | 1.14 | 0.44 | 0.31 | 0.94 | 0.44 | 0.39 | 1.29 |
| As(III) | 0.44 | 0.37 | 1.13 | 0.58 | 0.50 | 1.50 | 0.65 | 0.40 | 1.39 | 0.52 | 0.39 | 1.46 |
| As(V) | 0.06 | 0.03 | 0.13 | 0.08 | 0.05 | 0.19 | 0.07 | 0.04 | 0.16 | 0.08 | 0.05 | 0.20 |
| As org | 57.59 | 42.39 | 137.49 | 85.93 | 66.48 | 231.65 | 84.32 | 53.33 | 187.16 | 76.21 | 59.84 | 194.22 |
| As ${ }_{\text {inorg }}$ | 0.49 | 0.39 | 1.24 | 0.67 | 0.54 | 1.69 | 0.72 | 0.44 | 1.56 | 0.60 | 0.44 | 1.62 |
| $\mathrm{Hg}_{\text {T }}$ | 0.87 | 0.55 | 1.94 | 1.17 | 1.15 | 2.69 | 1.25 | 1.22 | 3.45 | 1.04 | 0.96 | 2.28 |
| MeHg | 0.88 | 0.57 | 1.93 | 1.17 | 1.17 | 2.69 | 1.26 | 1.31 | 3.45 | 1.07 | 1.02 | 2.27 |
| Cd | 3.50 | 2.32 | 7.23 | 5.00 | 5.04 | 12.70 | 4.15 | 4.33 | 9.69 | 4.64 | 4.17 | 11.82 |
| Pb | 0.27 | 0.21 | 0.65 | 0.35 | 0.31 | 0.79 | 0.37 | 0.26 | 0.66 | 0.31 | 0.26 | 0.78 |
| $\mathrm{OTC}_{T}$ | 0.11 | 0.07 | 0.21 | 0.14 | 0.10 | 0.34 | 0.13 | 0.09 | 0.34 | 0.13 | 0.09 | 0.34 |
| Butyl | 0.09 | 0.06 | 0.18 | 0.11 | 0.08 | 0.29 | 0.10 | 0.08 | 0.28 | 0.10 | 0.08 | 0.27 |
| Phényl | 0.02 | 0.01 | 0.03 | 0.02 | 0.02 | 0.06 | 0.02 | 0.02 | 0.05 | 0.02 | 0.02 | 0.04 |
| Octyl | 0.00 | 0.00 | 0.01 | 0.01 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.01 | 0.00 | 0.01 |

As : total arsenic, $\mathrm{As}_{\text {org: }}$ : organic arsenic, As inorg : inorganic arsenic, $\mathrm{Hg}_{\mathrm{T}}$ : total mercury, $\mathrm{OTC}_{\mathrm{T}}$ : All organostannic compounds, in $\mu \mathrm{g} \mathrm{Sn} / \mathrm{kg}$ bw/wk, Butyl: butyltin, Phenyl: phenyltin, Octyl: octyltin.

Table 40: Food exposure of the high fish and seafood consumers to trace elements - Lorient ( $\mu \mathrm{g} / \mathrm{kg} \mathrm{bw} /$ week)

|  | Adult men$\begin{gathered} (18-64 y) \\ n=52 \end{gathered}$ |  |  | $\begin{gathered} \text { Adult women } \\ \begin{array}{c} (18-64 y) \\ n=158 \end{array} \end{gathered}$ |  |  | $\begin{gathered} \text { Older subjects } \\ \text { (65 y and more) } \\ n=37 \end{gathered}$ |  |  | Women of childbearing age$\begin{gathered} (18-44 y) \\ n=76 \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 |
| $\mathrm{As}_{\text {T }}$ | 91.58 | 53.54 | 199.79 | 104.47 | 74.64 | 232.80 | 116.84 | 67.28 | 244.47 | 97.53 | 81.43 | 230.26 |
| AsB | 82.96 | 37.49 | 181.01 | 95.64 | 69.73 | 219.04 | 108.93 | 64.04 | 228.01 | 89.66 | 76.51 | 220.44 |
| MMA | 0.31 | 1.14 | 0.68 | 0.32 | 0.38 | 0.74 | 0.32 | 0.31 | 0.84 | 0.27 | 0.42 | 0.53 |
| DMA | 0.81 | 0.37 | 2.07 | 0.76 | 0.72 | 2.01 | 0.61 | 0.42 | 1.40 | 0.68 | 0.76 | 1.92 |
| As(III) | 0.64 | 0.27 | 1.38 | 0.75 | 0.57 | 1.63 | 0.72 | 0.40 | 1.49 | 0.68 | 0.62 | 1.54 |
| As(V) | 0.09 | 0.01 | 0.17 | 0.09 | 0.05 | 0.18 | 0.09 | 0.04 | 0.17 | 0.09 | 0.06 | 0.18 |
| As org | 84.08 | 39.00 | 183.50 | 96.72 | 70.55 | 220.57 | 109.85 | 64.53 | 229.72 | 90.61 | 77.47 | 222.68 |
| As ${ }_{\text {inorg }}$ | 0.72 | 0.29 | 1.54 | 0.85 | 0.62 | 1.82 | 0.81 | 0.43 | 1.67 | 0.77 | 0.67 | 1.71 |
| $\mathrm{Hg}_{\text {T }}$ | 1.40 | 0.21 | 3.11 | 1.63 | 1.13 | 3.75 | 1.74 | 0.89 | 3.32 | 1.50 | 1.15 | 2.79 |
| MeHg | 1.44 | 0.34 | 3.10 | 1.67 | 1.15 | 3.67 | 1.75 | 0.89 | 3.30 | 1.54 | 1.16 | 2.80 |
| Cd | 3.10 | 1.36 | 8.79 | 2.92 | 3.14 | 10.04 | 2.24 | 2.49 | 6.44 | 2.62 | 2.70 | 6.63 |
| Pb | 0.43 | 0.29 | 0.96 | 0.44 | 0.31 | 0.94 | 0.43 | 0.23 | 0.81 | 0.40 | 0.30 | 0.93 |
| OTC $_{\text {T }}$ | 0.07 | 0.03 | 0.12 | 0.07 | 0.04 | 0.16 | 0.07 | 0.03 | 0.12 | 0.07 | 0.05 | 0.15 |
| Butyl | 0.05 | 0.02 | 0.09 | 0.05 | 0.03 | 0.11 | 0.04 | 0.02 | 0.07 | 0.05 | 0.03 | 0.10 |
| Phényl | 0.01 | 0.01 | 0.03 | 0.02 | 0.01 | 0.04 | 0.02 | 0.01 | 0.03 | 0.02 | 0.01 | 0.04 |
| Octyl | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.01 |
| $\mathrm{As}_{\mathrm{T}}$ : total arsenic, $\mathrm{As}_{\text {org }}$ : organic arsenic, $\mathrm{As}_{\text {inorg }}$ : inorganic arsenic, $\mathrm{Hg}_{\mathrm{T}}$ : total mercury, OTC $\mathrm{C}_{\mathrm{T}}$ : All organostannic compounds, in $\mu \mathrm{g} \mathrm{Sn} / \mathrm{kg}$ bw/wk, Butyl: butyltin, Phenyl: phenyltin, Octyl: octyltin. |  |  |  |  |  |  |  |  |  |  |  |  |

Table 41: Food exposure of the high fish and seafood consumers to trace elements La Rochelle ( $\mu \mathrm{g} / \mathrm{kg} \mathrm{bw} / \mathrm{week}$ )

|  | Adult men$\begin{gathered} (18-64 y) \\ n=87 \end{gathered}$ |  |  | $\begin{aligned} & \text { Adult women } \\ & \begin{array}{c} (18-64 y) \\ n=122 \end{array} \end{aligned}$ |  |  | Older subjects (65 y and more) $\mathrm{n}=39$ |  |  | Women of childbearing age$\begin{gathered} (18-44 y) \\ n=78 \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 |
| $\mathrm{As}_{\text {T }}$ | 64.78 | 53.50 | 153.88 | 77.38 | 50.82 | 175.58 | 86.64 | 55.77 | 189.20 | 69.27 | 43.57 | 156.61 |
| AsB | 59.06 | 48.36 | 139.18 | 70.96 | 45.42 | 155.38 | 77.42 | 48.64 | 169.13 | 63.49 | 39.34 | 142.36 |
| MMA | 0.13 | 0.11 | 0.36 | 0.15 | 0.12 | 0.37 | 0.16 | 0.11 | 0.34 | 0.15 | 0.12 | 0.37 |
| DMA | 0.58 | 0.48 | 1.22 | 0.75 | 0.54 | 1.74 | 0.83 | 0.51 | 1.58 | 0.73 | 0.59 | 2.02 |
| As(III) | 0.32 | 0.22 | 0.69 | 0.42 | 0.28 | 0.93 | 0.39 | 0.28 | 0.93 | 0.41 | 0.30 | 0.91 |
| As(V) | 0.08 | 0.05 | 0.15 | 0.09 | 0.05 | 0.21 | 0.09 | 0.04 | 0.17 | 0.09 | 0.05 | 0.20 |
| As org | 59.77 | 48.73 | 140.48 | 71.86 | 45.75 | 157.58 | 78.41 | 48.91 | 170.04 | 64.38 | 39.68 | 143.68 |
| As sinorg | 0.40 | 0.26 | 0.79 | 0.52 | 0.33 | 1.09 | 0.48 | 0.32 | 1.07 | 0.50 | 0.34 | 1.09 |
| $\mathrm{Hg}_{\text {T }}$ | 1.39 | 1.29 | 3.01 | 1.59 | 1.15 | 3.52 | 1.75 | 1.06 | 3.58 | 1.39 | 0.92 | 3.03 |
| MeHg | 1.42 | 1.27 | 3.08 | 1.65 | 1.19 | 3.62 | 1.79 | 1.09 | 3.81 | 1.43 | 0.96 | 3.09 |
| Cd | 1.22 | 2.16 | 3.26 | 1.72 | 2.46 | 6.06 | 1.55 | 1.78 | 5.19 | 1.92 | 2.92 | 7.85 |
| Pb | 0.38 | 0.30 | 0.76 | 0.48 | 0.31 | 1.13 | 0.49 | 0.32 | 1.06 | 0.47 | 0.34 | 1.14 |
| OTC $_{T}$ | 0.07 | 0.05 | 0.14 | 0.09 | 0.05 | 0.20 | 0.08 | 0.04 | 0.16 | 0.08 | 0.05 | 0.18 |
| Butyl | 0.05 | 0.04 | 0.11 | 0.07 | 0.04 | 0.14 | 0.06 | 0.03 | 0.11 | 0.06 | 0.04 | 0.14 |
| Phényl | 0.01 | 0.01 | 0.03 | 0.02 | 0.01 | 0.04 | 0.02 | 0.01 | 0.03 | 0.01 | 0.01 | 0.03 |
| Octyl | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.02 | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.01 |
| $\mathrm{As}_{\mathrm{T}}$ : total arsenic, $\mathrm{As}_{\text {org }}$ : organic arsenic, $\mathrm{As}_{\text {inorg }}$ : inorganic arsenic, $\mathrm{Hg}_{\mathrm{T}}$ : total mercury, $\mathrm{OTC}_{\mathrm{T}}$ : All organostannic compounds, in $\mu \mathrm{g} \mathrm{Sn} / \mathrm{kg}$ bw/wk, Butyl: butyltin, Phenyl: phenyltin, Octyl: octyltin. |  |  |  |  |  |  |  |  |  |  |  |  |

Table 42: Food exposure of the high fish and seafood consumers to trace elements - Toulon ( $\mu \mathrm{g} / \mathrm{kg}$ bw/week)

|  | Adult men$\begin{gathered} (18-64 y) \\ n=60 \end{gathered}$ |  |  | $\begin{aligned} & \text { Adult women } \\ & \begin{array}{c} (18-64 y) \\ n=171 \end{array} \end{aligned}$ |  |  | $\begin{aligned} & \text { Older subjects } \\ & \text { (65 y and more) } \\ & n=21 \end{aligned}$ |  |  | Women of childbearing age$\begin{gathered} (18-44 y) \\ n=92 \end{gathered}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 | Mean | SD | P95 |
| $\mathrm{As}_{\text {T }}$ | 64.53 | 57.33 | 143.34 | 76.32 | 65.02 | 222.04 | 80.42 | 58.94 | 146.55 | 72.00 | 63.00 | 232.77 |
| AsB | 54.63 | 47.54 | 120.97 | 66.63 | 57.28 | 193.31 | 69.78 | 52.42 | 131.78 | 63.06 | 55.86 | 209.70 |
| MMA | 0.10 | 0.09 | 0.25 | 0.12 | 0.10 | 0.26 | 0.10 | 0.06 | 0.20 | 0.12 | 0.11 | 0.30 |
| DMA | 0.49 | 0.40 | 1.24 | 0.52 | 0.37 | 1.18 | 0.60 | 0.38 | 1.04 | 0.52 | 0.39 | 1.21 |
| As(III) | 0.50 | 0.49 | 1.58 | 0.58 | 0.87 | 1.61 | 0.49 | 0.45 | 1.53 | 0.46 | 0.41 | 1.45 |
| As(V) | 0.08 | 0.05 | 0.15 | 0.09 | 0.06 | 0.20 | 0.08 | 0.03 | 0.11 | 0.09 | 0.05 | 0.18 |
| As org | 55.23 | 47.96 | 122.08 | 67.27 | 57.61 | 194.36 | 70.48 | 52.69 | 132.74 | 63.70 | 56.16 | 211.56 |
| As ${ }_{\text {inorg }}$ | 0.58 | 0.54 | 1.81 | 0.67 | 0.91 | 1.78 | 0.56 | 0.47 | 1.63 | 0.55 | 0.45 | 1.67 |
| $\mathrm{Hg}_{\text {T }}$ | 1.54 | 1.31 | 4.73 | 1.71 | 1.44 | 4.11 | 1.54 | 1.13 | 3.05 | 1.61 | 1.27 | 3.87 |
| MeHg | 1.50 | 1.29 | 4.09 | 1.69 | 1.42 | 4.43 | 1.50 | 0.80 | 2.87 | 1.60 | 1.29 | 4.26 |
| Cd | 0.83 | 0.75 | 2.16 | 0.76 | 0.76 | 1.92 | 0.67 | 0.66 | 2.28 | 0.71 | 0.79 | 1.73 |
| Pb | 0.38 | 0.30 | 0.96 | 0.42 | 0.59 | 1.19 | 0.39 | 0.27 | 0.77 | 0.36 | 0.40 | 1.03 |
| OTC $_{T}$ | 0.10 | 0.05 | 0.18 | 0.11 | 0.08 | 0.27 | 0.09 | 0.04 | 0.14 | 0.11 | 0.08 | 0.27 |
| Butyl | 0.08 | 0.04 | 0.16 | 0.09 | 0.06 | 0.21 | 0.07 | 0.03 | 0.11 | 0.09 | 0.07 | 0.22 |
| Phényl | 0.01 | 0.01 | 0.03 | 0.02 | 0.01 | 0.04 | 0.01 | 0.01 | 0.02 | 0.02 | 0.01 | 0.03 |
| Octyl | 0.01 | 0.01 | 0.02 | 0.01 | 0.01 | 0.02 | 0.01 | 0.00 | 0.01 | 0.01 | 0.01 | 0.02 |
| $\mathrm{As}_{\mathrm{T}}$ : total arsenic, $\mathrm{As}_{\text {org }}$ : organic arsenic, $\mathrm{As}_{\text {inorg }}$ : inorganic arsenic, $\mathrm{Hg}_{\mathrm{T}}$ : total mercury, $\mathrm{OTC}_{\mathrm{T}}$ : All organostannic compounds, in $\mu \mathrm{g} \mathrm{Sn} / \mathrm{kg}$ bw/wk, Butyl: butyltin, Phenyl: phenyltin, Octyl: octyltin. |  |  |  |  |  |  |  |  |  |  |  |  |

Table 43: Food exposure of the high fish and seafood consumers to trace elements of all areas regardless of the age and sex ( $\mu \mathrm{g} / \mathrm{kg}$ bw/week, Mean $\pm$ SD)

|  | Le Havre $n=249$ | Lorient $n=247$ | La Rochelle $n=248$ | Toulon $n=252$ | All subjects n=996 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{As}_{\text {T }}$ | $84.2 \pm 65.9^{\text {a }}$ | $104 \pm 69.7^{\text {b }}$ | $74.4 \pm 52.9^{\text {a }}$ | $73.9 \pm 62.8^{\text {a }}$ | $84.0 \pm 64.2$ |
| As ${ }_{\text {inorg }}$ | $0.64 \pm 0.51^{\text {a }}$ | $0.81 \pm 0.56{ }^{\text {b }}$ | $0.47 \pm 0.31^{\text {c }}$ | $0.64 \pm 0.81^{\text {a }}$ | $0.64 \pm 0.59$ |
| Pb | $0.34 \pm 0.29{ }^{\text {a }}$ | $0.44 \pm 0.29^{\text {b }}$ | $0.45 \pm 0.31^{\text {b }}$ | $0.40 \pm 0.51^{\text {a.b }}$ | $0.41 \pm 0.37$ |
| Cd | $4.64 \pm 4.63^{\text {a }}$ | $2.86 \pm 3.01^{\text {b }}$ | $1.52 \pm 2.27^{\text {c }}$ | $0.77 \pm 0.74^{\text {d }}$ | $2.44 \pm 3.34$ |
| $\mathrm{Hg}_{\text {T }}$ | $1.12 \pm 1.08^{\text {a }}$ | $1.60 \pm 1.04^{\text {b }}$ | $1.55 \pm 1.19^{\text {b }}$ | $1.66 \pm 1.38^{\text {b }}$ | $1.48 \pm 1.20$ |
| $\mathbf{M e H g}$ | $1.13 \pm 1.11^{\text {a }}$ | $1.63 \pm 1.05^{\text {b }}$ | $1.59 \pm 1.21^{\text {b }}$ | $1.63 \pm 1.35^{\text {b }}$ | $1.49 \pm 1.20$ |
| OTC $_{\text {T }}$ | $0.13 \pm 0.09^{\text {a }}$ | $0.07 \pm 0.04{ }^{\text {b }}$ | $0.08 \pm 0.05^{\text {b }}$ | $0.11 \pm 0.07^{\text {c }}$ | $0.10 \pm 0.07$ |
| As $\mathrm{T}_{\mathrm{T}}$ : total arsenic, $\mathrm{As}_{\text {inorg }}$ : inorganic arsenic, $\mathrm{Hg}_{\mathrm{T}}$ : total mercury, $\mathrm{OTC}_{\mathrm{T}}$ : All organostannic compounds, in $\mu \mathrm{g} \mathrm{Sn} / \mathrm{kg} \mathrm{bw} / \mathrm{wk}$ Values in the same raw with different superscript letters are significantly different, $\mathrm{p}<0.05$ (Tukey's test) |  |  |  |  |  |

Table 43 indicates that the total arsenic exposure in Lorient is significantly higher than in the other zones ( $\mathrm{p}<0.05$ ). For inorganic arsenic, the most toxic form ( $\mathrm{As} \mathrm{s}_{\text {inorg }}$ ), the subjects in Lorient are also significantly more exposed and the subjects in La Rochelle are significantly less exposed than those in the other zones.

Regarding cadmium, we observe a significant north-south exposure gradient ( $\mathrm{p}<0.05$ ) with a maximum in Le Havre and a minimum in Toulon. On the other hand, for mercury ( HgT ) and more particularly methylmercury the exposure of subjects is significantly less in Le Havre than in the other zones.

Finally, for organostannic compounds (OTC ${ }_{T}$ ) we observe that subjects in Le Havre and Toulon are significantly more exposed than people in the other zones ( $\mathrm{p}<0.05$ ).

### 4.2.2 Biomarkers of exposure

Table 44 presents the trace element concentrations found in the blood of our tested subjects.
Lead : Twenty-two subjects ( $6 \%$ of the subjects for which a blood sample was taken) display a blood level exceeding the so-called "standard" ( $90 \mu \mathrm{~g} / \mathrm{L}$ for men, $70 \mu \mathrm{~g} / \mathrm{L}$ for women) ${ }^{120}$, but none exceed the concentration of $200 \mu \mathrm{~g} / \mathrm{L}$ above which medical monitoring is required. The average lead levels in the blood range from 27.1 to $52.3 \mu \mathrm{~g} / \mathrm{L}$ in individuals aged 18 to 64 years; they are slightly higher in elderly subjects at 40.7 to $77.2 \mu \mathrm{~g} / \mathrm{L}$.

Mercury : For total mercury, 13 subjects (3\%) exceed the "standard" of $10 \mu \mathrm{~g} / \mathrm{L}$ of blood ${ }^{120}$. In all the zones women of child-bearing age (under 45 years old) are the group with the lowest Hg and MeHg levels in the blood ( 1.91 to $4.13 \mu \mathrm{~g} / \mathrm{L}$ and 2.29 to $3.39 \mu \mathrm{~g} / \mathrm{L}$ respectively, depending on the zone). Elderly subjects constitute the group with the highest levels, except in Toulon ( 3.12 to $5.91 \mu \mathrm{~g} \mathrm{Hg} / \mathrm{L}$ and 3.85 to $5.34 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{L}$ ). The P95 level is $9.07 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{L}$ in Toulon and $9.69 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{L}$ in Le Havre.

Cadmium : The urinary cadmium level is the biomarker usually used (see Introduction). Nevertheless depending on the zone, the average cadmium levels in the blood range from 0.42 to $0.94 \mu \mathrm{~g} / \mathrm{L}$ for men or from 0.54 to $0.64 \mu \mathrm{~g} / \mathrm{L}$ for women aged 18 to 64 years. A concentration exceeding the "standard" ( $1 \mu \mathrm{~g} / \mathrm{L}$ for non-smokers and $2 \mu \mathrm{~g} / \mathrm{L}$ for smokers) ${ }^{120}$ was found in 18 individuals ( $4.6 \%$ ). No individuals exceeded the concentration associated with toxicity ( $20 \mu \mathrm{~g} / \mathrm{L}$ ).

Table 44: Blood concentrations in trace elements ( $\mu \mathrm{g} / \mathrm{L}$ )

|  |  |  | Cd | Pb | $\mathrm{Hg}_{\text {T }}$ | MeHg |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Le Havre | Adult men (18-64 y) $\mathrm{n}=18$ | Mean | 0.42 | 36.4 | 2.6 | 2.95 |
|  |  | SD | 0.34 | 19.6 | 2.54 | 2.40 |
|  | Adult women (18-64 y) $\mathrm{n}=59$ | Mean | 0.54 | 28.9 | 2.75 | 3.44 |
|  |  | SD | 0.44 | 14.9 | 2.05 | 2.53 |
|  | Older subjects (65 y and more) $\mathrm{n}=6$ | Mean | 0.49 | 40.7 | 3.12 | 3.85 |
|  |  | SD | 0.19 | 14.9 | 2.31 | 2.87 |
|  | Women of childbearing age (18-44 y) $\mathrm{n}=28$ | Mean | 0.62 | 20.6 | 1.91 | 2.29 |
|  |  | SD | 0.54 | 7.02 | 1.41 | 1.27 |
|  | All subjects $\mathrm{n}=83$ | Mean | 0.51 | 31.4 | 2.76 | 3.37 |
|  |  | SD | 0.41 | 16.3 | 2.16 | 2.51 |
|  |  | P95 | 1.43 | 57.6 | 7.49 | 9.69 |
|  | Adult men (18-64 y) $\mathrm{n}=21$ | Mean | 0.94 | 40.6 | 3.27 | 3.32 |
|  |  | SD | 1.00 | 20.1 | 1.64 | 1.82 |
|  | Adult women (18-64 y) $\mathrm{n}=84$ | Mean | 0.57 | 27.1 | 3.61 | 3.57 |
|  |  | SD | 0.44 | 13.1 | 2.35 | 2.27 |
|  | Older subjects (65 y and more) $\mathrm{n}=10$ | Mean | 0.55 | 54.1 | 4.83 | 5.34 |
|  |  | SD | 0.26 | 39.5 | 3.15 | 3.51 |
|  | Women of childbearing age (18-44 y) $\mathrm{n}=39$ | Mean | 0.65 | 21.1 | 2.54 | 2.55 |
|  |  | SD | 0.53 | 11.7 | 1.65 | 1.49 |
|  | All subjects $\mathrm{n}=115$ | Mean | 0.63 | 31.9 | 3.65 | 3.68 |
|  |  | SD | 0.59 | 19.8 | 2.33 | 2.36 |
|  |  | P95 | 1.72 | 69.6 | 8.40 | 8.34 |
|  | Adult men (18-64 y) $\mathrm{n}=38$ | Mean | 0.76 | 52.3 | 4.07 | 3.55 |
|  |  | SD | 0.90 | 22.7 | 2.55 | 2.18 |
|  | Adult women (18-64 y) $\mathrm{n}=46$ | Mean | 0.64 | 39.3 | 4.61 | 3.74 |
|  |  |  | 0.48 | 22.3 | 3.38 | 3.02 |
|  | Older subjects (65 y and more) $\mathrm{n}=14$ | Mean | 0.67 | 77.2 | 5.91 | 4.82 |
|  |  | SD | 0.33 | 40.6 | 4.23 | 3.35 |
|  | Women of childbearing age (18-44 y) $\mathrm{n}=28$ | Mean | 0.67 | 32.0 | 4.13 | 3.39 |
|  |  | SD | 0.60 | 18.6 | 3.11 | 2.94 |
|  | All subjects $\mathrm{n}=98$ | Mean | 0.69 | 49.7 | 4.59 | 3.82 |
|  |  | SD | 0.66 | 28.5 | 3.24 | 2.78 |
|  |  | P95 | 1.94 | 103.3 | 11.9 | 8.79 |
| $\begin{aligned} & \text { 등 } \\ & \frac{0}{2} \end{aligned}$ | Adult men (18-64 y) $\mathrm{n}=17$ | Mean | 0.73 | 48.0 | 2.96 | 3.57 |
|  |  |  | 0.50 | 28.4 | 2.29 | 2.80 |
|  | Adult women (18-64 y) $\mathrm{n}=69$ | Mean | 0.58 | 33.3 | 3.14 | 4.02 |
|  |  |  | 0.44 | 21.4 | 5.34 | 7.07 |
|  | Older subjects (65 y and more) $\mathrm{n}=9$ | Mean | 1.05 | 57.3 | 3.37 | 4.72 |
|  |  |  | 1.49 | 21.4 | 2.03 | 2.92 |
|  | Women of childbearing age (18-44 y) $\mathrm{n}=41$ | Mean | 0.66 | 26.7 | 2.35 | 2.84 |
|  |  | SD | 0.52 | 18.8 | 2.26 | 2.23 |
|  | All subjects $\mathrm{n}=95$ | Mean | 0.65 | 38.2 | 3.13 | 4.01 |
|  |  | SD | 0.63 | 24.0 | 4.68 | 6.19 |
|  |  | P95 | 1.53 | 84.0 | 7.01 | 9.07 |
|  | All subjects $\mathrm{n}=391$ | Mean | 0.62 | 37.8 | 3.57 | 3.73 |
|  |  | SD | 0.58 | 23.8 | 3.30 | 3.76 |
|  |  | P95 | 1.70 | 84.3 | 8.90 | 8.84 |
| $\mathrm{Hg}_{\mathrm{T}}$ : total mercury |  |  |  |  |  |  |

Tables 45 and 46 present the trace element concentrations found in the urine of our tested subjects.
12 subjects ( $3 \%$ ) have a total arsenic level less than the limit of quantification ( $5 \mu \mathrm{~g} / \mathrm{L}$ ); 242 subjects ( $63 \%$ ) have urinary cadmium levels less than the LOQ ( $0.5 \mu \mathrm{~g} / \mathrm{L}$ ); 335 subjects ( $87 \%$ ) have lead levels less than the LOQ ( $5 \mu \mathrm{~g} / \mathrm{L}$ ). However, in order not to underestimate the biological level, the averages presented in Table 45 include all the subjects, with levels below the LOQ taken to be equal to $1 / 2$ LOQ for each of the trace elements.

Cadmium : The measurements indicate that the cadmium levels in urine are in the "standard" range, less than $2 \mu \mathrm{~g} / \mathrm{g}$ creatinine, even for the high percentiles (P95). Only 12 people (3\%) exceed this norm, yet these are not the people having the highest cadmium concentrations in the blood.

Lead : The lead levels are also low for all age groups and both sexes, on average $5.7 \pm 4.4 \mu \mathrm{~g} / \mathrm{g}$ creatinine. Levels of $25 \mu \mathrm{~g} / \mathrm{g}$ creatinine (the "standard" ${ }^{120}$ ) or more were observed in four individuals ( $1 \%$ ), who were also among the people having the highest lead concentrations in the blood.

Arsenic : Table 46 presents the results of the arsenic speciation performed on the 101 subjects displaying the highest level of total arsenic in the urine ( $>75 \mu \mathrm{~g} / \mathrm{g}$ creatinine). 87 subjects ( $86 \%$ ) have levels of inorganic arsenic (As(III), As $(\mathrm{V})$ and its derivatives MMA(V) and DMA(V)) exceeding the "standard" of $10 \mu \mathrm{~g} / \mathrm{g}$ creatinine ${ }^{120}$. These forms of arsenic account for $16.1 \%$ of total arsenic in urine.

Table 45: Urinary concentrations in trace elements

|  |  |  | As <br> ( $\mu \mathrm{g} / \mathrm{g}$ creatinine) | Cd <br> ( $\mu \mathrm{g} / \mathrm{g}$ creatinine) | $\begin{gathered} \mathrm{Pb} \\ (\mu \mathrm{~g} / \mathrm{g} \text { creatinine) }) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Adult men (18-64 y) $\mathrm{n}=18$ | Mean $\pm$ SD | $39.0 \pm 50.4$ | $0.5 \pm 0.4$ | $4.2 \pm 3.6$ |
|  | Adult women ( $18-64 \mathrm{y}$ ) $\mathrm{n}=60$ | Mean $\pm$ SD | $58.4 \pm 87.4$ | $0.6 \pm 0.3$ | $4.8 \pm 2.5$ |
|  | Older subjects ( 65 y and more) $\mathrm{n}=6$ | Mean $\pm$ SD | $50.8 \pm 48.3$ | $1.0 \pm 0.7$ | $5.7 \pm 1.5$ |
|  | Women of childbearing age (18-44 y) $\mathrm{n}=29$ | Mean $\pm$ SD | $41.7 \pm 58.1$ | $0.5 \pm 0.2$ | $4.2 \pm 2.3$ |
|  | All subjects $\mathrm{n}=84$ | Mean $\pm$ SD | $53.7 \pm 78.4$ | $0.6 \pm 0.4$ | $4.8 \pm 2.7$ |
|  |  | P95 | 175 | 1.2 | 9.8 |
| + | Adult men ( $18-64 \mathrm{y}$ ) $\mathrm{n}=21$ | Mean $\pm$ SD | $55.9 \pm 101$ | $0.4 \pm 0.2$ | $3.7 \pm 1.9$ |
|  | Adult women ( $18-64 \mathrm{y}$ ) $\mathrm{n}=84$ | Mean $\pm$ SD | $84.7 \pm 152$ | $0.7 \pm 0.4$ | $6.5 \pm 4.3$ |
|  | Older subjects ( 65 y and more) $\mathrm{n}=10$ | Mean $\pm$ SD | $57.4 \pm 39.0$ | $1.0 \pm 0.5$ | $7.3 \pm 2.6$ |
|  | Women of childbearing age (18-44 y) $\mathrm{n}=39$ | Mean $\pm$ SD | $36.7 \pm 36.4$ | $0.5 \pm 0.5$ | $6.0 \pm 5.4$ |
|  | All subjects $\mathrm{n}=115$ | Mean $\pm$ SD | $77.1 \pm 137$ | $0.7 \pm 0.4$ | $6.0 \pm 4.0$ |
|  |  | P95 | 269 | 1.4 | 12.5 |
|  | Adult men ( $18-64 \mathrm{y}$ ) $\mathrm{n}=38$ | Mean $\pm$ SD | $69.9 \pm 92.0$ | $0.5 \pm 0.3$ | $3.7 \pm 1.7$ |
|  | Adult women ( $18-64 \mathrm{y}$ ) $\mathrm{n}=46$ | Mean $\pm$ SD | $160 \pm 449$ | $0.7 \pm 0.6$ | $6.5 \pm 6.2$ |
|  | Older subjects ( 65 y and more) $\mathrm{n}=13$ | Mean $\pm$ SD | $94.9 \pm 119$ | $0.6 \pm 0.3$ | $7.2 \pm 3.7$ |
|  | Women of childbearing age (18-44 y) $\mathrm{n}=28$ | Mean $\pm$ SD | $70.9 \pm 79.2$ | $0.6 \pm 0.6$ | $5.6 \pm 6.1$ |
|  | All subjects $\mathrm{n}=97$ | Mean $\pm$ SD | $116 \pm 319$ | $0.6 \pm 0.5$ | $5.5 \pm 4.8$ |
|  |  | P95 | 333 | 1.5 | 15.0 |
| $\begin{aligned} & \text { 들 } \\ & \text { 응 } \end{aligned}$ | Adult men ( $18-64 \mathrm{y}$ ) $\mathrm{n}=16$ | Mean $\pm$ SD | $52.2 \pm 68.7$ | $0.6 \pm 0.4$ | $4.0 \pm 2.4$ |
|  | Adult women (18-64 y) $\mathrm{n}=66$ | Mean $\pm$ SD | $100 \pm 292$ | $0.7 \pm 0.5$ | $6.0 \pm 4.8$ |
|  | Older subjects (65 y and more) $\mathrm{n}=9$ | Mean $\pm$ SD | $79.9 \pm 75.4$ | $1.3 \pm 0.4$ | $12.0 \pm 8.9$ |
|  | Women of childbearing age (18-44 y) $\mathrm{n}=39$ | Mean $\pm$ SD | $114 \pm 70.6$ | $0.5 \pm 0.3$ | $4.7 \pm 2.8$ |
|  | All subjects $\mathrm{n}=91$ | Mean $\pm$ SD | $89.7 \pm 251$ | $0.7 \pm 0.5$ | $6.2 \pm 5.4$ |
|  |  | P95 | 258 | 1.7 | 16.4 |
| All subjects $\mathrm{n}=387$ |  | Mean $\pm$ SD | $84.8 \pm 218$ | $0.7 \pm 0.5$ | $5.7 \pm 4.4$ |
|  |  | P95 | 288 | 1.5 | 12.7 |
| $\mathrm{AS}_{\mathrm{T}}$ : total arsenic |  |  |  |  |  |

Table 46: Arsenic speciation in the urine of the 101 subjects with the highest urinary total arsenic level (Mean $\pm$ SD)

|  | Le Havre $\mathrm{n}=15$ | Lorient n=33 | La Rochelle $\mathrm{n}=32$ | Toulon $\mathrm{n}=21$ | All subjects $\mathrm{n}=101$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{As}_{\mathrm{T}}$ ( $\mu \mathrm{g} / \mathrm{L}$ urine) | $193 \pm 127$ | $180 \pm 167$ | $315 \pm 377$ | $344 \pm 496$ | $259 \pm 331$ |
| Creatinine ( $\mathrm{g} / \mathrm{L}$ urine) | $1.20 \pm 0.60$ | $1.00 \pm 0.40$ | $1.40 \pm 0.70$ | $1.20 \pm 0.60$ | $1.20 \pm 0.60$ |
| $\mathrm{As}_{\mathrm{T}} /$ Creat ( $\mu \mathrm{g} / \mathrm{g}$ Creat) | $180 \pm 118$ | $201 \pm 210$ | $283 \pm 520$ | $304 \pm 469$ | $245 \pm 383$ |
| As $\mathrm{inorg}^{\text {( }} \mathrm{\mu} / \mathrm{L} / \mathrm{L}$ urine) | $24.6 \pm 21.5$ | $19.9 \pm 11.2$ | $31.8 \pm 26.7$ | $38.2 \pm 41.6$ | $28.2 \pm 26.8$ |
| As ${ }_{\text {inorg }} /$ Creat ( $\mu \mathrm{g} / \mathrm{g}$ ) | $25.7 \pm 24.8$ | $22.8 \pm 12.3$ | $26.0 \pm 24.1$ | $33.7 \pm 28.9$ | $26.5 \pm 22.3$ |
| \% As ${ }_{\text {inorg }}$ | $16.5 \pm 12.4$ | $16.3 \pm 10.4$ | $14.4 \pm 10.8$ | $18.3 \pm 11.9$ | $16.1 \pm 11.1$ |
| $\mathrm{As}_{\mathrm{T}}$ : total arsenic. $\mathrm{As}_{\text {inorg }}$ : inorganic arsenic ( $\mathrm{As}(\mathrm{III}), \mathrm{As}(\mathrm{V})$ and their metabolits MMA(V) and DMA(V)), Creat: creatinine |  |  |  |  |  |

Whatever the trace element, age group and sex considered, the subjects in Le Havre have the lowest blood levels, although there are no statistically significant differences between zones (table 47). These trends are consistent with the exposures calculated by the indirect approach, with the exception of cadmium. We calculated an average exposure in Le Havre of $4.64 \mu \mathrm{~g} \mathrm{Cd} / \mathrm{kg} \mathrm{bw} /$ week, which is 2 to 6 times more than in the other zones, whereas the biological results suggest an equivalent exposure in the four study zones (differences not statistically significant). The north-south gradient that appeared significantly with the indirect approach for cadmium is therefore not reflected in the biological level results. For lead, mercury and methylmercury we find similar trends between the level of dietary intake and the biological exposure, with a minimum in Le Havre and a maximum in La Rochelle. For lead in particular it would appear that the subjects in La Rochelle are often the most exposed (indirect approach and direct approach). This is confirmed by the statistical analysis ( $\mathrm{p}<0.05$ ).

In the cases of arsenic, cadmium and lead, there are no significant differences in concentrations in urine samples between the study zones, despite the fact that dietary exposure to inorganic arsenic is significantly higher in Lorient and significantly lower in La Rochelle.

Table 47: Blood and urinary concentrations in trace elements of the subjects of all areas regardless of the age and sex (Mean $\pm$ SD)

|  | Le Havre | Lorient | La Rochelle | Toulon | All subjects |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Blood concentrations | $\mathrm{n}=83$ | $\mathrm{n}=115$ | $\mathrm{n}=98$ | $\mathrm{n}=95$ | $\mathrm{n}=391$ |
| Cd ( $\mu \mathrm{g} / \mathrm{L}$ Blood) | $0.51 \pm 0.41^{\text {a }}$ | $0.63 \pm 0.59^{\text {a }}$ | $0.69 \pm 0.66^{\text {a }}$ | $0.65 \pm 0.63{ }^{\text {a }}$ | $0.62 \pm 0.58$ |
| Pb ( $\mu \mathrm{g} / \mathrm{L}$ Blood) | $31.7 \pm 16.3^{\text {a }}$ | $31.9 \pm 19.8{ }^{\text {a }}$ | $49.8 \pm 28.5{ }^{\text {b }}$ | $38.2 \pm 24.0^{\text {a }}$ | $37.8 \pm 23.8$ |
| $\mathrm{Hg}_{\mathrm{T}}$ ( $\mu \mathrm{g} / \mathrm{L}$ Blood) | $2.76 \pm 2.16^{\text {a }}$ | $3.65 \pm 2.33^{\text {a. b }}$ | $4.59 \pm 3.24{ }^{\text {b }}$ | $3.13 \pm 4.68{ }^{\text {a }}$ | $3.57 \pm 3.30$ |
| MeHg ( $\mu \mathrm{g} / \mathrm{L}$ Blood) | $3.37 \pm 2.51^{\text {a }}$ | $3.68 \pm 2.36{ }^{\text {a }}$ | $3.82 \pm 2.78{ }^{\text {a }}$ | $4.01 \pm 6.19^{\text {a }}$ | $3.73 \pm 3.76$ |
| Urinary concentrations | $\mathrm{n}=84$ | $\mathrm{n}=115$ | $\mathrm{n}=97$ | $\mathrm{n}=91$ | $\mathrm{n}=387$ |
| As $_{T} /$ Creat ( $\mu \mathrm{g} / \mathrm{g}$ Creat) | $53.7 \pm 78.4^{\text {a }}$ | $77.1 \pm 137.4^{\text {a }}$ | $116.1 \pm 318.5^{\text {a }}$ | $89.7 \pm 251.0^{\text {a }}$ | $84.8 \pm 217.5$ |
| Cd/Creat ( $\mu \mathrm{g} / \mathrm{g}$ Creat) | $0.6 \pm 0.4^{\text {a }}$ | $0.7 \pm 0.4^{\text {a }}$ | $0.6 \pm 0.5^{\text {a }}$ | $0.7 \pm 0.5^{\text {a }}$ | $0.7 \pm 0.5$ |
| Pb ( $\mu \mathrm{g} / \mathrm{g}$ Creat) | $4.8 \pm 2.7{ }^{\text {a }}$ | $6.0 \pm 4.0^{\text {a }}$ | $5.5 \pm 4.8{ }^{\text {a }}$ | $6.2 \pm 5.4^{\text {a }}$ | $5.7 \pm 4.4$ |
| Speciation in urines | $\mathrm{n}=15$ | $\mathrm{n}=33$ | $\mathrm{n}=32$ | $\mathrm{n}=21$ | $\mathrm{n}=101$ |
| As inorg ( $\mu \mathrm{g} / \mathrm{L}$ urine) | $24.6 \pm 21.5^{\text {a }}$ | $19.9 \pm 11.2^{\text {a }}$ | $31.8 \pm 26.7^{\text {a }}$ | $38.2 \pm 41.6^{\text {a }}$ | $28.2 \pm 26.8$ |
| As inorg $/$ Creat ( $\mu \mathrm{g} / \mathrm{g}$ Creat) | $25.7 \pm 24.8^{\text {a }}$ | $22.8 \pm 12.3^{\text {a }}$ | $26.0 \pm 24.1^{\text {a }}$ | $33.7 \pm 28.9^{\text {a }}$ | $26.5 \pm 22.3$ |

$A s_{T}$ : total arsenic. As $s_{\text {inorg }}$ : inorganic arsenic (As(III), As(V) and their metabolits MMA(V) and DMA(V)), HgT: total mercury, Creat: creatinine. Values in the same raw with different superscript letters are significantly different, p<0.05 (Tukey's test)

### 4.3 Exposure to persistent organic pollutants

### 4.3.1 Food exposure

These results correspond to the exposure to persistent organic pollutants through fish and seafood consumption only, not through total diet; they also exclude environmental exposure via the respiratory tract. However food is the principal vector accounting for more than $90 \%$ of total exposure of the population to PCDD/Fs and DL-PCBs ${ }^{121}$, and fish and seafood are found to be the most contaminated products (see Introduction).

Tables 48 and 49 present the exposure to dioxins and furans, DL-PCBs, i-PCBs and PBDEs of the different population groups in each study zone. The Appendix 5 and 6 presents for all the consumers in all the zones the main contributions (as percentages) to the total exposure for each class of pollutants and to TRV when they do exist.

PCDD/F and DL-PCB : Only the subjects in Toulon and women of child-bearing age in Lorient have an average exposure to dioxins, furans and DL-PCBs less than the WHO's PTMI of 70 pg TEQ $_{w H O} / \mathrm{kg} \mathrm{bw} / \mathrm{month}$. However $62 \%$ of the subjects have an exposure through their fish and seafood consumptions less than the PTMI. Clearly the average is strongly influenced by certain high values; the statistical distribution is not symmetric. This average exposure ranges from 9.70 to 20.0 pg TEQ $_{\text {who }} / \mathrm{kg}$ bw/week in adult males and from 11.9 to $27.1 \mathrm{pg} \mathrm{TEQ}_{\text {wHo }} / \mathrm{kg}$ bw/week in women. Elderly subjects have the highest exposure, in particular those in Le Havre (average 32.0 pg TEQ ${ }_{\text {who }} / \mathrm{kg}$ bw/week, or 109 pg TEQ ${ }_{\text {who }} / \mathrm{kg}$ bw/week at P95).

However we should underline that the real exposure to dioxins is almost certainly overestimated since the cooking of fish and seafood reduces the PCDDs level, as pointed out by Hori and his team ${ }^{122}$. The DL-PCBs account for $76 \%$ of the total exposure and the PCDD/F account for $24 \%$ which is consistent with the conclusions of the report of the Afssa (2006) ${ }^{121}$.

The main contributors to PCDD/Fs and DL-PCBs exposure are sardine (19\%), salmon (14\%), seabass (7\%), mackerel (7\%), and seabream (5\%). The swimcrab accounts for almost 5\% of the exposure on average in the 4 zones; but it is an important contributor to PCDD/Fs and DL-PCBs exposure only in Le Havre (16\%). Similarly, eel is a major contributor to DL-PCB exposure only in La Rochelle (18\%).
i-PCB : Only 278 people ( $28 \%$ ) have a i-PCBs exposure through their fish and seafood consumption less than the TDI of $0.02 \mu \mathrm{~g} / \mathrm{kg}$ bw/day. The highest average exposure is that of elderly people in Le Havre with more than $0.67 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} /$ week; the P95 level of these same subjects is $2.36 \mu \mathrm{~g} / \mathrm{kg}$ bw/week. The main contributors to this exposure are sardine ( $20 \%$ ), salmon ( $13 \%$ ), seabass ( $8 \%$ ) and mackerel ( $7 \%$ ). The eel is a major contributor in La Rochelle ( $26 \%$ ) and the swimcrab in Le Havre ( $16 \%$ ), but not in the other zones.

In subjects exceeding the TRV of PCDD/Fs and DL-PCBs or i-PCBs, we find the same major contributors to the exposure, the most important being the sardine ( $23 \%$ for PCDD/Fs, DL-PCBs and i-PCBs) and the eel ( $16 \%$ for PCDD/Fs and DL-PCBs, $12 \%$ for $i-P C B s$ ).

PBDE : The average exposure to PBDEs ( $28,47,99,100,153,154,183$ ) ranges from 1.61 to $2.39 \mathrm{ng} / \mathrm{kg}$ bw/day for men aged 18 to 64 years and from 1.98 to $2.58 \mathrm{ng} / \mathrm{kg}$ bw/day for women in the same age group. The most exposed adult females in Le Havre (P95) have an exposure of $5.95 \mathrm{ng} / \mathrm{kg}$ bw/day. The main contributors to this exposure are salmon (19\%), mackerel ( $9 \%$ ), cod ( $6 \%$ ), sardine ( $7 \%$ ) and tuna (5\%). Once again, in La Rochelle we find eel is a major contributor (12\%) to PBDEs exposure.

[^23]Table 48: Food exposure of high fish and seafood consumers to persistent organic pollutants

|  |  |  | $\begin{gathered} \text { PCDD/F } \\ \text { (pg TEQ } \\ \text { kg bws/week) } \end{gathered}$ | $\begin{gathered} \text { PCB-DL } \\ \text { (pg/TEQ oms } \\ \text { /kg bw/week) } \end{gathered}$ | Total diox (pg/TEQ ${ }_{\text {oms }}$ / kg bw/week) | iPCB <br> ( $\mu \mathrm{g} / \mathrm{kg}$ bw/week) | PBDE (ng/kg bw/day) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & 0 \\ & \frac{1}{2} \\ & \frac{\pi}{1} \\ & 9 \end{aligned}$ | Adult men$(18-64 y) n=45$ | Mean | 4.15 | 13.64 | 17.79 | 0.38 | 1.74 |
|  |  | SD | 3.91 | 13.41 | 17.18 | 0.38 | 1.06 |
|  |  | P95 | 8.88 | 43.47 | 51.79 | 1.21 | 3.93 |
|  | Adult women$(18-64 y) n=180$ | Mean | 5.78 | 21.31 | 27.09 | 0.51 | 2.34 |
|  |  | SD | 6.47 | 52.48 | 56.98 | 0.59 | 1.84 |
|  |  | P95 | 16.58 | 56.43 | 71.57 | 1.60 | 5.95 |
|  | Older subjects (65 y and more) $\mathrm{n}=\mathbf{2 6}$ | Mean | 7.67 | 24.31 | 31.98 | 0.67 | 2.23 |
|  |  | SD | 7.97 | 25.19 | 33.15 | 0.72 | 1.52 |
|  |  | P95 | 26.51 | 82.05 | 108.55 | 2.36 | 5.25 |
|  | Women of childbearing age (18-44 $y$ ) $n=98$ | Mean | 4.92 | 15.93 | 20.85 | 0.47 | 2.32 |
|  |  | SD | 5.16 | 17.63 | 22.53 | 0.56 | 2.04 |
|  |  | P95 | 15.94 | 56.68 | 70.69 | 1.59 | 7.04 |
| $\begin{aligned} & \text { を } \\ & \text { © } \\ & \hline 0 \end{aligned}$ | Adult men$(18-64 y) n=53$ | Mean | 4.78 | 15.27 | 20.05 | 0.40 | 2.17 |
|  |  | SD | 0.69 | 4.49 | 5.18 | 0.11 | 0.11 |
|  |  | P95 | 12.20 | 40.61 | 53.59 | 1.15 | 5.06 |
|  | Adult women$(18-64 y) n=159$ | Mean | 4.89 | 14.60 | 19.49 | 0.34 | 2.16 |
|  |  | SD | 4.04 | 13.95 | 17.71 | 0.35 | 1.60 |
|  |  | P95 | 11.89 | 42.80 | 56.74 | 0.97 | 4.31 |
|  | Older subjects (65 y and more) $n=37$ | Mean | 4.68 | 14.04 | 18.73 | 0.33 | 1.98 |
|  |  | SD | 3.00 | 9.72 | 12.66 | 0.26 | 1.04 |
|  |  | P95 | 11.26 | 35.97 | 48.05 | 0.86 | 3.59 |
|  | Women of childbearing age ( $18-44 \mathrm{y}$ ) $\mathrm{n}=77$ | Mean | 3.95 | 11.63 | 15.58 | 0.27 | 2.10 |
|  |  | SD | 3.21 | 13.21 | 16.17 | 0.35 | 1.75 |
|  |  | P95 | 9.10 | 24.90 | 33.79 | 0.64 | 4.04 |
|  | Adult men$(18-64 y) n=88$ | Mean | 3.94 | 15.98 | 19.92 | 0.54 | 2.39 |
|  |  | SD | 3.63 | 24.70 | 27.54 | 1.12 | 3.07 |
|  |  | P95 | 8.62 | 33.43 | 42.52 | 1.09 | 5.06 |
|  | Adult women$(18-64 y) n=125$ | Mean | 5.20 | 16.23 | 21.43 | 0.46 | 2.58 |
|  |  | SD | 3.72 | 14.58 | 17.60 | 0.52 | 1.88 |
|  |  | P95 | 12.41 | 48.70 | 56.76 | 1.42 | 6.08 |
|  | Older subjects (65 y and more) $n=40$ | Mean | 4.47 | 16.00 | 20.47 | 0.47 | 2.21 |
|  |  | SD | 3.13 | 11.49 | 14.07 | 0.41 | 1.40 |
|  |  | P95 | 9.77 | 38.73 | 43.95 | 1.53 | 4.59 |
|  | Women of childbearing age ( $18-44 \mathrm{y}) \mathrm{n}=79$ | Mean | 4.90 | 13.80 | 18.70 | 0.39 | 2.53 |
|  |  | SD | 3.96 | 13.94 | 17.19 | 0.49 | 2.04 |
|  |  | P95 | 12.81 | 38.83 | 51.83 | 1.36 | 6.17 |
| $\begin{aligned} & \text { 들 } \\ & \text { 을 } \end{aligned}$ | Adult men$(18-64 y) n=60$ | Mean | 2.04 | 7.65 | 9.69 | 0.22 | 1.61 |
|  |  | SD | 1.35 | 5.86 | 7.12 | 0.21 | 1.09 |
|  |  | P95 | 4.38 | 17.10 | 21.22 | 0.56 | 3.51 |
|  | Adult women$(18-64 y) n=177$ | Mean | 2.47 | 9.39 | 11.86 | 0.27 | 1.98 |
|  |  | SD | 1.89 | 9.61 | 11.26 | 0.38 | 1.60 |
|  |  | P95 | 6.67 | 25.48 | 31.89 | 0.85 | 4.77 |
|  | Older subjects ( 65 y and more) $\mathrm{n}=21$ | Mean | 2.44 | 9.23 | 11.67 | 0.25 | 1.58 |
|  |  | SD | 1.16 | 5.82 | 6.88 | 0.18 | 0.73 |
|  |  | P95 | 4.04 | 15.15 | 19.20 | 0.35 | 2.55 |
|  | Women of childbearing age (18-44 $y$ ) $n=96$ | Mean | 2.25 | 8.53 | 10.78 | 0.25 | 1.97 |
|  |  | SD | 1.80 | 10.54 | 12.02 | 0.45 | 1.76 |
|  |  | P95 | 5.49 | 19.79 | 24.44 | 0.72 | 3.74 |

Table 49: Food exposure to persistent organic pollutants of high fish and seafood consumers of all areas regardless of the age and sex (Mean $\pm$ SD)

|  | Le Havre $n=249$ | Lorient $n=247$ | La Rochelle $n=248$ | Toulon $n=252$ | All subjects n=996 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| PCDD/F (pg TEQ ${ }_{\text {oms }} / \mathrm{kg}$ bw/week) | $5.58 \pm 6.01^{\text {a }}$ | $4.84 \pm 3.79^{\text {a }}$ | $4.65 \pm 3.63{ }^{\text {a }}$ | $2.37 \pm 1.73^{\text {b }}$ | $4.34 \pm 4.25$ |
| PCB-DL (pg TEQ oms $^{\text {/kg bw/week) }}$ | $17.7 \pm 19.3^{\text {a }}$ | $14.7 \pm 13.0^{\text {a }}$ | $16.1 \pm 18.4{ }^{\text {a }}$ | $8.96 \pm 8.59^{\text {b }}$ | $14.3 \pm 15.7$ |
| Total PCDD/F et PCB-DL ( $\mathrm{pg} \mathrm{TEQ}_{\text {омs }} / \mathrm{kg} \mathrm{bw} /$ week) | $23.3 \pm 25.2^{\text {a }}$ | $19.5 \pm 16.6^{\text {a }}$ | $20.8 \pm 21.1^{\text {a }}$ | $11.3 \pm 10.1^{\text {b }}$ | $18.7 \pm 19.6$ |
| iPCB ( $\mu \mathrm{g} / \mathrm{kg}$ bw/week) | $0.53 \pm 0.58^{\text {a }}$ | $0.35 \pm 0.35^{\text {b }}$ | $0.49 \pm 0.77{ }^{\text {a }}$ | $0.26 \pm 0.33^{\text {b }}$ | $0.40 \pm 0.55$ |
| PBDE ( $\mathrm{ng} / \mathrm{kg}$ bw/day) | $2.23 \pm 1.70^{\text {a.b }}$ | $2.14 \pm 1.50{ }^{\text {a.b }}$ | $2.45 \pm 2.31{ }^{\text {a }}$ | $1.86 \pm 1.44^{\text {b }}$ | $2.17 \pm 1.78$ |
| Values in the same raw with different superscript letters are significantly different, $\mathrm{p}<0.05$ (Tukey's test) |  |  |  |  |  |

For all the classes of pollutant we observe a similar trend: the subjects in Toulon are less exposed than those in the other zones.

This trend is significant for dioxins, furans and DL-PCBs: the subjects in Toulon are significantly less exposed than the subjects in the other zones, without distinction of age or sex ( $p<0.05$ ). For the i-PCBs, subjects in both Toulon and Lorient have a significantly lower exposure than those in the two other zones. Moreover, for the dioxins, furans and i-PCBs, the subjects in Le Havre are more exposed than those in Lorient and La Rochelle, although this difference is not statistically significant.

Finally, the average exposure to PBDEs (28, 47, 99, 100, 153, 154, 183) is significantly less in Toulon than in La Rochelle ( $p<0.05$ ), but it is equivalent to that in the two other zones.

Generally for food exposure we find a north-south gradient like the one observed for seafood contamination by POPs.

## FIFTH PART

## Perception of risks

In addition to the food frequency questionnaire, questions on the perception of food risks in general and on the risks associated with seafood in particular were asked to the respondents in order to better appreciate their knowledge, their attitudes and their selection criteria, in particular those of coastal dwellers and high seafood consumers. The subjects were questioned during a period when no serious food crisis was in the news.

### 5.1 General dietary risks

A first general question on the perception of dietary risks formulated identically during the INCA 99 national study reveals the awareness of the existence of risks linked to food, although this is somewhat moderate.
"In your opinion, do today's food products present a health risk?"

Figure 13 : High seafood consumers' perception of the health risk of food products (\% of replies)


Almost half the participants believe there is a small risk, $35 \%$ think there is a major risk, and about $11 \%$ think there is no risk at all.
"Among the risks associated with food, in your opinion which, among the following proposals, is the one that today represents the most worrying health risk?"

## En premier?

Figure 14 : The most worrying health risks associated with food - first choice (\% of replies)


Almost 23\% of the respondents think that contamination of food by pesticides and chemical products represents the greatest health risk today (figure 14). This is followed by GMOs (17\%) then obesity (16\%) having received wide media coverage recently; mad cow disease is in 4th position (13\%). We note that during the pilot survey, mad cow disease topped this list of the most serious risks and was mentioned as by $30 \%$ of respondents. This illustrates the evolution of food concerns and the preoccupations of the population as regards health and food safety.

The other risks were all mentioned by less than $10 \%$ of the participants.

## Second choice?

Figure 15 : The most worrying health risks associated with food - second choice (\% of replies)


Chemical products appear to be a matter of concern for a majority of respondents ( $18 \%$ ) since these are found once again at the top of the list of the most serious health risks (figure 15). Environmental contaminants, the use of growth hormones in farms, cardiovascular diseases and GMOs are also mentioned by more than $10 \%$ of the participants.

The "Other risks" mentioned include cancer, avian influenza, food allergies and infantile obesity. The fact that avian influenza was little mentioned is explained by the fact that the survey was made prior to recent intense media coverage of this problem.

These results are consistent with those of the IRSN barometer of risk perception performed during the same year. In this study, the most frequently mentioned hazards or risks potentially linked to food are obesity and pesticides and, to a lesser degree, GMOs ${ }^{123}$.

### 5.2 Risks associated with marine pollution

"In your opinion, does marine pollution present a major health risk through seafood consumption?"

Of the people questioned, $81.1 \%$ believe marine pollution can present health risks through food consumption. This general awareness of the potential health risks relating to maritime pollution and possible contamination of marine products is no doubt due to the fact that the respondents reside in coastal areas. Only $13.6 \%$ of the respondents think that marine pollution presents no major health risk. This high level of risk awareness in relation to fish and seafood consumption, without prejudice to their seriousness, contrasts with the results of other studies that often reveal a phenomenon of habituation to risks, or even denial of these risks by the people exposed. This has been observed in risk perception studies around listed installations, notably nuclear power plants ${ }^{124}$, and in studies on the perception of risks associated with "mad cow disease" which revealed the absence of changes in consumption habits among the highest meat consumers during the crisis, whereas occasional consumers reduced their consumption ${ }^{125}$.

This general perception of the risk is accompanied by very good knowledge of the environmental contaminants contained in fish and seafood.
"In your opinion, which pollutants can be present in seafood?"(several replies possible). The results correspond to the aggregated replies.

Figure 16 : Pollutants present in the sea, according to the respondents (\% of replies)


[^24]Hydrocarbons are the best-known pollutants, in particular among people living near the coast who are very sensitive to marine pollution issues, no doubt due to their experience of oil spillages, notably that of the oil tanker "Erika" recently. Heavy metals (mercury and lead) are mentioned in second place. The frequent mention of lead is rather surprising in view of the moderate contribution of marine products to tolerable intake to lead ( $<2 \%$ PTWI, see Appendix 6). This can be explained by the bad reputation of lead as an environmental contaminant. Dioxins, which had been the subject of much media coverage a few months earlier, are also often mentioned as marine pollutants.

The "Other" pollutants mentioned include toxic algae, radioactivity, pesticides, general waste (plastics, metals) and bacteria.

### 5.3 Behaviour of seafood consumers

"Do you pay particular attention to the origin of the seafood you buy, in the sense that you know some coastal zones can be more polluted than others?"

More than the half the respondents ( $53 \%$, or 539 ) say they attach importance to the origin of the seafood they buy. Among them, $4 \%$ said the origin was the first criterion when selecting seafood and $22 \%$ said it was the second criterion. This sensitivity to the origin of the fishing is confirmed by OFIMER studies that reveal the importance of consumer information, notably for well-informed consumers who constitute a large part of our study population. However, we shall see later that the geographic origin is of little importance compared to the main criteria: appearance, freshness, price and season.
"Which seafood origins do you avoid?" (several replies possible). The results correspond to the aggregated replies.

Figure 17 : Seafood origins avoided by high consumers (\% of replies)


The Mediterranean Sea is the most avoided when purchasing seafood (18\% of respondents), followed by the northern oceans: Northwest and Northeast Atlantic and the Baltic mentioned in about 12\% of replies.
"Among the following criteria, to which do you attach the greatest importance when buying seafood?"

## First choice?

Figure 18 : Consumers' first selection criterion when purchasing seafood (\% of replies)


## Second choice?

Figure 19 : Consumers' second selection criterion when purchasing seafood (\% of replies)


Freshness and appearance are undeniably the prime purchasing criteria for $80 \%$ of respondents. Price is the second criterion for $32 \%$ of respondents and to a lesser degree the season (18\%), the taste (16\%) and the origin (13\%). These are therefore consumers who trust their own judgement when making their choices.

Very few people appear to be concerned by the nutritional value of the products bought. We have seen that the fatty acid composition of fish is nevertheless highly variable and could interest the consumer. It appears that for the moment this type of very important information for nutritionists, dieticians and public health professionals in general is not well integrated in consumers' criteria when selecting products, despite the increasing reputation of omega 3 fatty acids.

The "Other" criteria mentioned include product availability, impulsive desire, presence/absence of bones, use-by date and odour. Ease of preparation is not often mentioned, although has been shown elsewhere that this is an important selection criterion for consumers in general. Perhaps the fact that our study population are high fish and seafood consumers means they are not put off by the preparation of whole fish, a supposition well supported by OFIMER studies ${ }^{126}$.

### 5.4 Consumer information

"To ensure effective information of consumers, since 1st January 2002 community regulations require that seafood labelling includes the commercial denomination of the species, the production method and the zone of capture. Are you aware of this?"

Almost the two thirds of the people questioned ( $62.8 \%$ ) are aware of the new regulation now applicable to seafood labelling. This result is not surprising in view of the high frequency of purchases of fish and seafood by the respondents in the study.
"Do you think this measure will affect your buying habits?"
Of the people questioned, $47 \%$ or 473 think that the new regulation (of which $71 \%$ were aware) will influence their buying habits, while $46 \%$ or 460 (of which $60 \%$ were aware of the regulation) do not think so.
"Do you think that the controls carried out at seafood points of sale are sufficient or insufficient?"
The controls made on seafood appear to reassure only one third of the respondents. More than half of them ( $50.7 \%$ ) believe these controls to be insufficient, and $16 \%$ have no opinion on the subject. This result no doubt reflects the visible preoccupation of this population of high fish and seafood consumers as regards possible environmental contamination of the fish, crustaceans or molluscs they consume.
"Do you think that consumption of cultivated fish is safer than consumption of wild fish?"
Figure 20 : High consumers' opinions as regards the risks of consuming cultivated fish (\% of replies)


A majority of respondents believe consumption of wild fish is safer ( $41 \%$ ), or at least no less safe than consuming cultivated fish (38\%) (figure 20).

Various media alerts, notably that stirred up by the publication in January 2004 of an article in "Science" about the contamination of farmed salmon in Europe, explains this undercurrent of suspicion as regards fish farming. In this media context, however, we note that 4 out of 10 of our subjects consider that farmed fish and wild fish are equivalent.
"If so, why?" (several replies possible) The results correspond to the aggregated replies.

Figure 21 : Reasons given for believing that consumption of farmed fish is safer then consumption of wild fish (\% of replies)


Almost half the respondents believe that consumption of farmed fish is safer thanks to better control and regular surveillance of the fish (figure 21). A quarter of them are reassured by the fact that the marine milieu in which these fish live is not only controlled and regularly analysed but also by the fact that this environment is less exposed to external pollution (boats, toxic waste, etc.). More than 20\% of them believe eating farmed fish is better for health: more natural, more healthy, better monitoring, etc.

About 5\% of respondents support the idea that consumption of farmed fish is safer, but without knowing precisely why. Finally, 4\% give philosophical reasons, prejudices or information read or heard in the media.

In conclusion, this part of the study on perceptions and attitudes of high fish and seafood consumers as regards the possible risks of their consumption shows that this population has extensive knowledge of fish and seafood. They are consumers seeking information, preoccupied, but they remain somewhat detached from public controversy on this subject.

## SIXTH PART



### 6.1 Fatty acids

The lipids of the erythrocyte membranes of the subjects contain $0.74 \%$ of EPA, $1.71 \%$ of DPA and $4.02 \%$ of DHA, or $6.47 \%$ of $n-3$ LC-PUFAs. Few data are available on this subject in the scientific literature, and when they do exist they are not comparable with ours since they generally come from intervention studies involving omega 3 supplementation of subjects, or they do not concern the general population, or they do not use the same biomarker of exposure.

In controls of an intervention study, Weill et al. in 2002 found DHA accounted for $4.8 \pm 0.89 \%$ of total lipids in the red blood cell membrane, and EPA for $0.5 \pm 0.12 \%{ }^{88}$. In 2004, Payet et al. measured $4.8 \pm 1.2 \%$ of DHA and $1.1 \pm 0.81 \%$ of EPA in elderly people before supplementation ${ }^{127}$. In 2003, Dewailly and his team reported $\mathrm{n}-3$ LC-PUFAs levels of $1.8 \%, 3.9 \%$ and $8.0 \%$ in circulating phospholipids in plasma in three Canadian population groups consuming respectively $13 \mathrm{~g}, 60 \mathrm{~g}$ and 131 g of fish per day ${ }^{128}$.

The trends observed in the results of the measurements in the erythrocytes for the different zones (Tables 34 to 38 ) are not completely in agreement with the intakes calculated by crossing consumption and composition data (Tables 29 to 33 ).

The relationship between n-3 LC-PUFAs intake and the erythrocyte composition is not proportional, which implies that many other factors are acting. Knowing that $\mathrm{n}-3$ fatty acids are subject to betaoxidation, the quantity of total lipids in the diet influences the oxidation of the n-3 LC-PUFAs. We note a Pearson correlation coefficient of 0.26 ( $p<0.0001$ ) between the $n-3$ LC-PUFAs levels in the erythrocytes and the proportion of $n-3$ LC-PUFAs in the dietary lipids.

Moreover, since age influences peroxidation of lipids, in particular the n-3 LC-PUFAs, the correlation between age and the erythrocyte $n-3$ LC-PUFAs levels is 0.23 ( $p<0.0001$ ). For smokers the number of cigarettes smoked per day correlates negatively with the blood DHA ( $r=-0.22, p=0.02$ ).

It is particularly difficult to find a simple mathematical model linking the erythrocyte composition to n 3 LC-PUFAs consumption, which may be explained in several ways:

- The first hypothesis is that the fatty acids composition of the blood tissue (erythrocyte) is not a good marker for all $n-3$ fatty acids. It is accepted that the DHA of the erythrocytes reflects very poorly the ingested quantity, unlike other n-3 PUFAs ${ }^{129}$. In man and general population, we must nevertheless settle for this accessible tissue.
- The second hypothesis is the variability between individuals related to beta-oxidation of fatty acids and its multiple regulation factors. In the case of fatty acids, although they are essential, their availability is highly dependant on their energy usage by beta-oxidation, which is not the case for many other essential nutrients, including vitamins and minerals. This beta-oxidation, which does not spare the $n-3$ PUFAs, can represent up to $90 \%$ of the $n-3$ metabolism; it is influenced by the physiological and physiopathological status of the energy expenditure and the fatty acids composition (quantity of saturated acids) and other energy-providing nutrients in the diet.
- The third hypothesis is that the presence of the precursor alpha-linolenic acid (ALA) in the diet can also bring non-negligible additional quantities of $n-3$ LC-PUFAs (EPA and DHA) ${ }^{88}$. This could explain why the French population, even heavy fish consumers, has an n-3 LC-PUFAs status lower than other

[^25]populations, in view of the small quantities of ALA consumed ${ }^{4}$ and the richness of the diet in $n-6$ fatty acids that inhibit the ALA conversion. By way of comparison, we could assume that a population consuming the precursor ALA (Canada, Japan) and little LA would increase its very long-chain n-3 status by ALA conversion, in this case non-negligible. The study of Weill et al. suggests this idea since the n-3 LC-PUFA status is high with a diet rich in ALA (flax-eating animals products) but in which fish is absent ${ }^{88}$. In our population, those most consumed oils are sunflower and olive which provide $1.0 \pm 1.0 \mathrm{~g} \mathrm{ALA} /$ week (high estimation) and $23.2 \pm 22.6 \mathrm{~g}$ LA/week compared to $11.2 \pm 8.9 \mathrm{~g} \mathrm{ALA} /$ week and $23.5 \pm 18.6 \mathrm{~g}$ LA/week if the subjects consumed only rapeseed oil.

- A fourth explanation is that the variability over a long period of the biological exposure in the general population is less than that found in the intervention studies over short periods. The literature on intervention studies relates that the variations in the erythrocyte membrane composition can be large. Harris reports a $4.3 \%$ increase of EPA + DHA resulting from supplementation of one gram of EPA + DHA over 6 months. For a higher n-3 LC-PUFAs intake ( $0.14 \mathrm{~g} /$ day $)$ Weill noted a $0.69 \%$ increase in the n-3 LC-PUFAs in the erythrocyte membrane. Our subjects are high fish and seafood consumers, but the large panel of products and the differences in the nutritional composition of the products could lead to considerable variation in intakes depending on the products consumed (see the chapter on "Seafood composition and contamination"). Furthermore the biological samples were taken between October and December 2004, which means they reflect summer consumptions, whereas the samples were taken between January and April 2005, a period during which most of the fish tend to display higher total lipid and fatty acid levels than those measured in summer. In addition to this seasonal variation of lipids levels in fish there may also be some seasonality in the consumption of oily fish. These factors could well imply that the erythrocyte membrane is not an ideal marker of long-term n-3 LC-PUFAs consumption, but in fact only a short-term marker as suggested by certain studies ${ }^{130}$. Intervention studies presenting significant results are often made over a few days or even a few weeks, but rarely over a duration corresponding to the hematopoietic cycle.
- A last hypothesis is that the relationship between the DHA consumed and the proportion of DHA in the erythrocyte membrane is not linear. It could be logarithmic, polynomial...


### 6.2 Trace elements

Arsenic: Concerning the levels found in food samples, we noted that the total of the contents of the different arsenic forms was not always equal to the $A s_{T}$ level. When the sum is less than the As level, the difference may be explained by the non-detection of certain arsenic species which induces a slight underestimation of the total of the arsenic species relative to the total level. In the opposite case, this is mainly due to the fact that the $\mathrm{As}_{\mathrm{T}}$ and the forms of speciation were quantified using two distinct analytical techniques, although the observed differences generally remain within the estimated limits of measuring uncertainty of the methods.

The total arsenic concentrations measured in urine are relatively high. This is not surprising in view of the diet rich in seafood of the subjects tested. In effect, organic arsenic is of dietary origin (AsB in particular), which accounts for a large part of total arsenic in the urine, reflects the intake at the last meal. It is therefore important to take the arsenic speciation into account here.

Regarding indirect exposure, note that the inorganic arsenic corresponds to the forms As(III) and As $(\mathrm{V})$. In toxicology and occupational medicine (direct exposure or biomarkers) inorganic arsenic is understood to mean the $\mathrm{As}(\mathrm{III})$ and $\mathrm{As}(\mathrm{V})$ forms and their mono- and dimethyl metabolic derivatives, MMA and DMA of redox potential V. Nevertheless, these methylated forms are in a minority.

There is no significant correlation between arsenic (total and inorganic) consumed and arsenic found in the urine, but the quantity of predator fish consumed correlates significantly with the inorganic arsenic measured in the urine, weighted by the creatinine ( $r=0.25, p=0.01$ ).

These results are highly consistent in view of the very low inorganic arsenic intakes from fish and seafood compared to other sources (water, etc.).

Mercury : The analyses of mercury in seafood products, either to determine the total level or the methylmercury, were self-checked on each series of tests by internal quality controls (IQC) performed using suitable certified reference material (CRM). In $83 \%$ of cases, the average measured concentration of total Hg tallies with that of the MeHg concentrations, allowing for the respective enlarged uncertainties of the two analytical methods used. The dispersion of these values is identical in both cases and is Gaussian. Nevertheless in $17 \%$ of cases there subsists significant differences between the total and methylmercury concentrations. To explain these differences and notably the fact that the MeHg levels can appear higher than the total mercury, we must take into consideration the possible existence of a analytical bias in the isotopic dilution method, although this approach is recognised as one of the best. In current research work it is now envisaged that native mercury may behave differently from mercury added during the isotopic dilution method. The analytical biases usually corrected by this method are therefore no longer corrected perfectly.

Generally speaking, the calculated exposures are higher than the estimations found in the international literature which often indicate that very few consumers exceed the PTWI. However the calculations are often made for populations that are not high fish consumers, and the studies often concern only a small number of products (less than 30, versus 84 in our study). In 1995 Buzina et al. reported, for Adriatic populations consuming fish and seafood 2 to 6 times per week, exposures ranging from 132 to $294 \mu \mathrm{~g}$ $\mathrm{Hg} /$ week on average ${ }^{131}$, or 1.9 to $4.2 \mu \mathrm{Hg} / \mathrm{kg}$ bw/week for individuals of 70 kg aged 15 to 59 years, which is of the same order of magnitude as the results of our study. This comparison therefore illustrates well that our study concerns mainly high seafood consumers. However, the consumption study method based on a food frequency questionnaire, it is probable that the consumptions and therefore the exposures are overestimated.

We note that the calculated mercury exposure $\left(\mathrm{Hg}_{T}\right)$ of the subjects and more particularly the methylmercury exposure appears significantly lower in Le Havre than in the other zones (Table 43). This could be explained by lower consumption of the main contributors to this exposure which are predator fish, notably fresh tuna (mean consumption and consumer rate, see Appendix 2).

As regards the MeHg biomarker of exposure data, these confirm the consequences of high consumption of seafood products and predator fish in particular. In view of the quality procedure applied in the $\mathrm{Hg}_{\boldsymbol{T}}$ analysis of the blood samples, the values of total Hg level obtained are considered to be reliable. Two series of independent measurements were performed by two specialized analytical laboratories. The measurements were made using a reference analytical technique recognised to be reliable and sensitive, and the results tally satisfactorily for at least $80 \%$ of the results around the tolerated confidence interval (Cl) (between $70 \%$ and $130 \%$ of the determined $\mathrm{Hg}_{\top}$ level). The rest of the data outside this interval are essentially low biomarker values, but not high values, which increases the confidence in the interpretation of the data.

On the other hand, the mercury speciation results in the blood samples are more difficult to interpret in that the technique was developed during this study.

The average ratio between the MeHg and HgT concentrations is $115 \%$, i.e. a $15 \%$ difference. About $30 \%$ of the samples analysed have quantified MeHg levels that lie outside the tolerated confidence interval $\mathrm{Cl}, \mathrm{Cl}$ being determined from the estimated uncertainties of the two methods used in order to establish a pertinent internal quality control (IQC). Of these $30 \%, 10 \%$ of the observed differences are less than the lower limit of 0.7 of the Cl , probably due to the small concentrations close to the estimated quantification limit, and $20 \%$ are higher than the upper limit of this CI. This overestimation of the results appear to be linked to the global composition of the blood matrix resulting in different behaviour between added mercury, present in the dissolved phase, and native mercury which is chelated by proteins that can precipitate in an acid milieu (isotopic dilution method). Actually, the quantification by isotopic dilution used during this study offers the possibility of developing a primary reference method enabling greater precision and reliability ${ }^{132} 133134$, however subject to verifying that the enriched isotope added is not extracted differently from the analyte present in the matrix ${ }^{135}$. This is probably what we observe in $20 \%$ of cases on the blood matrix. Further detailed investigations are necessary in order to confirm this and to attempt to correct it.

The MeHg measured in the blood correlates significantly with the quantity of MeHg absorbed from fish and seafood ( $r=0.36, p<0.0001$ ). This conclusion is borne out by the fact that the quantity of predator fish consumed correlates positively with the MeHg measured in the blood ( $r=0.26, p<0.0001$ ). Furthermore we also note a significant correlation between the MeHg level in the blood and the ages of the subjects ( $r=0.25, \mathrm{p}<0.0001$ ). We should bear in mind that our population is relatively homogeneous (only high consumers) and that it is difficult to obtain a very good correlation contrary to a heterogeneous population representing a wide range of consumption and exposure levels.

Moreover the blood analysis results are consistent with the results of the study of Bjornberg et al. (2005) concerning Swedish women of child-bearing age and who are high fish consumers ${ }^{136}$. Such women in our study have an average MeHg level in the blood of 2.3 to $3.4 \mu \mathrm{~g} / \mathrm{l}$ depending on the zone, with a median for all the zones combined of $2.4 \mu \mathrm{~g} / \mathrm{L}$. The Swedish study indicates a median value of $1.7 \mu \mathrm{~g} / \mathrm{L}$ for these high seafood consumers. For comparison, the NHANES study indicates for the general American population (not high fish consumers) a geometric average of $1.02 \mu \mathrm{~g} / \mathrm{L}$ for women aged 16 to 49 years ${ }^{137}$.

The blood analyses results for MeHg in the subjects of our study are reassuring. We recall that a PTWI of $1.6 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} /$ week was established by the JECFA ${ }^{22}$. This corresponds to a concentration of 14 mg of $\mathrm{Hg} / \mathrm{kg}$ in the mother's hair, or $56 \mu \mathrm{~g}$ of $\mathrm{MeHg} / \mathrm{L}$ in the mother's blood (mean ration hair/blood of 250 ), having no adverse effects on the foetus. The PTVI includes an uncertainty factor of 2 corresponding to the inter-individual variability of the relation between MeHg concentration measured in hair and that measured in blood. In fact, even among the very high consumers of fish and seafood, consuming up to 4.5 kg of product per week and with exposures calculated from these high consumptions of as much as

[^26]$9.6 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} /$ week, we find MeHg blood levels well below the value of $56 \mu \mathrm{~g} / \mathrm{L}$, the maximum being 18 $\mu \mathrm{g} / \mathrm{L}$. In fact only one person displayed a concentration exceeding this value, and this can be partly explained by this person's use of a medication affecting the renal and hepathic functions.

Given that the biological results do not indicate that the total Hg and MeHg values associated with toxicity are exceeded, whereas the exposure calculation indicates that a third of the subjects exceed the recommended PTWI for MeHg , to characterise the risk associated with exposure to MeHg we exploit the biomarker results of our study. The measurements in the biological matrices (blood or hair) enable us to calculate, using a pharmacokinetic model, a "steady state dietary intake" which relates the daily intake of MeHg to the concentration in blood or hair, as described by international scientific bodies such as JECFA, EPA, FDA, NRC and $\mathrm{WHO}^{19}{ }^{138}$. The use of biomarker data to estimate exposures requires that the methylmercury concentrations in the blood of our population be effectively in a steady state, which we believe to be the case in view of our recruitment criteria and the homogeneous dietary habits stable over time. Taking only the individuals for which blood analyses were made ( $n=385$ ), by crossing the consumption and contamination data we calculate an average exposure of $1.61 \pm 1.28 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{kg}$ bw/week for the general population excluding women of child-bearing age, and for such women an average exposure of $1.34 \pm 0.92 \mu \mathrm{~g} / \mathrm{kg}$ bw/week. The use of blood concentrations and the singlecompartment pharmacokinetic model indicate for the general population, excluding women of child-bearing age, an average exposure of $0.65 \pm 0.64 \mu \mathrm{~g} / \mathrm{kg}$ bw/week, and for women of child-bearing age, the subjects most "at risk", an average exposure of $0.39 \pm 0.29 \mu \mathrm{~g} / \mathrm{kg}$ bw/week. These average exposures can be compared with the PTWI which takes into account a safety factor of 3.2 corresponding to the inter-individual variability of the pharmacokinetic model. Applying the pharmacokinetic model at individual level this variability is clearly not integrated, but we can consider that, in view of the size of our population, this variability can be ignored for the average of the population. The average intakes of MeHg of $0.39 \pm 0.29 \mu \mathrm{~g} / \mathrm{kg}$ bw/week estimated for women of child-bearing age, and of $0.65 \pm 0.64$ $\mu \mathrm{g} / \mathrm{kg} \mathrm{bw} /$ week estimated for the general population are only half the estimated intakes via indirect exposure.

Table 50 : Modelling of methylmercury exposure at the steady state of subjects participating in the biological part, and quantification of the probability of exceeding the PTWI of $1.6 \mu \mathrm{~g} / \mathrm{kg}$ bw/week

|  | Intake ( $\mu \mathrm{g} / \mathrm{kg}$ bw/week) |  |  | Blood concentrations ( $\mu \mathrm{g} / \mathrm{L}$ ) |  | Exposure calculated with the pharmacokinetic model "Steady State Ingestion" ( $\mu \mathrm{g} / \mathrm{kg}$ bw/week) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean $\pm$ SD | P95 | \%>PTWI | Mean $\pm$ SD | P95 | Mean $\pm$ SD | P95 |
| General population excluding women of childbearing age. $\mathrm{n}=252$ | $1.61 \pm 1.28$ | 3.87 | 37\% | $4.27 \pm 4.34$ | 9.91 | $0.65 \pm 0.64$ | 1.49 |
| Women of childbearing age $(18-44 y), n=132$ | $1.34 \pm 0.92$ | 2.87 | 32\% | $2.70 \pm 2.00$ | 5.61 | $0.39 \pm 0.29$ | 0.85 |

The mean of the individual ratios of calculated exposure to steady-state dietary intake is 4.3 , with a minimum of 0.2 and a P95 of 14 for all the people who were sampled. This ratio is lower, close to $1-2$, when the calculated exposure and the blood MeHg level are high. Two explanations are possible: first, it is possible that the consumptions declared on the food frequency questionnaire by the highest consumers in our study (the people most exposed) are closer to the reality and overestimate the real consumptions less than the declared consumptions of low consumers. Another hypothesis is possible when the consumption overestimation on the questionnaire is the same for all consumers. The absorption of MeHg , its distribution in the blood and its excretion are perhaps dependant on the quantity ingested. This point is interesting to underlined as the linearity of absorption as never been demonstrated. In this case the model describing the steady state would be more suitable for high consumptions (high exposures) than low consumptions.

This comparison between food exposure calculated by consumptions and exposure estimated by application of the pharmacokinetic model shows that the factor of 3.2 applied by the JECFA to take account of the inter-individual pharmacokinetic variability is somewhat protective with regard to high biomarker levels for which the ratio of calculated exposure to steady state dietary intake is around 1 or 2. This is reassuring from a public health point of view.

The fact that the factor between the calculated dietary intakes based on consumptions and the results of the pharmacokinetic model is higher in the JECFA study tends to support the assumption that food frequency questionnaire overestimates consumptions, a point that should be underlined since many exposure studies use this type of questionnaire.

The Codex at its meeting on April 2005 underlined the need to better define the target populations to which the PTWI is applicable, in particular to know whether its PTWI of $1.6 \mu \mathrm{~g} / \mathrm{kg}$ bw$/$ week established in 2003 should be used as a reference toxicological value for the general adult population or if a different PTWI should be defined ${ }^{139}$. This toxicological issue has been submitted for clarification at the next meeting of the JECFA's expert committee scheduled for June $2006{ }^{140}$. The FSA report on the benefits and risks of fish consumption ${ }^{35}$ states that there has been no new published information suggesting that the previous PTWI of $3.3 \mu \mathrm{~g} / \mathrm{kg}$ bw/week established in 2000 was not sufficiently protective for the general population. In this case, on the basis of the calculated exposures in our study for the general population, excluding women of child-bearing age, only $7.9 \%$ of our blood-sampled consumers would exceed the PTWI, instead of $37 \%$.

[^27]Lead : As regards lead, we noted that 22 subjects (6\%) have blood levels higher than the standard (70 to $90 \mu \mathrm{~g} / \mathrm{L}$ ) and that 4 others have urine levels higher than the standard ( $25 \mu \mathrm{~g} / \mathrm{g}$ creatinine). For most of these people, this high concentration can be at least partly explained by professional or leisure activities exposing them to lead (welding, paints, manipulation of metals, hunting, etc.) and/or by the fact that their homes were built before 1948 (after which lead paint was forbidden), although no direct statistical link can be established.

There is a significant correlation between the quantity of lead consumed in seafood and the lead measured in the blood ( $r=0.18, p=0.0005$ ). However this correlation is less marked than for mercury. This is explained by the fact that other sources of lead intake (water, other foods) have not been taken into account here. Age also correlates positively with the presence of lead in the blood ( $\mathrm{r}=0.46, \mathrm{p}<0.0001$ )

Cadmium : The urine analyses reveal that 12 people (3\%) have a cadmium concentration exceeding 2 $\mu \mathrm{g} / \mathrm{g}$ of creatinine, among which 7 are smokers or former smokers with an average age of 52.

As for MeHg we observe that the cadmium results in the biological matrices do not lead to the same interpretations as those for the calculated food exposures. According to our calculations, $8.5 \%$ of the subjects exceed the PTWI, whereas the results of the biological analyses indicate that the cadmium levels remain below the standards.

Some factors have major impact on the cadmium levels measured in the biological matrices. For example the number of cigarettes smoked per day correlates strongly with cadmium level in the blood ( $r=0.62$, $p<0.0001$ ). For non-smokers, age also correlates with these cadmium levels ( $r=0.38, p<0.0001$ ), which is normal for an element that accumulates in the body over time.

Age correlates with cadmium level expressed in $\mu \mathrm{g} \mathrm{Cd} / \mathrm{g}$ creatinine, with a correlation of 0.34 ( $\mathrm{p}<0.0001$ ), which means that the individuals with the highest levels ( $>2 \mu \mathrm{~g} / \mathrm{g}$ creatinine) are over 50 years old.

We note a correlation between the urinary cadmium level and the dietary exposure to cadmium from seafood ( $r=0.32, \mathrm{p}<0.0001$ ). On the other hand, the blood cadmium level does not correlate with the exposure to cadmium ( $p=0.65$ ). This phenomenon may be explained by the fact that seafood are not the only foods contributing to cadmium exposure. The total diet study shows that the main food sources of cadmium are vegetables and potatoes, a long way ahead of crustaceans and molluscs, bread, poultry and offal ${ }^{3}$. Substitutions between consumptions of terrestrial meat products and fish and seafood consumption may explain the absence of a relationship between seafood consumption and blood cadmium level.

Another parameter to be taken into account when comparing the differences between the biological results and the calculated dietary exposures is the difficulty in quantifying the real contamination variability of the seafood products consumed. The contribution of beach fishing to the provisioning, in particular in Lorient and La Rochelle for molluscs and crustaceans, can induce non-negligible variability in the contamination of the consumed foods. Indeed the IFREMER monitoring plans indicate contaminations levels that can correspond to large differences in concentrations from one point of control to another, which is not the case in Toulon for example. It is therefore possible that the subjects
in Toulon are more likely to consume products with trace elements levels close to the average, whereas in La Rochelle and Lorient the food contamination can be much more variable, depending on the provisioning. This may be due to the fact that the ports of Lorient and La Rochelle (and more generally Brittany and the Atlantic coast) commercialise products of more varied origins (fishing zones, foreign boats, etc.) than Mediterranean ports such as Toulon. Consequently applying an average contamination to products in Toulon is without doubt more in line with the reality than doing so in other zones.

### 6.3 Persistent organic polluants

In view of the blood volumes already taken for the analysis of trace elements and the large quantities of blood necessary we preferred not to perform blood analyses. However the calculations of exposure already enable us to raise some points for discussion relative to the existing literature.

PCDD/Fs, DL-PCBs and iPCBs: We find that $39 \%$ of the individuals exceed the PTMI of 70 pg TEQ ${ }_{\text {who }} / \mathrm{kg}$ $\mathrm{bw} / \mathrm{month}$ fixed for PCDD/Fs and DL-PCBs, and that $72 \%$ of them exceed the TDI of $0.02 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} / \mathrm{day}$ fixed for i-PCBs. We must remember that other foods also contribute to the intake of PCDD/Fs, DLPCBs and i-PCBs and that consequently the total exposures are higher. The PCDD/F, DL-PCB and i-PCB contaminations of seafood products are comparable to those measured by the monitoring plans of Administrations. And the PCDD/F and DL-PCB contaminations are within the interval reported by the European Authority in 2005: between 0.3 and 5.8 pg TEQ/g fresh weight ${ }^{141}$, except in the case of the very heavily contaminated eel sample from the Netherlands and some crustaceans rarely consumed, such as swimcrab. The study confirms that even when consuming fish and seafood that comply with the European maximum contamination limits, a high consumer can exceed the JECFA's PTMI, a fact that has already been revealed by other studies ${ }^{142}$. This demonstrates the need to make an effort to reach target values lower than the regulatory limits as rapidly as possible, which is what the new European regulation proposes.

The biggest contributors are oily fish (appendix 5 and 6). The lower exposure of the subjects in Toulon is concomitant with the lower POP contaminations measured in the Toulon samples. Moreover, although the consumption of fish with the highest contaminations is generally equivalent in the four study zones, it is found that in Toulon the consumption of the most contaminated crustaceans (swimcrab, crab, spider crab) is less.

However in no zone in particular do we find a dietary exposure significantly higher than in the other zones.

PBDEs : The average exposure to PBDEs ( $28,47,99,100,153,154,183$ ) is $2.17 \pm 1.78 \mathrm{ng} / \mathrm{kg}$ bw/day, all zones and all subjects included, which is consistent with exposures recently estimated in other countries.

Total Diet Studies (TDS) published in several countries (Canada, USA, Finland, Netherlands, Spain, Sweden, United Kingdom and Japan) report average exposures of 13 to 228 ng PBDE/day ${ }^{114}$. The levels found in our population of high fish consumers range from 139 to 161 ng PBDE/day with an average of $150 \mathrm{ng} / \mathrm{day}$. Our results are therefore very consistent with those other studies using similar methodologies. Our study population consumes on average four times more fish and seafood than the only consumers of the French adult population in the INCA survey (Appendix 3a). We have applied an average deterministic exposure model exploiting the INCA consumption data and our contamination data for fish and seafood

[^28]and those of other product groups contributing to PBDE exposure taken from European studies. This calculation yields an estimated PBDE exposure of about 63 to 142 ng PBDE/day for the French population compared to 172 to $250 \mathrm{ng} /$ day for our study population ${ }^{143}$.

In its evaluation in 2005, the JECFA concluded that the observed exposure of the general population is estimated to be about $4 \mathrm{ng} / \mathrm{kg}$ bw/day, which corresponds to $240 \mathrm{ng} /$ day for a person weighing 60 kg , or slightly more than our calculated exposure. This result is very consistent since the JECFA estimation was not based on fish consumption alone. The JECFA considered that in view of the consequent margin of exposure for a non-genotoxic compound, the current intakes do not appear to be a cause for concern as regards public health ${ }^{114}$.

Today it remains very difficult to measure the PBDEs in biological matrices. Gas phase chromatography coupled with mass spectrometry (GC-NCI-MS or GC-EI-HRMS) is for the moment the most suitable method for detecting PBDEs in matrices of dietary or human origin. But precise evaluation of PBDEs levels in these matrices is hampered by two serious problems: possible contamination of the samples and the technical difficulty of measuring the heaviest compounds.

Nevertheless, there have been some studies of PBDEs levels in maternal milk. The data indicate values ranging from less than $0.1 \mathrm{ng} / \mathrm{g}$ lipids in the earliest studies (in the 1970 s ) to 1.7 to $3.8 \mathrm{ng} / \mathrm{g}$ lipids more recently (1997 to 2003) ${ }^{143}$. Other studies have measured the PBDEs in blood and adipose tissue. The average levels range from 0.4 to $5.6 \mathrm{ng} / \mathrm{g}$ in blood or plasma, and from 0.5 to $11.6 \mathrm{ng} / \mathrm{g}$ of lipids in adipose tissue ${ }^{143}$. Sjodin and his team have moreover showed that a correlation exists between a fish-rich diet and elevation of the measured plasma levels for certain PBDEs ${ }^{144}$. We note that whatever the matrix considered the tissual concentrations in American studies are always much higher than in other studies.

Globally, the data presently available on the concentration in biological tissue are insufficiently documented to enable to establish a relation with the dietary exposure we have observed.

### 6.4 Characterisation of benefits and risks

Table 51 summarises the EPA and DHA intakes and the exposures to various contaminants of the population studied, and the probability of exceeding the recommended intakes for n-3 LC-PUFAs or the TRV.

[^29]Table 51 : Distribution of exposure to omega 3 and toxic elements and probability for high consumers to exceed the recommendations and TRV

| Element | TRV or recommendation | P2.5 | $\begin{aligned} & \text { Exposure } \\ & \text { P50 } \end{aligned}$ | P97.5 | $\begin{aligned} & \%>\text { TRV } \\ & \text { or reco } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| EPA + DHA (mg/d) : | 500 (ISSFAL. 2004) | 255 | 1006 | 1500 | 84 |
| Trace element ( $\mu \mathrm{g} / \mathrm{kg} \mathrm{bw} / \mathrm{week}$ ) : |  |  |  |  |  |
| $\mathrm{As}_{\text {T }}$ | 350 | 15.18 | 66.57 | 254.42 | 0.7 |
| As sinorg | 15 (JECFA. 1989) | 0.14 | 0.51 | 1.92 | 0.0 |
| $\mathrm{Hg}_{T}$ | 5 | 0.28 | 1.21 | 4.68 | 2.1 |
| MeHg | 1.6 (JECFA. 2003) | 0.29 | 1.19 | 4.46 | 34 |
| Pb | 25 (JECFA. 1987) | 0.05 | 0.32 | 1.19 | 0.0 |
| Cd | 7 (JECFA. 2001) | 0.06 | 1.28 | 11.38 | 8.5 |
| Organotin ( $\mu \mathrm{g} \mathrm{Sn/kg} \mathrm{bw/wk)} \mathrm{:}$ |  |  |  |  |  |
| TBT. DBT. TPT and DOT* | 0.72 (AESA. 2004) | 0.01 | 0.05 | 0.20 | 0.0 |
| Persistent organic pollutants : |  |  |  |  |  |
| PCDD/F and PCB-DL (pg TEQomskg bw/month) | 70 (JECFA. 2001) | 8.36 | 54.4 | 381 | 39 |
| iPCB ( $\mu \mathrm{g} / \mathrm{kg}$ bw/day) | 0.02 (WHO. 2003) | 0.006 | 0.034 | 0.42 | 72 |

$\mathrm{As}_{T}$ : total arsenic, $\mathrm{As}_{\text {inorg }}$ : inorganic arsenic, $\mathrm{Hg}_{\mathrm{T}}$ : total mercury, reco: recommendation

* Tributyltin, Dibutyltin, Triphenyltin and Dioctyltin

The vast majority ( $84 \%$ ) of the individuals in our study have EPA and DHA intakes exceeding the recommendations (Table 51), with a daily average of $1,238 \pm 961 \mathrm{mg}$. The people with an EPA and DHA intake less than $500 \mathrm{mg} /$ day consume fish and seafood at least twice a week, which therefore qualifies them as high consumers. Their low intake of $n-3$ LC-PUFAs is explained by the fact that they consume products containing little of these fatty acids, on average 596 g of fish and seafood products per week including 52 g of fatty fish, versus $1,277 \mathrm{~g}$ of fish and seafood including 277 g of fatty fish for the people whose intake exceeds the recommendations

We also note cases in which the PTWI of trace elements is slightly exceeded (except for MeHg ( $34 \%$ ) discussed previously) and in which the TRV of POPs is exceeded.

After characterising the individuals for which the calculated exposure exceeds the TRV for a given contaminant (Table 51), the consumption levels of the main foods contributing to these high exposures were analysed. Table 52 presents the results for the adult population exceeding the TRVs of PCDD/Fs, DL-PCBs and i-PCBs, and for women of child-bearing age exceeding the PTWI of MeHg. In parallel, the consumptions of the main foods contributing to the omega 3 intake of individuals whose EPA and DHA intake reaches the recommendations are presented.

Table 52 : Consumptions (g/week) of the major contributors ( $>5 \%$ exposure) to the exposure to persistent organic pollutants, MeHg and Omega 3 by subjects who have an exposure above the TRV or recommendations

| Major Contributors | Subjects who have an exposure to PCDD/F and PCB-DL > TRV** |  |  | Subjects who have an exposure to iPCB $>$ TRV** |  |  | Women of childbearing age who have an exposure MeHg > TRV** |  |  | Subjects who have an intake of EPA and DHA > Recommendation** |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | \% contrib | Mean $\pm$ SD | P95 | \% contri | Mean $\pm$ SD | P95 | \% contrib | Mean $\pm$ SD | P95 | \% contrib | Mean $\pm$ SD | P95 |
| Eel* | 16 | $9 \pm 30$ | 37 | 12 | $4 \pm 19$ | 19 | - | - | - | - | - | - |
| Seabass* | 7 | $40 \pm 80$ | 175 | 9 | $25 \pm 56$ | 102 | - | - | - | - | - | - |
| Sea bream* | 5 | $38 \pm 84$ | 197 | 5 | $25 \pm 61$ | 114 | - | - | - | - | - | - |
| Swimcrab | 9 | $19 \pm 48$ | 100 | 6 | $8 \pm 32$ | 50 | - | - | - | - | - | - |
| Mackerel | 7 | $18 \pm 105$ | 256 | 7 | $42 \pm 75$ | 173 | - | - | - | 12 | $47 \pm 79$ | 181 |
| Sardine | 23 | $60 \pm 57$ | 164 | 23 | $35 \pm 45$ | 138 | - | - | - | 9 | $39 \pm 47$ | 147 |
| Salmon | 9 | $102 \pm 126$ | 288 | 11 | $73 \pm 95$ | 225 | - | - | - | 23 | $82 \pm 98$ | 230 |
| Cod | - | - | - | - | - | - | 5 | $93 \pm 103$ | 212 | - | - | - |
| Swordfish* | - | - | - | - | - | - | 7 | $13 \pm 38$ | 50 | - | - | - |
| Ling | - | - | - | - | - | - | 5 | $19 \pm 37$ | 98 | - | - | - |
| Whiting | - | - | - | - | - | - | 5 | $36 \pm 62$ | 137 | - | - | - |
| Hake | - | - | - | - | - | - | 5 | $38 \pm 73$ | 190 | - | - | - |
| Sole | - | - | - | - | - | - | 6 | $65 \pm 98$ | 250 | - | - | - |
| Tuna* | - | - | - | - | - | - | 26 | $139 \pm 123$ | 335 | - | - | - |
| Anchovy | - | - | - | - | - | - | - | - | - | 5 | $40 \pm 89$ | 180 |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 <br> ** PTMI for PCDD/F and PCB-DL $=70 \mathrm{pg}$ TEQ $_{\text {oms }} / \mathrm{kg} \mathrm{bw} / \mathrm{month}$ (JECFA, 2001). TDI for $\mathrm{iPCB}=0.02 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} / \mathrm{day}$ (WHO, 2003). PTWI for $\mathrm{MeHg}=1.6 \mu \mathrm{~g} / \mathrm{kg}$ bw/week (JECFA, 2003). Recommended intake for EPA + DHA $=500 \mathrm{mg} / \mathrm{d}$ (ISSFAL, 2004) |  |  |  |  |  |  |  |  |  |  |  |  |

Table 52 and the tables in Appendix 5 listing the contributions to the exposures do not reveal any particular contaminated fish species contributing significantly to the exposure to contaminants covered by this study. Although it is difficult to weigh the nutritional benefits of omega 3 in fish and seafood against the risks associated with their contamination, some points can nevertheless be underlined.

It is useful to recall the correlation between the MeHg content of fish and seafood products (excepting canned and smoked products and prepared dishes) and the n-3 LC-PUFAs content (EPA, DPA, DHA): $r=0.23$ ( $p=0.03$ ). Nevertheless, the major contributors to MeHg exposure in women of child-bearing age (tuna, cod, ling, sole, whiting and hake) are not the main contributors to n-3 LC-PUFAs intake (salmon, mackerel, sardine, anchovy and herring). More particularly, the contribution to n-3 LC-PUFAs intake of the main vectors of MeHg is negligible, representing less than $5 \%$.

Figures 11 and 12 (Chapter 3: Seafood composition and contamination) show clearly the nutritional benefits of salmon, mackerel, sardine, anchovy and halibut, accompanied by low MeHg content.

Nevertheless, we should remember that these species, in particular salmon, sardine and mackerel which are oily fish and therefore rich in omega 3, also contain high levels of POPs and are, regardless of the congeners considered, the main vectors of these pollutants. These same species also contribute to lead, cadmium and organotin exposure, although this is not necessarily due to heavy contamination, but often to their high level of consumption which is visible in the regional data.

From one zone to another we tend to find the same major contributors to contaminant and omega 3 fatty acid intakes. However some geographic differences are seen (Appendix 5).

These disparities are largest as regards the trace elements (other than MeHg ) and it clearly appears that the average contributions in all the zones are not representative of each region. The differences are due to particularly high contaminations in a given zone and/or regional food consumptions differences.

- Contamination : crab accounts for $53 \%$ of the Cd exposure in Lorient, but less than $5 \%$ in the other zones, which is due to heavy contamination of the composite sample in Lorient ( $12 \mu \mathrm{~g} / \mathrm{g}$ versus less than $1 \mu \mathrm{~g} / \mathrm{g}$ in the other sampling zones). Similarly, shrimp accounts for $60 \%$ of Cd exposure in Le Havre, but less than $2 \%$ in the other zones, due to heavy contamination of the composite sample in Le Havre ( $4 \mu \mathrm{~g} / \mathrm{g}$ versus less than $0.05 \mu \mathrm{~g} / \mathrm{g}$ in the other zones). The same observation can be made concerning whelks in La Rochelle ( $21 \%$ of the exposure and a Cd level of $2 \mu \mathrm{~g} / \mathrm{g}$ versus less than 1 $\mu \mathrm{g} / \mathrm{g}$ elsewhere). Ray, a fish relatively contaminated by $\mathrm{As}_{\mathrm{T}}$ and $A s_{\text {inorg }}$, appears in some zones as a majority contributor; similarly for the great scallop in Toulon (14\%). The swimcrab, which appears to be a majority contributor to PCDD/F, DL-PCB and i-PCB exposure in general in the four zones, is in fact a major contributor only in Le Havre, due to the heavy contamination of the composite sample in Le Havre.
- Consumption : the great scallop appears to contribute heavily to Pb exposure in Le Havre (22\%) due to high consumption, as do mussels in La Rochelle (16\%) and sea urchin in Toulon (14\%). The great scallop is also a majority contributor to As inorg exposure in Le Havre (15\%), again due to high consumption, along with cod, less contaminated than ray but widely consumed. The sea urchin in Toulon is also a majority contributor to $\mathrm{As}_{\text {inorg }}$ exposure (12\%) since it is highly consumed compared to the other zones. The eel appears to be the only heavy contributor to PBDE, DL-PCB and i-PCB exposure, but only in La Rochelle due to high consumption and a high consumer rate in all the population groups there, compared to the other zones. Finally, as regards the organotins, in view of the low levels of each compound in the samples, the contributions are explained by the different consumptions from one region to another.
- Contamination and consumption : other large regional contributions to Cd exposure - faithe in Lorient and Toulon, whelks in La Rochelle, anchovy in Lorient and Toulon, great scallop in Le Havre and Toulon - are due not only to slightly higher contamination in these zones, but also to higher consumption of these products. Similarly, faithe, a fish widely consumed, appears to be a main contributor (28\%) to Pb exposure in Lorient in particular since the composite sample is more contaminated in this zone ( $0.2 \mu \mathrm{~g} / \mathrm{g}$ versus less than $0.002 \mu \mathrm{~g} / \mathrm{g}$ elsewhere).

On the basis of our analysis is appears that consumption recommendations should take account of all the data presented here and, if they are based on synthetic results, they should also take into account the fact that consumption levels vary for given species and above that the contamination levels for certain species and certain contaminants can vary greatly from one region to another, even from one sampling point to another within the same region. It is therefore important to incorporate in the analysis the provisioning methods, local ones in particular, of certain products, and encourage consumers to diversify their provisioning origins for local species (and species bought locally but not of local origin) subject to the highest contaminations.

Lastly the analysis of the consumption and exposure data shows that for many subjects exceeding the TRV by the indirect approach, particularly for MeHg , this exceeding is not due to the consumption of highly contaminated products but to their high consumption, in quantities and variety. Low contaminated products consumed in high quantities can lead to exceeding of TRV. For example we can consider an "average fish" including all the 81 different products consumed by subjects exceeding the PTWI for MeHg , weighting the contaminations of those products by the mean consumption of those subjects. The contamination of this "average fish" is $0.096 \mu \mathrm{~g} \mathrm{MeHg} / \mathrm{g}$ fresh weight. A consumption of this fish higher than $1,167 \mathrm{~g}$ per week would lead to an exceeding of the TRV for a subject of 70 kg bw . In other words a high consumer who would consume a important variety of products with about 8 portions of fish and seafood per week or more might present a risk of exceeding the PTWI for MeHg by the indirect approach calculation.

In view of the possible overestimation of consumptions by the data collection method - the food frequency questionnaire - used in this survey, we calculated a correction factor applicable to the consumption data of this survey, using the data collected from the feasibility study ${ }^{86}$.

Having calculated contaminants and Omega 3 contributions of each seafood product, the products for which at least one intake was greater than or equal to $5 \%$ of the total intake were selected. These products and their contribution to the total intake are shown in the Appendix 5.

During the feasibility study, two methods of collecting dietary consumption data had been employed: the food frequency questionnaire and a 7-day diary record. Among the seafood products consumed in the feasibility survey and common to the two methods, we selected the products also common with the list of contributors of the full-scale survey identified previously.

Three categories of products were constituted:

- 1: Fish consumed only fresh or frozen
- 2: Molluscs and crustaceans consumed only fresh or frozen
- 3: Fish and crustaceans consumed fresh, frozen, smoked or canned

For each individual in the database of the pilot survey, the ratios between the quantity of product noted in the 7-day diary record and the quantity declared on the food frequency questionnaire were calculated. Table 53 presents the results of these calculations.

Table 53 : Consumption correction coefficient between the FFQ and the 7 day diary record in the feasibility study for the three categories defined

| Cl 95\% |  |  |  |
| :---: | :---: | :---: | :---: |
| Categorie | Mean | Inf. limit | Sup. limit |
| 1 | 0.47 | 0.20 | 0.73 |
| 2 | 0.58 | 0.31 | 0.85 |
| 3 | 0.65 | 0.43 | 0.86 |
| CI : Confidence interval |  |  |  |

The results show clearly here that the food frequency questionnaire tends to overestimate consumptions, compared to the 7 -day diary record, by a factor 1.5 to 2 depending on the category.

Applying these correction coefficients, the consumptions of all the products concerned for the 996 individuals of the CALIPSO survey were calculated and compared with the consumption data of fish and seafood collected among consumers only in the INCA 99 survey (Appendix 3b).

The difference factor between the consumption values from the two surveys is less than in the first comparison (Part 2.2.) for fish and molluscs and crustaceans where it falls from about 2.5 to 2 . On the other hand, it is reduced much less, from 1.5 to about 1.3, for other seafood products (canned, smoked and others). For the total consumption, the difference factor between the two surveys falls from about 2.5 to 3 .

After correction of the consumptions, it appears that the probability of exceeding the TRVs for all contaminants is less. In particular, before correction the i-PCBs exposure of $72 \%$ of the subjects exceeded the TDI; after correction this figure falls to $58 \%$. Similarly, before correction $39 \%$ of the subjects exceeded the PTMI for PCDD/Fs and DL-PCBs, but only $26 \%$ after correction. Concerning cadmium exposure, the percentage of people exceeding the PTWI drops from $8.5 \%$ to $2.2 \%$, and for arsenic from $0.7 \%$ to $0.03 \%$.

Finally, while $34 \%$ of the subjects in our study have an MeHg exposure exceeding the PTWI of $1.6 \mu \mathrm{~g} / \mathrm{kg}$ $\mathrm{bw} /$ week, after correction this figure falls to $20 \%$. If we consider the two PTWIs ( $1.6 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} / \mathrm{week}$ for women of child-bearing age, $3.3 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} / \mathrm{week}$ for other adult people), the cases exceeding the PTWI fall from $16 \%$ to $7 \%$. Moreover, if we apply to these new exposures the average correction factor of 4.3 (exposure / steady state) we obtain a new average steady state of $0.27 \pm 0.24 \mu \mathrm{~g} / \mathrm{kg} \mathrm{bw} / \mathrm{week}$ (versus $0.35 \pm 0.88 \mu \mathrm{~g} / \mathrm{kg}$ bw/week before correction) and $0.67 \mu \mathrm{~g} / \mathrm{kg}$ bw/week at P95 ( 0.82 before correction).

The correlations between dietary exposure and blood levels are not improved after correction of the consumptions, either for fatty acids or contaminants.

## SEVENTH PART

Unlike traditional so-called "indirect" exposure studies based on ingestions, the CALIPSO study enables finer characterisation of the risks and benefits associated with fish and seafood consumption by measuring actual biological levels. The study reveals that French coastal populations, generally high seafood consumers, are well informed and have sound knowledge of these foods. They appreciate information on this subject which is a source of concern, yet they tend to be ambivalent as regards the public controversy on this issue.

The study shows that the contaminant levels measured in provisioned fish and seafood are globally satisfactory relative to currently applicable regulations, with the exception of a few products. For trace elements this "background" contamination level is relatively homogeneous all along the French coast, whereas for persistent organic pollutants a North-South contamination gradient is observed.

From a "benefits" point of view, the study shows that consuming fish alone at least twice a week (including some oily fish) ensures the recommended intake of omega 3 long-chain polyunsaturated fatty acids. Consequently, in view of the measured data and current scientific knowledge of the benefits of consuming omega 3, notably prevention of cardio-vascular diseases, it is legitimate to conclude that such effects are observed in our study population, even if the study reveals that physiological, nutritional and behavioural factors also affect the homeostasic regulation of omega 3.

As regards risks, only the highest consumers of our study population present a non-negligible probability of exceeding the reference toxicological values, notably for methylmercury, cadmium, dioxins and PCBs ("dioxin-like" or not). For these persistent organic pollutants (POP), other foods not taken into account in this study are also vectors. However, the study of biomarkers and the rectification of the consumptions revealed by the food consumption survey shows that these calculated excess levels are difficult to interpret owing to the uncertainties inherent in all indirect exposure studies and the existence of safety factors. Even when the reference toxicological values are exceeded, the levels remain relatively close to these values, in particular for methylmercury and cadmium. Nevertheless these results demonstrate the need to pursue the efforts being made to reduce exposure (by reducing pollution), especially to dioxins and all PCBs.

With the exception of a few fish, the foods contributing most to omega 3 intake and to exposure to persistent organic pollutants are often the same specifies, in particular salmon, mackerel and sardine, owing mainly to their high fat content and their high consumers rate. For trace elements, the contributing foods are different: for example tuna and swordfish for methylmercury, and shrimp, crab, anchovy, great scallop and periwinkle for cadmium, mainly due to a higher level of contamination and/or consumption in certain regions.

Finally, concerning the global question of weighing health risks against nutritional benefits, the study results confirm the validity of the recommendations formulated by various national scientific bodies: that the general population should consume fish at least twice a week, including some oily fish, and that pregnant or breast-feeding women should consume predator fish not more than once a week.

Looking beyond these general recommendations, this study highlights the advantages of diversifying the consumed fish and seafood species in terms of proportions and provisioning origins in order to ensure a rational balance between benefits and risks compatible with nutritional and toxicological recommendations.

## EIGHTH PART



Appendix 1: Fish and seafood sampled in the 4 study zones

|  | Fish |  | Mollusc, crustacean |
| :---: | :---: | :---: | :---: |
| Le Havre | Angler fish* | Pollack | Crab |
|  | Catshark* | Ray* | Great scallop |
|  | Cod | Saithe / Coalfish | Mussel |
|  | Dab | Salmon | Oyster |
|  | Eel* | Sardine | Periwinkle |
|  | Grenadier / hoki* | Seabass* | Scampi |
|  | Hake | Sea bream* | Shrimp |
|  | Halibut* | Sole | Squid |
|  | Ling | Swordfish* | Swimcrab |
|  | Mackerel | Tuna* | Whelk |
|  | Plaice | Whiting |  |
| Lorient | Angler fish* | Mackerel | Cockle |
|  | Catshark* | Plaice | Crab |
|  | Cod | Pollack | Great scallop |
|  | Dab | Ray* | Mussel |
|  | Emperor* | Saithe / Coalfish | Oyster |
|  | Goatfish | Salmon | Periwinkle |
|  | Grenadier / hoki* | Sardine | Scampi |
|  | Gurnard | Seabass* | Shrimp |
|  | Haddock | Sea bream* | Spider crab |
|  | Hake | Sole | Squid |
|  | Halibut* | Swordfish* | Swimcrab |
|  | John dory | Pout Tuna* |  |
|  | Ling | Whiting |  |
| La Rochelle | Angler fish * | Pollack | Calico scallop |
|  | Catshark* | Ray* | Cockle |
|  | Cod | Saithe / Coalfish | Crab |
|  | Dab | Salmon | Cuttle fish |
|  | Emperor* | Sardine | Great scallop |
|  | Goatfish | Seabass* | Mussel |
|  | Grenadier / hoki* | Sea bream * | Oyster |
|  | Haddock | Sole | Periwinkle |
|  | Hake | Swordfish* | Scampi |
|  | Halibut* | Tuna* | Shrimp |
|  | Ling | Whiting | Squid |
|  | Mackerel |  | Whelk |
| Toulon | Angler fish* | Ray* | Cuttle fish |
|  | Catshark* | Saithe / Coalfish | Great scallop |
|  | Cod | Scorpion fish | Lobster |
|  | Dab | Seabass* | Mussel |
|  | Emperor* | Sea bream * | Octopus |
|  | Goatfish | Salmon | Oyster |
|  | Grenadier / hoki* | Sardine | Sea urchin |
|  | Hake | Sole | Shrimp |
|  | Halibut* | Swordfish* | Squid |
|  | John dory | Tuna* | Whelk |
|  | Ling | Whiting |  |
|  | Mackerel |  |  |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 |  |  |  |

## Samples common to the four zones

| Fresh anchovy | Smoked haddock |
| :--- | :--- |
| Canned anchovy | Smoked herring |
| Preserved anchovy | Smoked mackerel |
| Canned mackerel | Smoked salmon |
| Canned pilchard |  |
| Canned sardines | Tarama |
| Canned Yellow fin tuna* | Surimi |
| Canned Albacore tuna* | Dehydrated fish soup |
| Canned Skipjack tuna* | Liquid fish soup |
| Canned flaked Yellow fin tuna* | Paella |
| Canned flaked tuna (without further details)* |  |
| Canned crab |  |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 |  |

Appendix 2 : Consumptions of fish and seafood by high consumers per zone ( $\mathrm{g} /$ week)
Consumptions of fresh and frozen fish by high consumers - Le Havre ( $\mathbf{g} /$ week)

| Fish | Adult men ( $18-64 \mathrm{y})$$\mathrm{n}=44$n |  |  | Adult women (18-64 y$\mathrm{n}=179$ |  |  | Older subjects (65 y and more) |  |  | Women of childbearing age (18-44 y) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\mathrm{n}=26$ | $\mathrm{n}=98$ |  |  |
|  | Mean | P95 | \%*** |  |  |  | Mean | P95 | \%*** | Mean | P95 | \%*** | Mean | P95 | \%*** |
| Anchovy | 8.4 | 37.5 | 13.6\% | 0.8 | 0.0 | 2.2\% | 1.9 | 18.8 | 7.7\% | 0.5 | 0.0 | 2.0\% |
| Angler fish* | 10.1 | 37.5 | 22.7\% | 11.7 | 50.0 | 36.3\% | 18.9 | 56.3 | 38.5\% | 7.8 | 37.5 | 28.6\% |
| Catshark* | 17.2 | 125.0 | 27.3\% | 20.6 | 125.0 | 38.0\% | 14.6 | 50.0 | 34.6\% | 20.7 | 125.0 | 36.7\% |
| Cod | 73.3 | 220.0 | 81.8\% | 101.4 | 300.0 | 95.5\% | 111.5 | 231.3 | 92.3\% | 77.8 | 200.0 | 95.9\% |
| Dab | 2.3 | 18.8 | 9.1\% | 13.5 | 100.0 | 20.7\% | 8.4 | 37.5 | 23.1\% | 9.9 | 100.0 | 15.3\% |
| Eel* | 0.5 | 0.0 | 23\% | 9.7 | 10.0 | 5.6\% | 0.0 | 0.0 | 0.0\% | 1.7 | 10.0 | 6.1\% |
| Emperor* | 2.1 | 18.8 | 6.8\% | 3.0 | 25.0 | 10.1\% | 1.9 | 25.0 | 7.7\% | 2.0 | 25.0 | 7.1\% |
| Goatish | 1.6 | 12.5 | 11.4\% | 4.3 | 22.5 | 15.1\% | 10.8 | 45.0 | 26.9\% | 2.2 | 12.5 | 8.2\% |
| Grenadier / hoki* | 3.7 | 25.0 | 11.4\% | 14.2 | 100.0 | 25.1\% | 5.0 | 25.0 | 26.9\% | 11.4 | 100.0 | 20.4\% |
| Grouper | 0.0 | 0.0 | 0.0\% | 02 | 0.0 | 1.1\% | 0.0 | 0.0 | 0.0\% | 0.3 | 0.0 | 1.0\% |
| Gurnard | 0.3 | 0.0 | 23\% | 52 | 0.0 | 4.5\% | 0.9 | 0.0 | 3.8\% | 8.0 | 0.0 | 2.0\% |
| Haddock | 4.7 | 36.3 | 13.6\% | 5.5 | 25.0 | 8.4\% | 18.6 | 100.0 | 30.8\% | 2.5 | 0.0 | 2.0\% |
| Hake | 0.5 | 0.0 | 23\% | 12.0 | 93.8 | 15.6\% | 6.8 | 62.5 | 15.4\% | 9.9 | 93.8 | 16.3\% |
| Halibut* | 2.7 | 18.8 | 9.1\% | 19.5 | 100.0 | 24.6\% | 14.6 | 62.5 | 30.8\% | 11.8 | 50.0 | 18.4\% |
| Hering | 15.2 | 62.5 | 38.6\% | 22.5 | 120.0 | 41.3\% | 129.3 | 76.0 | 53.8\% | 15.1 | 100.0 | 29.6\% |
| John Dory | 0.0 | 0.0 | 0.0\% | 2.6 | 22.5 | 12.8\% | 4.6 | 25.0 | 7.7\% | 0.7 | 0.0 | 4.1\% |
| Ling | 13.6 | 125.0 | 20.5\% | 18.9 | 100.0 | 39.1\% | 37.5 | 137.5 | 73.1\% | 9.4 | 47.5 | 31.6\% |
| Mackerel | 25.7 | 109.4 | 61.4\% | 33.7 | 150.0 | 49.7\% | 41.7 | 237.5 | 73.1\% | 41.0 | 237.5 | 46.9\% |
| Mullet | 1.3 | 0.0 | 23\% | 13 | 0.0 | 3.4\% | 0.0 | 0.0 | 0.0\% | 1.7 | 0.0 | 3.1\% |
| Plaice | 19.4 | 100.0 | 27.3\% | 30.3 | 125.0 | 34.6\% | 39.0 | 250.0 | 50.0\% | 22.2 | 100.0 | 23.5\% |
| Pollack | 16.6 | 22.5 | 9.1\% | 11.1 | 50.0 | 212\% | 18.9 | 100.0 | 46.2\% | 6.6 | 45.0 | 16.3\% |
| Pout | 0.4 | 0.0 | 23\% | 0.8 | 0.0 | 4.5\% | 0.0 | 0.0 | 0.0\% | 0.7 | 0.0 | 4.1\% |
| Ray* | 23.9 | 125.0 | 45.5\% | 29.7 | 125.0 | 53.1\% | 46.8 | 125.0 | 57.7\% | 24.0 | 100.0 | 45.9\% |
| Redfish | 0.0 | 0.0 | 0.0\% | 02 | 0.0 | 0.6\% | 0.0 | 0.0 | 0.0\% | 0.4 | 0.0 | 1.0\% |
| Saithe / coalfish | 48.4 | 200.0 | 65.9\% | 49.8 | 200.0 | 67.6\% | 55.5 | 200.0 | 65.4\% | 54.2 | 200.0 | 82.7\% |
| Salmon | 63.8 | 220.0 | 70.5\% | 75.1 | 220.0 | 79.9\% | 72.3 | 220.0 | 88.5\% | 66.4 | 190.0 | 79.6\% |
| Sardine | 11.9 | 62.5 | 25.0\% | 82 | 62.5 | 26.8\% | 11.2 | 62.5 | 34.6\% | 9.1 | 62.5 | 24.5\% |
| Scorpion fish | 0.0 | 0.0 | 0.0\% | 0.8 | 0.0 | 2.8\% | 1.7 | 0.0 | 3.8\% | 1.2 | 0.0 | 4.1\% |
| Seabass* | 19.5 | 120.0 | 38.6\% | 12.0 | 62.5 | 302\% | 16.6 | 118.8 | 42.3\% | 9.1 | 50.0 | 28.6\% |
| Sea bream* | 16.2 | 47.5 | 22.7\% | 8.8 | 47.5 | 20.1\% | 10.5 | 47.5 | 23.1\% | 9.4 | 47.5 | 22.4\% |
| Smelt | 0.4 | 0.0 | 23\% | 23 | 12.5 | 7.3\% | 1.0 | 6.3 | 11.5\% | 2.2 | 12.5 | 6.1\% |
| Sole | 38.7 | 118.8 | 43.2\% | 32.6 | 171.9 | 48.0\% | 33.8 | 150.0 | 42.3\% | 38.6 | 200.0 | 50.0\% |
| Sprat | 0.0 | 0.0 | 0.0\% | 0.5 | 0.0 | 2.8\% | 0.2 | 0.0 | 3.8\% | 0.5 | 0.0 | 2.0\% |
| Swordfish* | 1.7 | 0.0 | 4.5\% | 3.7 | 0.0 | 3.9\% | 21.8 | 55.0 | 15.4\% | 4.2 | 18.8 | 5.1\% |
| Tuna* | 12.5 | 97.5 | 18.2\% | 13.3 | 87.5 | 22.9\% | 11.3 | 48.8 | 23.1\% | 14.4 | 87.5 | 26.5\% |
| Turbot | 3.4 | 50.0 | 6.8\% | 2.5 | 18.8 | 9.5\% | 4.6 | 25.0 | 7.7\% | 1.8 | 12.5 | 7.1\% |
| Whiting | 14.6 | 125.0 | 22.7\% | 24.5 | 118.8 | 44.1\% | 25.9 | 125.0 | 50.0\% | 20.9 | 100.0 | 40.8\% |
| Other** | 8.2 | 62.5 | 13.6\% | 52 | 45.0 | 9.5\% | 9.6 | 112.5 | 7.7\% | 5.5 | 62.5 | 9.2\% |
| TOTAL | 483.0 | 928.8 | 100.0\% | 6121 | 1,629 | 100.0\% | 807.7 | 1,701 | 100.0\% | 525.9 | 1,303 | 100.0\% |

Consumptions of molluscs and crustaceans by high consumers - Le Havre (g/week)

| Mollusc, orustacean | Adult men (18-64 y) |  |  | Adult women (18-64 y) |  |  | Older subjects (65 y and more) |  |  | Women of childbearing age (18-44 y ) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{n}=44$ |  |  | $\mathrm{n}=179$ |  |  | $\mathrm{n}=26$ |  |  | $\mathrm{n}=98$ |  |  |
|  | Mean | P95 | \%* | Mean | P95 | \%* | Mean | P95 | \%* | Mean | P95 | \%* |
| Abalone | 0.0 | 0.0 | 0.0\% | 03 | 0.0 | 1.1\% | 1.3 | 0.0 | 3.8\% | 0.4 | 0.0 | 1.0\% |
| Calico scallop | 6.2 | 56.3 | 13.6\% | 73 | 45.0 | 21.8\% | 11.9 | 56.3 | 26.9\% | 8.0 | 56.3 | 20.4\% |
| Carpet shell | 0.0 | 0.0 | 23\% | 03 | 2.0 | 5.6\% | 0.6 | 6.0 | 7.7\% | 0.4 | 2.5 | 5.1\% |
| Clam | 0.3 | 0.0 | 23\% | 0.1 | 0.0 | 1.1\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Codkle | 1.0 | 6.3 | 13.6\% | 1.6 | 8.8 | 19.6\% | 0.9 | 6.3 | 11.5\% | 1.6 | 12.5 | 20.4\% |
| Crab | 6.6 | 20.0 | 59.1\% | 93 | 50.0 | 59.8\% | 8.0 | 40.0 | 61.5\% | 7.9 | 40.0 | 55.1\% |
| Crayfish | 1.6 | 8.8 | 13.6\% | 2.5 | 12.5 | 20.1\% | 1.3 | 10.0 | 11.5\% | 2.0 | 12.5 | 20.4\% |
| Cuttle fish | 1.9 | 16.3 | 6.8\% | 2.4 | 32.5 | 8.4\% | 5.8 | 16.3 | 11.5\% | 2.4 | 32.5 | 8.2\% |
| Donax clam | 0.0 | 0.0 | 0.0\% | 02 | 0.0 | 0.6\% | 0.5 | 0.0 | 3.8\% | 0.0 | 0.0 | 0.0\% |
| Great scallop | 47.6 | 156.3 | 72.7\% | 57.9 | 250.0 | 83.8\% | 58.2 | 187.5 | 88.5\% | 51.8 | 300.0 | 77.6\% |
| Grooved sea squirt | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Hard clam | 1.0 | 6.3 | 13.6\% | 12 | 7.5 | 19.0\% | 3.1 | 15.0 | 23.1\% | 0.9 | 7.5 | 14.3\% |
| Limpet | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.6\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Lobster | 4.1 | 22.5 | 18.2\% | 5.9 | 22.5 | 17.9\% | 5.2 | 45.0 | 15.4\% | 7.9 | 45.0 | 20.4\% |
| Mussel | 18.4 | 50.0 | 86.4\% | 16.9 | 50.0 | 83.8\% | 18.8 | 43.8 | 88.5\% | 15.4 | 50.0 | 80.6\% |
| Octopus | 0.4 | 0.0 | 4.5\% | 1.9 | 16.3 | 7.3\% | 0.6 | 0.0 | 3.8\% | 1.9 | 10.0 | 7.1\% |
| Oyster | 19.1 | 90.0 | 40.9\% | 17.9 | 90.0 | 51.4\% | 33.0 | 90.0 | 69.2\% | 11.1 | 72.0 | 48.0\% |
| Periwinkle | 4.3 | 12.5 | 50.0\% | 3.7 | 15.0 | 49.7\% | 4.0 | 25.0 | 50.0\% | 2.9 | 12.5 | 45.9\% |
| Queen scallop | 0.3 | 0.0 | 23\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Razor clam | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Scampi | 7.1 | 45.0 | 38.6\% | 62 | 30.0 | 38.5\% | 6.3 | 30.0 | 38.5\% | 4.3 | 30.0 | 30.6\% |
| Sea urchin | 0.0 | 0.0 | 0.0\% | 0.2 | 0.0 | 0.6\% | 0.5 | 0.0 | 3.8\% | 0.4 | 0.0 | 1.0\% |
| Shrimp | 41.5 | 93.8 | 90.9\% | 49.6 | 150.0 | 95.5\% | 43.8 | 150.0 | 88.5\% | 44.7 | 150.0 | 95.9\% |
| Slipper lobster | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Spider crab | 2.3 | 10.0 | 15.9\% | 22 | 10.0 | 18.4\% | 2.4 | 10.0 | 26.9\% | 1.1 | 10.0 | 14.3\% |
| Spiny lobster | 1.8 | 12.5 | 15.9\% | 2.1 | 12.5 | 162\% | 0.7 | 6.3 | 11.5\% | 3.0 | 12.5 | 17.3\% |
| Squid | 19.9 | 81.3 | 52.3\% | 14.3 | 50.0 | 51.4\% | 15.5 | 81.3 | 38.5\% | 12.6 | 50.0 | 50.0\% |
| Swimgrab | 11.2 | 62.5 | 25.0\% | 14.0 | 62.5 | 32.4\% | 33.2 | 187.5 | 53.8\% | 8.3 | 62.5 | 23.5\% |
| Whelk | 24.2 | 62.5 | 52.3\% | 36.3 | 150.0 | 58.1\% | 17.6 | 75.0 | 50.0\% | 27.5 | 130.0 | 51.0\% |
| TOTAL | 220.9 | 526.5 | 100.0\% | 254.3 | 631.3 | 100.0\% | 273.1 | 529.1 | 100.0\% | 216.5 | 588.3 | 100.0\% |
| * Consumers rate |  |  |  |  |  |  |  |  |  |  |  |  |

Consumptions of canned food, smoked fish and seafood-based dishes by high consumers - Le Havre (g/week)

| Other seafood | Adult men (18-64 y) |  |  | Adult women (18-64 y) |  |  | Older subjects (65 y and more) |  |  | Women of childbearing age (18-44 y) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** |
| Canned food |  |  |  |  |  |  |  |  |  |  |  |  |
| Anchovy | 8.5 | 45.0 | 27.3\% | 93 | 37.5 | 22.3\% | 22.1 | 45.0 | 38.5\% | 3.1 | 30.0 | 14.3\% |
| Crab | 17.3 | 56.3 | 68.2\% | 19.2 | 60.0 | 63.1\% | 10.6 | 37.5 | 57.7\% | 22.9 | 180.0 | 65.3\% |
| Madkerel | 2.6 | 13.1 | 11.4\% | 1.6 | 7.5 | 173\% | 2.5 | 18.8 | 15.4\% | 1.3 | 7.5 | 12.2\% |
| Pilchard | 15.1 | 43.8 | 77.3\% | 11.7 | 40.0 | 62.0\% | 13.7 | 40.0 | 84.6\% | 11.8 | 40.0 | 62.2\% |
| Sardine | 75.6 | 180.0 | 97.7\% | 50.4 | 180.0 | 95.5\% | 17.0 | 56.3 | 88.5\% | 59.6 | 315.0 | 95.9\% |
| Tuna* | 2.7 | 22.5 | 20.5\% | 8.1 | 30.0 | 402\% | 4.6 | 22.5 | 42.3\% | 10.6 | 37.5 | 37.8\% |
| Total canned food | 121.9 | 302.5 | 100.0\% | 100.3 | 371.3 | 98.3\% | 70.5 | 171.3 | 96.2\% | 109.4 | 405.0 | 100.0\% |
| Smoked fish |  |  |  |  |  |  |  |  |  |  |  |  |
| Haddock | 1.3 | 11.3 | 15.9\% | 23 | 11.3 | 162\% | 0.7 | 3.8 | 7.7\% | 1.8 | 7.5 | 15.3\% |
| Herring | 7.8 | 56.3 | 36.4\% | 4.3 | 22.5 | 28.5\% | 3.8 | 15.0 | 30.8\% | 4.9 | 22.5 | 28.6\% |
| Mackerel | 1.3 | 11.3 | 11.4\% | 2.7 | 11.3 | 15.1\% | 1.6 | 5.0 | 7.7\% | 4.4 | 15.0 | 19.4\% |
| Salmon | 12.7 | 40.0 | 88.6\% | 13.1 | 40.0 | 872\% | 12.2 | 37.5 | 80.8\% | 13.4 | 40.0 | 86.7\% |
| Total smoked fish | 23.2 | 61.3 | 88.6\% | 22.4 | 72.5 | 89.4\% | 18.3 | 46.3 | 92.3\% | 24.4 | 80.0 | 88.8\% |
| Seafood-based dish |  |  |  |  |  |  |  |  |  |  |  |  |
| Fish soup | 20.2 | 100.0 | 29.5\% | 27.0 | 100.0 | 313\% | 25.7 | 75.0 | 46.2\% | 27.6 | 187.5 | 28.6\% |
| Paella | 2.6 | 12.5 | 25.0\% | 3.9 | 20.0 | 32.4\% | 3.6 | 31.3 | 23.1\% | 5.4 | 25.0 | 37.8\% |
| Surimi | 35.3 | 70.0 | 72.7\% | 49.8 | 210.0 | 87.7\% | 9.0 | 37.5 | 57.7\% | 50.1 | 245.0 | 87.8\% |
| Tarama | 53.5 | 175.0 | 68.2\% | 56.0 | 200.0 | 66.5\% | 18.8 | 125.0 | 26.9\% | 63.4 | 200.0 | 69.4\% |
| Total seafood-based dish | 111.6 | 245.0 | 86.4\% | 136.7 | 395.0 | 96.6\% | 57.2 | 162.5 | 84.6\% | 146.6 | 442.5 | 98.0\% |
| Total | 256.6 | 572.5 | 100.0\% | 259.4 | 765.0 | 100.0\% | 145.9 | 361.3 | 100.0\% | 280.4 | 792.5 | 100.0\% |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 ** Consumers rate |  |  |  |  |  |  |  |  |  |  |  |  |

Consumptions of fresh and frozen fish by high consumers－Lorient（g／week）

|  | me |  |  | It women（18．64 |  |  | OIder subiects（65y yand more） |  |  | Women of dilidbearing age（1844y） |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fsh | Mean | P95 | \％$\%$＊＊ | Mean | P95 | \％＊＊＊＊ | Mean | n＝37 |  |  | p95 |  |
| $w$ | 16.6 | 187.5 | 6\％ |  |  |  |  | 56.3 | 108\％ | 20.4 | 550 | 55\％ |
|  | 2.1 |  | \％ | 23,9 22 | ${ }_{17.5}^{1250}$ |  | $\underset{\substack{24.4 \\ 1.6}}{ }$ | ${ }_{25,5}^{112.5}$ |  | 21.5 1.8 1.8 | 938 0.0 0.0 |  |
|  | 936 996 | 3375 | （30．8\％ | 65.2 160 160 | 2000 <br> 1000 <br> $\substack{200}$ | cinem | 790 376 376 | $\underset{\substack{23,5 \\ 150.0}}{\substack{12,\\}}$ |  | ${ }_{\substack{662 \\ 156 \\ 156}}$ | 2000 <br> 1000 <br> $\substack{20}$ |  |
| Eelt | 4.6 | ${ }_{31,5}^{31,5}$ | 13.50 | 0.7 | 0．0 | ${ }_{44 \%}$ | 1.0 | 0.0 | ${ }^{27 \% \%}$ | 0.2 | 0.0 | 13\％ |
|  | ${ }_{9.5}^{12}$ | 12.5 <br> 35.0 <br> 0.0 | ${ }_{\text {cke }}^{58.5 \%}$ | （11．8 | 250 <br> 450 | 132\％ | 3,8 <br> 388 <br> 18 | cois |  | a， 4.8 4.8 | ${ }_{450}^{0.0}$ |  |
|  | 16.1 0.0 | 12.5 <br> 0.0 <br> 125 |  | ${ }_{0.4}^{24.1}$ | － 10.0 | Stiche | ． | （18．8 | come | $\xrightarrow{22.0}$ | 10.0 <br> 0.0 | 年．1．9\％ |
| derd | 4．0 | 27，5 | －9．9\％ | 6.0 | 23,5 <br>  <br>  <br> 285 | 70， | 207 | ${ }_{12125}^{1215}$ | 21，50\％ |  | 0.0 | 9\％\％ |
|  |  | 2000 $\substack{20.0}$ 200 | 59．6\％ | ${ }_{4}^{475}$ | 1820 <br>  <br> 275 <br> 125 |  | （185 | $\underset{\substack{2000 \\ 250}}{ }$ |  | 38.1 <br> 4.0 | （1500 |  |
| Hering |  |  | 23，19\％ | ${ }_{29}^{28}$ | －1258 |  | 0.0 | ＋120 | 䢒 | ${ }^{12,5}$ | O． |  |
| Ung | － 28.4 | ${ }^{212,50}$ | ${ }^{46,2 \%}$ | 26.0 | ${ }^{1255}$ | ${ }_{\text {cke }}^{3920 \%}$ | 250 | $\underset{1250}{1200}$ | 退 | 22， | ${ }_{1250}^{1250}$ |  |
| Nater | ${ }_{4}^{6.6}$ | 280．5 | 5ism\％ | 412 | 20．0 | \％ | （ | 10．0 | co． | 30．5 0.5 | － | ${ }^{4.35 \%}$ |
|  | ${ }_{17,9}^{29}$ | cois | 58. | ${ }_{29,0}^{43}$ | ${ }_{1250}^{2250}$ | \％ | 8， | 50．0 |  |  | ${ }^{10.0}$ | \％ |
| Paut | 4.4 | ${ }_{23,4}^{238}$ | 11．5\％ | ${ }^{6.5}$ | ${ }^{72,5}$ |  | 3， 3 | ${ }_{1550}$ | ${ }_{8}^{8.19 \%}$ | ${ }^{45}$ | － | 10．5\％ |
|  | 0 | 0.0 |  | O． |  |  |  |  |  |  |  | \％ |
|  | 70.4 <br> 46.5 | ${ }^{200.0}$ | 5， | 51．2 | ${ }_{200.0}^{2000}$ |  | 57， | － |  | 624 | ${ }_{20,8}^{20,0}$ | 67．1\％ |
| $\substack{\text { saraine } \\ \text { scorion }}$ | ${ }^{357}$ | ${ }_{0}^{10,9}$ | ${ }_{7}^{759.9}$ | 22， | 398 <br> 08 | 59．90\％ | 313 | ${ }_{10.0}^{1500}$ | ${ }^{784 \%}$ | 17．0 |  | $\underset{\substack{48.70 \% \\ 1306}}{ }$ |
| Seanest | －240 | ${ }^{950}$ |  | ${ }_{212}^{21.2}$ | cos |  |  | 180.0 | 40．5\％ | ${ }^{17.4}$ | ${ }_{9}^{950}$ | 50．8\％ |
| smett | 1.4 | 18.8 | 77.78 | 02 | ${ }^{0.0}$ | 133\％ | 20 | 250 | 隹 | ${ }_{0}^{0.2}$ | 0， | ${ }^{1,5 \%}$ |
| Sprat | －0．0 | 0．0 | 0．0\％ | 0.1 | 0.0 | 0．6\％ | 0.5 | 0.0 | ${ }^{2727 \%}$ |  | 0．0 | 0．0\％ |
| Sters | ${ }^{624}$ | ${ }^{238,8}$ |  | ${ }^{33}{ }^{43}$ | （1926 | 5859\％ |  | ${ }^{137.5}$ | （2120\％ | ${ }_{3}^{24.6}$ | 1906 | 隹 |
| Whiting | ${ }^{33,7}$ | 17.9 | ${ }^{325 \%}$ | ${ }^{331}$ | 150.0 | 48，7\％ | 34.7 | 125.0 | 5955\％ | 362 | 200.0 | 4887\％ |
| Tot | ${ }_{7} 16.5$ | ${ }^{0.09}$ | come | ${ }_{663}^{29}$ | ${ }^{0.009}$ | 900\％ | ${ }_{7}^{093}$ | ${ }^{2.31}$ | －0．0\％ | ${ }_{\text {coser }}^{0.0}$ | ${ }_{1}^{0.003}$ | 0．0\％ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |

Consumptions of molluscs and crustaceans by high consumers - Lorient (g/week)

| Mollusc, orustacean | Adult men (18-64 y)$n=52$ |  |  | Adult women (18-64 y)$\mathrm{n}=158$ |  |  | Older subjects ( 65 y and more)$\mathrm{n}=37$ |  |  | Women of childbearing age (18-44 y)$\mathrm{n}=76$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Mean | P95 | \%* | Mean | P95 | \%* | Mean | P95 | \%* | Mean | P95 | \%* |
| Abalone | 0.5 | 0.0 | 1.9\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Calico scallop | 7.6 | 45.0 | 21.2\% | 72 | 37.5 | 272\% | 12.7 | 112.5 | 24.3\% | 5.9 | 28.1 | 23.7\% |
| Carpet shell | 10.3 | 37.5 | 59.6\% | 5.4 | 30.0 | 43.7\% | 5.5 | 30.0 | 56.8\% | 4.4 | 30.0 | 32.9\% |
| Clam | 0.1 | 0.0 | 1.9\% | 02 | 0.0 | 3.8\% | 0.0 | 0.0 | 0.0\% | 0.1 | 0.0 | 3.9\% |
| Codkle | 6.7 | 37.5 | 48.1\% | 7.9 | 50.0 | 49.4\% | 6.1 | 31.3 | 40.5\% | 4.9 | 17.5 | 38.2\% |
| Crab | 15.3 | 50.0 | 76.9\% | 10.7 | 50.0 | 70.9\% | 9.2 | 50.0 | 59.5\% | 9.4 | 40.0 | 73.7\% |
| Crayfish | 0.8 | 0.0 | 3.8\% | 03 | 0.0 | 2.5\% | 0.0 | 0.0 | 0.0\% | 0.4 | 0.0 | 2.6\% |
| Cuttle fish | 3.3 | 30.0 | 11.5\% | 1.8 | 0.0 | 3.8\% | 1.8 | 16.3 | 5.4\% | 1.7 | 0.0 | 1.3\% |
| Donax clam | 0.0 | 0.0 | 0.0\% | 0.6 | 0.0 | 1.9\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Great scallop | 22.1 | 62.5 | 78.8\% | 40.6 | 156.3 | 84.8\% | 52.0 | 125.0 | 78.4\% | 33.6 | 125.0 | 84.2\% |
| Grooved sea squirt | 0.0 | 0.0 | 0.0\% | 03 | 0.0 | 0.6\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Hard clam | 1.8 | 12.5 | 23.1\% | 2.1 | 9.4 | 17.7\% | 3.9 | 37.5 | 24.3\% | 1.4 | 9.4 | 13.2\% |
| Limpet | 0.0 | 0.0 | 0.0\% | 1.0 | 0.0 | 3.2\% | 0.0 | 0.0 | 0.0\% | 0.7 | 0.0 | 1.3\% |
| Lobster | 4.5 | 45.0 | 11.5\% | 3.9 | 45.0 | 13.9\% | 2.1 | 22.5 | 10.8\% | 3.3 | 22.5 | 11.8\% |
| Mussel | 23.1 | 50.0 | 96.2\% | 22.8 | 70.0 | 83.5\% | 21.7 | 70.0 | 81.1\% | 21.0 | 70.0 | 76.3\% |
| Octopus | 1.9 | 25.0 | 7.7\% | 1.4 | 0.0 | 1.9\% | 1.4 | 0.0 | 2.7\% | 1.1 | 0.0 | 1.3\% |
| Oyster | 43.9 | 180.0 | 75.0\% | 33.3 | 144.0 | 63.9\% | 57.8 | 180.0 | 83.8\% | 31.3 | 144.0 | 52.6\% |
| Periwinkle | 9.6 | 40.0 | 75.0\% | 8.1 | 40.0 | 68.4\% | 12.0 | 40.0 | 83.8\% | 5.9 | 25.0 | 56.6\% |
| Queen scallop | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.6\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 1.3\% |
| Razor clam | 0.6 | 0.0 | 3.8\% | 0.1 | 0.0 | 0.6\% | 0.0 | 0.0 | 0.0\% | 0.2 | 0.0 | 1.3\% |
| Sea urchin | 0.8 | 0.0 | 1.9\% | 12 | 0.0 | 3.8\% | 0.2 | 0.0 | 2.7\% | 0.1 | 0.0 | 1.3\% |
| Scampi | 43.0 | 120.0 | 84.6\% | 37.9 | 144.0 | 80.4\% | 43.0 | 135.0 | 81.1\% | 27.9 | 75.0 | 76.3\% |
| Shrimp | 38.4 | 100.0 | 94.2\% | 39.8 | 125.0 | 88.0\% | 41.1 | 125.0 | 83.8\% | 45.2 | 150.0 | 88.2\% |
| Slipper lobster | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Spider crab | 11.7 | 50.0 | 75.0\% | 12.4 | 50.0 | 71.5\% | 10.3 | 40.0 | 64.9\% | 7.8 | 40.0 | 63.2\% |
| Spiny lobster | 0.0 | 0.0 | 0.0\% | 0.5 | 6.3 | 8.2\% | 0.1 | 0.0 | 2.7\% | 0.7 | 6.3 | 10.5\% |
| Squid | 25.9 | 125.0 | 55.8\% | 16.4 | 81.3 | 50.0\% | 10.4 | 81.3 | 29.7\% | 17.7 | 125.0 | 46.1\% |
| Swimcrab | 11.9 | 62.5 | 23.1\% | 11.0 | 75.0 | 19.0\% | 3.0 | 37.5 | 8.1\% | 3.6 | 25.0 | 9.2\% |
| Whelk | 4.8 | 37.5 | 15.4\% | 2.6 | 12.5 | 10.1\% | 2.0 | 0.0 | 2.7\% | 1.8 | 12.5 | 6.6\% |
| TOTAL | 288.6 | 700.9 | 100\% | 269.6 | 667.3 | 100.0\% | 296.4 | 623.8 | 100.0\% | 229.9 | 641.0 | 100.0\% |
| *Consumers rate |  |  |  |  |  |  |  |  |  |  |  |  |

Consumptions of canned food, smoked fish and seafood-based dishes by high consumers - Lorient (g/week)

| Other seafood | Adult men (18-64 y) |  |  | Adult women (18-64 y) |  |  | Older subjects (65 y and more) |  |  | Women of childbearing age (18-44 y) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** |
| Canned food |  |  |  |  |  |  |  |  |  |  |  |  |
| Anchovy | 26.4 | 150.0 | 34.6\% | 26.2 | 150.0 | 42.4\% | 11.9 | 75.0 | 27.0\% | 17.1 | 90.0 | 34.2\% |
| Crab | 24.3 | 90.0 | 67.3\% | 11.6 | 56.3 | 55.1\% | 10.2 | 56.3 | 56.8\% | 11.9 | 56.3 | 52.6\% |
| Mackerel | 4.5 | 56.3 | 13.5\% | 1.6 | 7.5 | 8.9\% | 0.5 | 0.0 | 2.7\% | 1.7 | 7.5 | 5.3\% |
| Pilchard | 22.6 | 60.0 | 76.9\% | 13.4 | 50.0 | 64.6\% | 22.1 | 60.0 | 86.5\% | 11.5 | 50.0 | 52.6\% |
| Sardine | 51.9 | 90.0 | 90.4\% | 37.2 | 120.0 | 88.6\% | 34.3 | 105.0 | 81.1\% | 44.1 | 180.0 | 90.8\% |
| Tuna* | 0.9 | 7.5 | 7.7\% | 2.4 | 15.0 | 15.8\% | 2.2 | 18.8 | 21.6\% | 3.0 | 18.8 | 18.4\% |
| Total canned food | 130.6 | 420.0 | 98.1\% | 92.4 | 268.8 | 96.8\% | 81.1 | 201.3 | 94.6\% | 89.2 | 268.8 | 97.4\% |
| Smoked fish |  |  |  |  |  |  |  |  |  |  |  |  |
| Haddock | 1.2 | 7.5 | 7.7\% | 0.7 | 5.6 | 7.0\% | 0.9 | 11.3 | 8.1\% | 0.3 | 0.0 | 2.6\% |
| Herring | 11.4 | 56.3 | 42.3\% | 52 | 22.5 | 31.0\% | 3.4 | 22.5 | 24.3\% | 4.5 | 22.5 | 25.0\% |
| Mackerel | 6.1 | 56.3 | 21.2\% | 2.6 | 22.5 | 11.4\% | 0.7 | 7.5 | 13.5\% | 3.9 | 37.5 | 14.5\% |
| Salmon | 9.2 | 31.3 | 76.9\% | 8.4 | 37.5 | 77.8\% | 5.4 | 25.0 | 54.1\% | 9.3 | 37.5 | 85.5\% |
| Total smoked fish | 27.7 | 93.8 | 88.5\% | 16.9 | 71.3 | 84.8\% | 10.4 | 40.0 | 73.0\% | 18.1 | 70.0 | 89.5\% |
| Seafood-based dish |  |  |  |  |  |  |  |  |  |  |  |  |
| Fish soup | 61.2 | 350.0 | 53.8\% | 60.3 | 300.0 | 46.8\% | 70.3 | 400.0 | 62.2\% | 67.7 | 350.0 | 44.7\% |
| Paella | 5.4 | 31.3 | 44.2\% | 6.9 | 31.3 | 42.4\% | 3.5 | 35.0 | 16.2\% | 8.1 | 31.3 | 44.7\% |
| Surimi | 17.1 | 70.0 | 57.7\% | 20.3 | 70.0 | 69.0\% | 13.3 | 120.0 | 56.8\% | 23.2 | 70.0 | 72.4\% |
| Tarama | 89.1 | 200.0 | 90.4\% | 54.5 | 200.0 | 59.5\% | 40.6 | 125.0 | 43.2\% | 67.7 | 312.5 | 71.1\% |
| Total seafood-based dish | 172.8 | 518.8 | 98.1\% | 1420 | 460.0 | 92.4\% | 127.7 | 545.0 | 83.8\% | 166.7 | 537.5 | 93.4\% |
| Total | 331.2 | 798.8 | 100.0\% | 251.3 | 648.8 | 99.4\% | 219.2 | 622.5 | 100.0\% | 273.9 | 726.3 | 98.7\% |

Consumptions of fresh and frozen fish by high consumers - La Rochelle (g/week)

| Fish | Adult men (18-64 y |  |  | Adult women (18-64 y)$n=122$ |  |  | Older subjects ( 65 y and more)$n=39$ |  |  | Women of childbearing age (18-44 y)$n=78$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean |  |  |  |  |  | Mean | P95 | \%*** |
| Anchovy | 4.2 | 18.8 | 92\% | 13.0 | 22.5 | 9.0\% |  |  |  | 3.6 | 45.0 | 10.3\% | 19.7 | 45.0 | 11.5\% |
| Angler fish* | 13.2 | 50.0 | 34.5\% | 19.0 | 93.8 | 32.8\% | 12.1 | 125.0 | 23.1\% | 15.6 | 68.8 | 29.5\% |
| Catshark* | 10.4 | 62.5 | 23.0\% | 73 | 36.3 | 13.9\% | 14.7 | 100.0 | 25.6\% | 5.5 | 35.0 | 10.3\% |
| cod | 103 | 325.0 | 83.9\% | 94.7 | 245.0 | 86.1\% | 108.4 | 380.0 | 87.2\% | 82.7 | 220.0 | 88.5\% |
| Dab | 9.1 | 50.0 | 18.4\% | 9.5 | 37.5 | 19.7\% | 17.6 | 93.8 | 25.6\% | 9.9 | 45.0 | 21.8\% |
| Eel* | 14.9 | 50.0 | 32.2\% | 63 | 35.0 | 21.3\% | 8.0 | 50.0 | 28.2\% | 5.1 | 35.0 | 19.2\% |
| Emperor* | 6.7 | 18.8 | 10.3\% | 3.4 | 18.8 | 5.7\% | 5.0 | 37.5 | 17.9\% | 0.5 | 0.0 | 2.6\% |
| Goattish | 11.0 | 62.5 | 23.0\% | 9.7 | 45.0 | 24.6\% | 22.3 | 180.0 | 30.8\% | 8.6 | 45.0 | 24.4\% |
| Grenadier/hoki* | 12.4 | 93.8 | 21.8\% | 16.4 | 93.8 | 37.7\% | 41.9 | 380.0 | 41.0\% | 15.0 | 87.5 | 35.9\% |
| Grouper | 0.8 | 0.0 | 2.3\% | 1.6 | 0.0 | 3.3\% | 1.0 | 0.0 | 2.6\% | 0.7 | 0.0 | 2.6\% |
| Gurnard | 8.0 | 25.0 | 11.5\% | 3.9 | 25.0 | 7.4\% | 12.6 | 112.5 | 23.1\% | 3.8 | 37.5 | 9.0\% |
| Haddock | 10.2 | 25.0 | 11.5\% | 73 | 45.0 | 13.9\% | 14.4 | 180.0 | 20.5\% | 6.1 | 45.0 | 11.5\% |
| Hake | 41.6 | 171.9 | 52.9\% | 38.2 | 150.0 | 51.6\% | 49.3 | 180.0 | 71.8\% | 28.4 | 180.0 | 44.9\% |
| Halibut* | 9.5 | 37.5 | 24.1\% | 14.1 | 62.5 | 31.1\% | 18.4 | 300.0 | 7.7\% | 13.1 | 93.8 | 29.5\% |
| Herring | 4.2 | 25.0 | 13.8\% | 6.8 | 25.0 | 16.4\% | 11.6 | 93.8 | 17.9\% | 8.9 | 50.0 | 17.9\% |
| John Dory | 3.5 | 18.8 | 10.3\% | 4.6 | 18.8 | 10.7\% | 3.5 | 37.5 | 12.8\% | 2.7 | 18.8 | 7.7\% |
| Ling | 27.4 | 125.0 | 32.2\% | 17.3 | 68.8 | 34.4\% | 21.1 | 150.0 | 30.8\% | 14.8 | 62.5 | 34.6\% |
| Mackerel | 13.9 | 70.0 | 33.3\% | 14.6 | 95.0 | 28.7\% | 33.8 | 190.0 | 41.0\% | 11.9 | 95.0 | 24.4\% |
| Mullet | 3.7 | 22.5 | 10.3\% | 7.6 | 37.5 | 9.8\% | 3.0 | 37.5 | 10.3\% | 7.1 | 50.0 | 10.3\% |
| Plaice | 9.3 | 25.0 | 10.3\% | 4.4 | 35.0 | 11.5\% | 9.0 | 93.8 | 15.4\% | 4.3 | 37.5 | 11.5\% |
| Pollack | 13.2 | 62.5 | 23.0\% | 73 | 37.5 | 123\% | 15.2 | 100.0 | 28.2\% | 6.9 | 50.0 | 12.8\% |
| Pout | 2.7 | 19.5 | 6.9\% | 1.6 | 0.0 | 3.3\% | 4.4 | 37.5 | 12.8\% | 1.5 | 0.0 | 2.6\% |
| Ray* | 30.2 | 150.0 | 55.2\% | 27.4 | 125.0 | 56.6\% | 56.2 | 200.0 | 64.1\% | 23.9 | 125.0 | 53.8\% |
| Redfish | 1.5 | 0.0 | 2.3\% | 1.1 | 0.0 | 1.6\% | 1.9 | 0.0 | 2.6\% | 1.6 | 0.0 | 2.6\% |
| Saithe / coalfish | 50.4 | 150.0 | 57.5\% | 54.1 | 200.0 | 62.3\% | 48.9 | 190.0 | 48.7\% | 61.9 | 220.0 | 71.8\% |
| Salmon | 60.5 | 190.6 | 60.9\% | 78.8 | 220.0 | 77.9\% | 51.1 | 220.0 | 53.8\% | 91.0 | 220.0 | 83.3\% |
| Sardine | 25.1 | 125.0 | 57.5\% | 30.0 | 150.0 | 50.0\% | 38.2 | 150.0 | 61.5\% | 21.7 | 150.0 | 41.0\% |
| Scorpion fish | 3.0 | 18.8 | 5.7\% | 4.0 | 18.8 | 9.0\% | 3.2 | 25.0 | 12.8\% | 2.5 | 18.8 | 6.4\% |
| Seabass* | 31.9 | 100.0 | 55.2\% | 30.6 | 112.5 | 44.3\% | 21.9 | 125.0 | 41.0\% | 19.3 | 95.0 | 46.2\% |
| Sea bream* | 24.2 | 125.0 | 33.3\% | 16.3 | 90.6 | 32.0\% | 30.3 | 275.0 | 43.6\% | 7.2 | 37.5 | 24.4\% |
| Smelt | 5.7 | 50.0 | 11.5\% | 3.8 | 21.9 | 12.3\% | 2.4 | 37.5 | 7.7\% | 5.3 | 25.0 | 17.9\% |
| Sole | 55.6 | 200.0 | 65.5\% | 47.7 | 171.9 | 623\% | 65.6 | 275.0 | 71.8\% | 33.6 | 150.0 | 55.1\% |
| Sprat | 1.1 | 0.0 | 1.1\% | 02 | 0.0 | 0.8\% | 1.6 | 25.0 | 5.1\% | 0.2 | 0.0 | 1.3\% |
| Swordfish* | 6.5 | 27.5 | 13.8\% | 52 | 37.5 | 11.5\% | 3.1 | 27.5 | 12.8\% | 4.0 | 38.1 | 11.5\% |
| Tuna* | 24.7 | 137.5 | 40.2\% | 27.5 | 112.5 | 45.9\% | 39.2 | 137.5 | 56.4\% | 26.7 | 137.5 | 43.6\% |
| Turbot | 6.2 | 37.5 | 8.0\% | 1.9 | 18.8 | 6.6\% | 5.8 | 50.0 | 12.8\% | 1.6 | 18.1 | 5.1\% |
| Whiting | 23.0 | 125.0 | 33.3\% | 14.4 | 62.5 | 262\% | 17.9 | 112.5 | 28.2\% | 11.9 | 62.5 | 25.6\% |
| Other** | 1.3 | 0.0 | 23\% | 2.9 | 18.8 | 8.2\% | 1.7 | 27.5 | 5.1\% | 2.9 | 27.5 | 10.3\% |
| TOTAL | 684.1 | 1,580 | 98.9\% | 654.4 | 1,410 | 100.0\% | 819.6 | 2,009 | 100.0\% | 588.2 | 1,303 | 100.0\% |

Consumptions of molluscs and crustaceans by high consumers - La Rochelle (g/week)

| Mollusc, orustacean | Adult men (18-64 y) |  |  | Adult women (18-64 y) |  |  | Older subjects (65 y and more) |  |  | Women of childbearing age (18-44 y) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{n}=87$ |  |  | $\mathrm{n}=122$ |  |  | $\mathrm{n}=39$ |  |  | $\mathrm{n}=78$ |  |  |
|  | Mean | P95 | \%* | Mean | P95 | \%* | Mean | P95 | \%* | Mean | P95 | \%* |
| Abalone | 0.4 | 0.0 | 1.1\% | 0.1 | 0.0 | 0.8\% | 4.5 | 0.0 | 2.6\% | 0.2 | 0.0 | 1.3\% |
| Calico scallop | 30.4 | 75.0 | 32.2\% | 32.7 | 150.0 | 39.3\% | 50.4 | 300.0 | 61.5\% | 40.5 | 281.3 | 37.2\% |
| Carpet shell | 3.4 | 25.0 | 32.2\% | 2.8 | 15.0 | 24.6\% | 3.1 | 12.0 | 41.0\% | 3.0 | 15.0 | 26.9\% |
| Clam | 0.1 | 0.0 | 4.6\% | 0.2 | 0.0 | 4.1\% | 0.4 | 0.0 | 2.6\% | 0.3 | 2.0 | 6.4\% |
| Codkle | 1.9 | 8.8 | 24.1\% | 3.4 | 17.5 | 262\% | 2.7 | 30.0 | 15.4\% | 3.9 | 18.8 | 30.8\% |
| Crab | 9.1 | 25.0 | 67.8\% | 92 | 25.0 | 63.9\% | 10.1 | 50.0 | 53.8\% | 8.9 | 25.0 | 65.4\% |
| Crayfish | 1.6 | 10.0 | 13.8\% | 2.0 | 15.0 | 123\% | 0.2 | 0.0 | 2.6\% | 2.9 | 31.3 | 16.7\% |
| Cuttle fish | 16.8 | 81.3 | 49.4\% | 19.0 | 81.3 | 46.7\% | 11.3 | 81.3 | 35.9\% | 18.0 | 81.3 | 50.0\% |
| Donax clam | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Great scallop | 33.5 | 100.0 | 66.7\% | 31.2 | 100.0 | 63.9\% | 27.7 | 156.3 | 56.4\% | 29.6 | 125.0 | 62.8\% |
| Grooved sea squirt | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% |
| Hard clam | 1.2 | 9.4 | 12.6\% | 0.5 | 0.0 | 4.9\% | 1.3 | 12.5 | 15.4\% | 0.2 | 0.0 | 3.8\% |
| Limpet | 0.0 | 0.0 | 0.0\% | 0.3 | 0.0 | 1.6\% | 1.3 | 0.0 | 2.6\% | 0.4 | 0.0 | 2.6\% |
| Lobster | 2.8 | 22.5 | 12.6\% | 5.1 | 22.5 | 19.7\% | 3.8 | 22.5 | 12.8\% | 4.6 | 22.5 | 17.9\% |
| Mussel | 26.5 | 70.0 | 80.5\% | 32.3 | 70.0 | 87.7\% | 36.6 | 70.0 | 100.0\% | 32.8 | 70.0 | 91.0\% |
| Octopus | 1.6 | 16.3 | 8.0\% | 0.6 | 0.0 | 2.5\% | 1.0 | 20.0 | 5.1\% | 1.0 | 0.0 | 3.8\% |
| Oyster | 66.8 | 216.0 | 81.6\% | 57.8 | 180.0 | 77.9\% | 80.5 | 288.0 | 79.5\% | 49.0 | 180.0 | 73.1\% |
| Queen scallop | 2.6 | 0.0 | 3.4\% | 0.0 | 0.0 | 0.0\% | 0.7 | 9.4 | 5.1\% | 0.0 | 0.0 | 0.0\% |
| Periwinkle | 2.2 | 10.0 | 46.0\% | 42 | 15.0 | 48.4\% | 1.9 | 12.5 | 25.6\% | 5.6 | 31.3 | 51.3\% |
| Razor clam | 0.5 | 0.0 | 3.4\% | 1.6 | 0.0 | 4.9\% | 0.0 | 0.0 | 0.0\% | 2.4 | 25.0 | 6.4\% |
| Sea urchin | 0.6 | 0.0 | 3.4\% | 0.0 | 0.0 | 0.0\% | 1.5 | 13.1 | 5.1\% | 0.0 | 0.0 | 0.0\% |
| Scampi | 22.0 | 90.0 | 69.0\% | 26.5 | 112.5 | 73.8\% | 34.4 | 180.0 | 69.2\% | 26.6 | 135.0 | 71.8\% |
| Shrimp | 34.0 | 75.0 | 93.1\% | 43.3 | 125.0 | 91.8\% | 34.8 | 100.0 | 76.9\% | 47.1 | 150.0 | 91.0\% |
| Spiny lobster | 1.2 | 6.3 | 13.8\% | 0.3 | 0.0 | 4.9\% | 2.4 | 12.5 | 23.1\% | 0.4 | 6.3 | 6.4\% |
| Slipper lobster | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 0.0\% | 0.0 | 0.0 | 2.6\% | 0.0 | 0.0 | 0.0\% |
| Spider crab | 4.0 | 20.0 | 25.3\% | 2.0 | 10.0 | 18.0\% | 1.3 | 10.0 | 20.5\% | 1.8 | 10.0 | 17.9\% |
| Squid | 17.7 | 81.3 | 50.6\% | 14.0 | 50.0 | 41.8\% | 11.4 | 50.0 | 41.0\% | 13.0 | 50.0 | 37.2\% |
| Swimcrab | 7.4 | 50.0 | 14.9\% | 7.9 | 12.5 | 8.2\% | 0.8 | 12.5 | 5.1\% | 4.3 | 18.8 | 7.7\% |
| Whelk | 17.3 | 100.0 | 50.6\% | 22.8 | 125.0 | 46.7\% | 6.7 | 37.5 | 23.1\% | 27.3 | 150.0 | 50.0\% |
| TOTAL | 305.7 | 730.3 | 99.1\% | 319.9 | 974.3 | 992\% | 330.6 | 721.3 | 100.0\% | 323.9 | 1,163 | 98.7\% |
| * Consumers rate |  |  |  |  |  |  |  |  |  |  |  |  |

Consumptions of canned food, smoked fish and seafood-based dishes by high consumers - La Rochelle (g/week)

| Other seafood | Adult men (18-64 y) |  |  | Adult women (18-64 y) |  |  | Older subjects (65 y and more) |  |  | Women of childbearing age (18-44 y) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** |
| Canned food |  |  |  |  |  |  |  |  |  |  |  |  |
| Anchovy | 19.8 | 112.5 | 33.3\% | 12.8 | 60.0 | 27.0\% | 12.6 | 150.0 | 28.2\% | 9.3 | 60.0 | 26.9\% |
| Crab | 17.5 | 60.0 | 62.1\% | 13.7 | 56.3 | 54.9\% | 11.8 | 60.0 | 53.8\% | 14.0 | 56.3 | 59.0\% |
| Mackerel | 2.6 | 22.5 | 10.3\% | 1.6 | 15.0 | 9.0\% | 0.6 | 7.5 | 7.7\% | 1.1 | 7.5 | 7.7\% |
| Pilchard | 11.5 | 40.0 | 60.9\% | 12.3 | 40.0 | 69.7\% | 13.8 | 60.0 | 71.8\% | 10.9 | 40.0 | 61.5\% |
| Sardine | 30.8 | 90.0 | 87.4\% | 29.5 | 90.0 | 88.5\% | 14.3 | 56.3 | 74.4\% | 31.9 | 90.0 | 91.0\% |
| Tuna* | 4.8 | 22.5 | 24.1\% | 2.8 | 15.0 | 22.1\% | 1.1 | 9.4 | 17.9\% | 3.2 | 15.0 | 19.2\% |
| Total canned food | 87.0 | 240.6 | 97.7\% | 72.7 | 168.8 | 98.4\% | 54.1 | 182.5 | 89.7\% | 70.4 | 172.5 | 97.4\% |
| Smoked fish |  |  |  |  |  |  |  |  |  |  |  |  |
| Haddock | 1.0 | 7.5 | 10.3\% | 2.5 | 11.3 | 14.8\% | 0.2 | 0.0 | 2.6\% | 2.3 | 11.3 | 17.9\% |
| Herring | 6.7 | 28.1 | 41.4\% | 3.9 | 18.8 | 27.9\% | 7.1 | 45.0 | 46.2\% | 4.4 | 22.5 | 33.3\% |
| Mackerel | 2.5 | 15.0 | 19.5\% | 1.7 | 12.5 | 13.1\% | 0.9 | 7.5 | 5.1\% | 1.7 | 11.3 | 15.4\% |
| Salmon | 9.0 | 31.3 | 65.5\% | 7.9 | 30.0 | 69.7\% | 4.5 | 20.0 | 59.0\% | 8.2 | 37.5 | 66.7\% |
| Total smoked fish | 19.2 | 58.1 | 79.3\% | 16.1 | 52.5 | 76.2\% | 12.7 | 56.3 | 74.4\% | 16.6 | 52.5 | 76.9\% |
| Seafood-based dish |  |  |  |  |  |  |  |  |  |  |  |  |
| Fish soup | 77.4 | 250.0 | 66.7\% | 73.6 | 250.0 | 61.5\% | 64.6 | 312.5 | 71.8\% | 86.1 | 400.0 | 60.3\% |
| Paella | 3.6 | 18.8 | 31.0\% | 5.0 | 21.9 | 33.6\% | 0.9 | 8.8 | 17.9\% | 6.6 | 40.0 | 37.2\% |
| Surimi | 25.9 | 80.0 | 71.3\% | 23.3 | 70.0 | 68.9\% | 7.9 | 40.0 | 33.3\% | 27.3 | 70.0 | 76.9\% |
| Tarama | 54.9 | 175.0 | 50.6\% | 33.6 | 125.0 | 41.0\% | 14.6 | 80.0 | 25.6\% | 38.7 | 175.0 | 44.9\% |
| Total seafood-based dish | 161.8 | 475.0 | 87.4\% | 135.4 | 368.8 | 91.0\% | 88.0 | 312.5 | 82.1\% | 158.7 | 462.5 | 89.7\% |
| Total | 268.1 | 736.3 | 98.9\% | 224.1 | 574.4 | 98.4\% | 154.8 | 476.3 | 97.4\% | 245.6 | 607.5 | 97.4\% |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 ** Consumers rate |  |  |  |  |  |  |  |  |  |  |  |  |

Consumptions of fresh and frozen fish by high consumers - Toulon (g/week)

| Fish |  |  |  | Adult women (18-64$\mathrm{n}=171$ |  |  | Older subjects (65 y and more) |  |  | Women of childbearing age (18-44 y ) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\mathrm{n}=21$ |  |  | $\mathrm{n}=92$ |  |
|  | Mean | P95 | \%*** |  |  |  | Mean | P95 | \%*** | Mean | P95 | \%*** | Mean | P95 | \%*** |
| Anchovy | 4.4 | 37.5 | 11.7\% | 32 | 22.5 | 11.1\% | 2.9 | 15.0 | 9.5\% | 1.6 | 15.0 | 7.6\% |
| Angler fish* | 13.3 | 68.8 | 38.3\% | 14.5 | 50.0 | 43.9\% | 35.7 | 93.8 | 42.9\% | 9.2 | 47.5 | 32.6\% |
| Catshark* | 2.1 | 17.5 | 83\% | 8.6 | 37.5 | 21.6\% | 4.5 | 25.0 | 19.0\% | 10.2 | 50.0 | 21.7\% |
| Cod | 90.9 | 325.0 | 78.3\% | 109.9 | 300.0 | 86.0\% | 69.6 | 190.0 | 71.4\% | 115.2 | 275.0 | 87.0\% |
| Dab | 14.5 | 92.2 | 26.7\% | 17.6 | 100.0 | 26.3\% | 28.0 | 145.0 | 19.0\% | 15.5 | 100.0 | 18.5\% |
| Eel* | 0.9 | 0.0 | 1.7\% | 1.9 | 12.5 | 5.3\% | 1.2 | 0.0 | 4.8\% | 2.7 | 25.0 | 6.5\% |
| Emperor* | 2.7 | 25.0 | 83\% | 3.5 | 18.1 | 7.6\% | 6.0 | 0.0 | 4.8\% | 1.3 | 0.0 | 4.3\% |
| Goattish | 11.7 | 62.8 | 28.3\% | 13.2 | 62.5 | 45.0\% | 22.3 | 90.0 | 52.4\% | 11.0 | 62.5 | 39.1\% |
| Grenadier / hoki* | 1 | 6.3 | 5.0\% | 5.8 | 37.5 | 12.9\% | 0.0 | 0.0 | 0.0\% | 6.0 | 50.0 | 9.8\% |
| Grouper | 1.5 | 18.4 | 6.7\% | 4.7 | 35.0 | 11.1\% | 1.7 | 12.5 | 9.5\% | 3.6 | 31.3 | 9.8\% |
| Gurnard | 1.4 | 18.1 | 6.7\% | 3.5 | 18.8 | 8.8\% | 1.1 | 0.0 | 4.8\% | 0.9 | 0.0 | 3.3\% |
| Haddock | 8.8 | 37.5 | 10.0\% | 2.8 | 12.5 | 7.0\% | 0.6 | 0.0 | 4.8\% | 1.6 | 6.3 | 5.4\% |
| Hake | 3.7 | 12.5 | 5.0\% | 12.3 | 93.8 | 19.9\% | 0.0 | 0.0 | 0.0\% | 13.7 | 100.0 | 21.7\% |
| Halibut* | 10.7 | 50.0 | 23.3\% | 11.5 | 50.0 | 23.4\% | 12.2 | 50.0 | 28.6\% | 11.2 | 50.0 | 21.7\% |
| Herring | 1.8 | 9.4 | 6.7\% | 2.0 | 12.5 | 7.0\% | 7.1 | 0.0 | 4.8\% | 3.2 | 25.0 | 9.8\% |
| John Dory | 4.3 | 24.4 | 18.3\% | 6.4 | 37.5 | 21.6\% | 12.0 | 68.8 | 33.3\% | 3.2 | 25.0 | 14.1\% |
| Ling | 6.8 | 25.0 | 23.3\% | 11.2 | 93.8 | 20.5\% | 22.3 | 125.0 | 23.8\% | 8.6 | 50.0 | 18.5\% |
| Mackerel | 11.8 | 47.5 | 20.0\% | 8.7 | 62.5 | 19.9\% | 6.3 | 25.0 | 14.3\% | 7.8 | 45.0 | 15.2\% |
| Mullet | 1.3 | 3.1 | 5.0\% | 9.4 | 45.0 | 8.8\% | 0.0 | 0.0 | 0.0\% | 5.6 | 18.8 | 5.4\% |
| Plaice | 2.9 | 25.0 | 10.0\% | 3.6 | 25.0 | 8.8\% | 14.3 | 100.0 | 9.5\% | 2.9 | 18.8 | 6.5\% |
| Pollack | 1.7 | 12.5 | 5.0\% | 83 | 45.0 | 12.3\% | 0.6 | 0.0 | 4.8\% | 8.1 | 37.5 | 13.0\% |
| Pout | 1.0 | 0.0 | 3.3\% | 02 | 0.0 | 1.2\% | 0.0 | 0.0 | 0.0\% | 0.1 | 0.0 | 1.1\% |
| Ray* | 9.4 | 50.0 | 20.0\% | 14.9 | 50.0 | 27.5\% | 8.8 | 47.5 | 19.0\% | 10.7 | 50.0 | 22.8\% |
| Redfish | 0.0 | 0.0 | 0.0\% | 22 | 0.0 | 3.5\% | 0.0 | 0.0 | 0.0\% | 0.1 | 0.0 | 1.1\% |
| Saithe / coalfish | 48.2 | 162.5 | 56.7\% | 70.0 | 200.0 | 71.9\% | 40.5 | 145.0 | 42.9\% | 81.8 | 200.0 | 79.3\% |
| Salmon | 50.5 | 210.0 | 60.0\% | 65.9 | 220.0 | 69.0\% | 45.8 | 220.0 | 47.6\% | 64.4 | 220.0 | 68.5\% |
| Sardine | 24.0 | 101.6 | 53.3\% | 21.2 | 100.0 | 43.9\% | 16.7 | 37.5 | 38.1\% | 11.4 | 78.1 | 29.3\% |
| Scorpion fish | 8.0 | 52.5 | 25.0\% | 73 | 45.0 | 292\% | 24.2 | 125.0 | 38.1\% | 4.6 | 25.0 | 23.9\% |
| Seabas** | 38.9 | 125.0 | 61.7\% | 31.8 | 112.5 | 61.4\% | 46.5 | 95.0 | 66.7\% | 21.7 | 95.0 | 53.3\% |
| Sea bream* | 41.1 | 160.0 | 53.3\% | 42.4 | 180.0 | 62.0\% | 81.5 | 400.0 | 61.9\% | 30.8 | 150.0 | 54.3\% |
| Smelt | 1.3 | 0.0 | 33\% | 2.5 | 22.5 | 8.2\% | 0.0 | 0.0 | 0.0\% | 1.5 | 12.5 | 5.4\% |
| Sole | 51.2 | 250.0 | 61.7\% | 30.7 | 171.9 | 48.0\% | 118.6 | 400.0 | 61.9\% | 34.7 | 171.9 | 45.7\% |
| Sprat | 0.2 | 0.0 | 1.7\% | 0. | 0.0 | 1.2\% | 0.0 | 0.0 | 0.0\% | 0.1 | 0.0 | 1.1\% |
| Swordfish* | 14.7 | 49.4 | 26.7\% | 93 | 48.8 | 19.9\% | 4.7 | 22.5 | 23.8\% | 8.6 | 48.8 | 17.4\% |
| Tuna* | 32.9 | 144.1 | 51.7\% | 29.7 | 137.5 | 48.0\% | 23.0 | 55.0 | 42.9\% | 23.2 | 137.5 | 35.9\% |
| Turbot | 3.3 | 12.5 | 5.0\% | 2.4 | 22.5 | 9.9\% | 1.1 | 0.0 | 4.8\% | 2.1 | 22.5 | 8.7\% |
| Whiting | 28.8 | 125.0 | 48.3\% | 29.8 | 125.0 | 49.7\% | 34.9 | 125.0 | 33.3\% | 26.9 | 125.0 | 48.9\% |
| Other** | 1.9 | 13.8 | 5.0\% | 1.6 | 12.5 | 6.4\% | 0.0 | 0.0 | 0.0\% | 1.2 | 9.4 | 5.4\% |
| * Predatory fish as described in the Commission Regulation (EC) No $78 / 2005$ of 19 January 2005 ** Other : Perch, bogue, trout, scabbard-fish, sea wrasse, conger eel, needlefish, bream coriphene, ling, bassbu *** Comsumers rate |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Consumptions of molluscs and crustaceans by high consumers - Toulon (g/week)


Consumptions of canned food, smoked fish and seafood-based dishes by high consumers - Toulon (g/week)

| Other seafood | Adult men (18-64 y) |  |  | Adult women (18-64 y) |  |  | Older subjects (65 y and more) |  |  | Women of childbearing age (18-44 y) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** | Mean | P95 | \%** |
| Canned food |  |  |  |  |  |  |  |  |  |  |  |  |
| Anchovy | 70.4 | 262.5 | 60.0\% | 41.0 | 187.5 | 51.5\% | 42.9 | 93.8 | 61.9\% | 37.9 | 187.5 | 41.3\% |
| Crab | 12.1 | 77.8 | 41.7\% | 11.3 | 56.3 | 46.2\% | 5.7 | 18.8 | 47.6\% | 14.8 | 60.0 | 45.7\% |
| Mackerel | 1.6 | 7.5 | 6.7\% | 0.7 | 0.0 | 4.7\% | 0.9 | 0.0 | 4.8\% | 1.2 | 7.5 | 5.4\% |
| Pilchard | 16.3 | 65.0 | 66.7\% | 10.8 | 40.0 | 66.1\% | 21.2 | 60.0 | 85.7\% | 11.2 | 40.0 | 62.0\% |
| Sardine | 66.9 | 315.0 | 95.0\% | 57.8 | 210.0 | 92.4\% | 19.4 | 60.0 | 71.4\% | 80.0 | 315.0 | 95.7\% |
| Tuna* | 9.3 | 56.3 | 46.7\% | 11.1 | 56.3 | 49.7\% | 3.0 | 9.4 | 38.1\% | 15.7 | 90.0 | 47.8\% |
| Total canned food | 176.6 | 532.5 | 100.0\% | 1327 | 360.0 | 98.2\% | 93.1 | 167.5 | 95.2\% | 160.8 | 667.5 | 100.0\% |
| Smoked fish |  |  |  |  |  |  |  |  |  |  |  |  |
| Haddock | 0.5 | 5.6 | 8.3\% | 1.6 | 7.5 | 12.3\% | 0.4 | 0.0 | 4.8\% | 0.8 | 7.5 | 7.6\% |
| Herring | 6.8 | 51.6 | 30.0\% | 5.1 | 22.5 | 322\% | 2.1 | 15.0 | 23.8\% | 3.7 | 22.5 | 25.0\% |
| Mackerel | 2.6 | 22.5 | 10.0\% | 1.2 | 7.5 | 11.7\% | 0.4 | 0.0 | 4.8\% | 1.3 | 7.5 | 13.0\% |
| Salmon | 10.8 | 37.5 | 91.7\% | 10.8 | 40.0 | 86.5\% | 8.6 | 25.0 | 66.7\% | 9.5 | 40.0 | 83.7\% |
| Total smoked fish | 20.7 | 85.0 | 95.0\% | 18.7 | 70.0 | 89.5\% | 11.5 | 28.8 | 85.7\% | 15.3 | 56.3 | 87.0\% |
| Seafood-based dish |  |  |  |  |  |  |  |  |  |  |  |  |
| Fish soup | 84.2 | 250.0 | 68.3\% | 73.2 | 250.0 | 68.4\% | 82.1 | 250.0 | 90.5\% | 74.3 | 400.0 | 62.0\% |
| Paella | 8.5 | 50.0 | 43.3\% | 11.2 | 50.0 | 52.0\% | 3.1 | 2.5 | 9.5\% | 14.3 | 100.0 | 45.7\% |
| Surimi | 28.1 | 70.0 | 80.0\% | 44.0 | 140.0 | 78.9\% | 18.9 | 43.8 | 38.1\% | 57.0 | 210.0 | 85.9\% |
| Tarama | 83.0 | 312.5 | 83.3\% | 59.6 | 200.0 | 74.9\% | 36.2 | 125.0 | 52.4\% | 69.9 | 312.5 | 78.3\% |
| Total seafood-based dish | 203.7 | 407.5 | 98.3\% | 188.0 | 540.0 | 98.8\% | 140.3 | 275.0 | 95.2\% | 215.5 | 606.3 | 98.9\% |
| Total | 401.0 | 987.2 | 100.0\% | 339.4 | 843.8 | 100.0\% | 244.9 | 472.5 | 100.0\% | 391.6 | 937.5 | 100.0\% |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 ** Consumers rate |  |  |  |  |  |  |  |  |  |  |  |  |

## Appendix 3a:

Comparison between consumption data from Calipso and INCA before correction

| Seafood group | $\begin{gathered} \text { Calipso }(n=243) \\ \text { Adult men }(18-64 y) \end{gathered}$ |  | INCA ( $\mathrm{n}=509$ )Consumers only ( $83,5 \%$ ) |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Mean | P95 | Mean | P95 |
| Fish | 633.0 | 1,491 | 267.2 | 705.0 |
| Mollusc, crustacean | 270.3 | 703.4 | 115.1 | 285.0 |
| Seafood-based dishes | 312.3 | 798.8 | 223.2 | 730.0 |
| Total | 1,216 | 2,486 | 365.9 | 1,075 |
|  | $\begin{gathered} \text { Calipso }(\mathrm{n}=630) \\ \text { Adult women }(18-64 \mathrm{y}) \end{gathered}$ |  | INCA ( $\mathrm{n}=610$ )Consumers only ( $84.3 \%$ ) |  |
| Fish | 636.5 | 1,522 | 229.0 | 580.0 |
| Mollusc, crustacean | 259.9 | 665.3 | 109.8 | 258.0 |
| Seafood-based dishes | 272.2 | 742.5 | 168.1 | 640.0 |
| Total | 1,169 | 2,588 | 304.5 | 800.0 |
|  | $\begin{gathered} \text { Calipso ( } n=123 \text { ) } \\ \text { Older subjects ( } 65 \mathrm{y} \text { and more) } \end{gathered}$ |  | INCA ( $\mathrm{n}=243$ )Consumers only ( $86.1 \%$ ) |  |
| Fish | 787.8 | 1,783 | 290.2 | 695.0 |
| Mollusc, crustacean | 279.3 | 648.8 | 115.3 | 330.0 |
| Seafood-based dishes | 187.7 | 472.5 | 127.0 | 410.0 |
| Total | 1,255 | 2,764 | 332.5 | 850.5 |
|  | Calipso ( $\mathrm{n}=344$ ) <br> Women of childbearing age (18-44 y) |  | INCA ( $\mathrm{n}=404$ )Consumers only ( $82.7 \%$ ) |  |
| Fish | 569.4 | 1286.9 | 224.3 | 580.0 |
| Mollusc, crustacean | 235.1 | 607.4 | 109.3 | 250.0 |
| Seafood-based dishes | 300.8 | 795.0 | 174.4 | 650.0 |
| Total | 1,105 | 2,401 | 300.3 | 785.0 |

## Appendix 3b:

Comparison between consumption data from Calipso and INCA after correction

| Seafood group | $\begin{gathered} \text { Calipso }(\mathrm{n}=243) \\ \text { Adult men }(18-64 \mathrm{y}) \end{gathered}$ |  | INCA ( $\mathrm{n}=509$ )Consumers only ( $83,5 \%$ ) |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Mean | P95 | Mean | P95 |
| Fish | 441.8 | 1,076 | 267.2 | 705.0 |
| Mollusc, crustacean | 205.7 | 511.9 | 115.1 | 285.0 |
| Seafood-based dishes | 268.4 | 695.0 | 223.2 | 730.0 |
| Total | 915.8 | 1,967 | 365.9 | 1,075 |
|  | $\begin{gathered} \text { Calipso }(\mathrm{n}=630) \\ \text { Adult women }(18-64 \mathrm{y}) \end{gathered}$ |  | $\begin{gathered} \text { INCA }(n=610) \\ \text { Consumers only (84.3\%) } \end{gathered}$ |  |
| Fish | 451.8 | 1,155 | 229.0 | 580.0 |
| Mollusc, crustacean | 194.7 | 544.7 | 109.8 | 258.0 |
| Seafood-based dishes | 236.1 | 641.0 | 168.1 | 640.0 |
| Total | 882.6 | 1,943 | 304.5 | 800.0 |
|  | $\begin{aligned} & \text { Calipso ( } n=123 \text { ) } \\ & \text { Older subjects ( } 65 \mathrm{y} \text { and more) } \end{aligned}$ |  | $\begin{gathered} \text { INCA ( } \mathrm{n}=243 \text { ) } \\ \text { Consumers only ( } 86.1 \% \text { ) } \end{gathered}$ |  |
| Fish | 575.6 | 1,413 | 290.2 | 695.0 |
| Mollusc, crustacean | 208.5 | 535.6 | 115.3 | 330.0 |
| Seafood-based dishes | 161.9 | 396.9 | 127.0 | 410.0 |
| Total | 946.0 | 2,148 | 332.5 | 850.5 |
|  | Calipso ( $\mathrm{n}=344$ ) <br> Women of childbearing age (18-44 y) |  | $\begin{gathered} \text { INCA ( } n=404 \text { ) } \\ \text { Consumers only ( } 82.7 \% \text { ) } \end{gathered}$ |  |
| Fish | 391.9 | 870.5 | 224.3 | 580.0 |
| Mollusc, crustacean | 174.1 | 487.9 | 109.3 | 250.0 |
| Seafood-based dishes | 262.4 | 691.9 | 174.4 | 650.0 |
| Total | 828.4 | 1,806 | 300.3 | 785.0 |

## Appendix 4: Distribution of provisioning per site (\% of consumed product)

Fish

| Le Havre | Fished or collected | Bought at the port | Bought at the market | Bought from a fishmonger |  | Consumed only outside the home |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchovy (excluding canned) | - | - | 4.0 | 57.3 | 38.7 | - |
| Angler fish* | 0.4 | 7.6 | 2.1 | 20.0 | 68.1 | 1.9 |
| Catshark* | 0.5 | 19.0 | 2.0 | 12.1 | 65.6 | 0.8 |
| Cod | 0.1 | 4.3 | 2.6 | 14.1 | 76.7 | 2.2 |
| Dab | 4.1 | 12.8 | 16.4 | 18.1 | 44.9 | 3.7 |
| Eel* | 3.9 | 16.0 | 5.7 | 65.5 | 7.9 | 1.0 |
| Emperor* | - | 0.5 | 12.7 | 26.5 | 60.3 | - |
| Goatfish | 2.0 | 15.8 | 2.6 | 21.2 | 42.1 | 16.3 |
| Grenadier / Hoki* | - | 0.9 | 0.9 | 13.5 | 84.5 | - |
| Grouper | 57.1 | - | - | - | 42.9 | - |
| Gurnard | - | 4.6 | - | 25.7 | 69.7 | - |
| Haddock | - | 0.2 | 6.0 | 15.4 | 78.5 | - |
| Hake | 1.0 | 0.3 | 14.4 | 6.6 | 77.7 | - |
| Halibut* | - | 0.6 | 0.3 | 4.5 | 93.2 | 1.4 |
| Herring (excluding smoked) | 1.7 | 10.7 | 5.8 | 30.7 | 50.7 | 0.4 |
| John dory | 3.3 | 9.6 | - | 34.3 | 52.9 | - |
| Ling | 0.5 | 3.2 | 2.8 | 16.7 | 75.1 | 1.8 |
| Mackerel | 10.5 | 37.4 | 1.7 | 11.0 | 37.3 | 2.2 |
| Mullet | 69.2 | - | - | - | 22.6 | 8.1 |
| Plaice | 2.4 | 38.1 | 5.1 | 12.9 | 41.4 | - |
| Pollack | 2.2 | 4.4 | 8.4 | 13.9 | 55.9 | 15.2 |
| Pout | 15.5 | 43.8 | - | 7.8 | 32.9 | - |
| Ray* | 4.1 | 6.4 | 3.3 | 16.4 | 65.0 | 4.9 |
| Redfish | - | - | - | - | 100.0 | - |
| Saithe / Coalfish |  | 3.9 | 2.1 | 10.9 | 82.4 | 0.8 |
| Salmon (excluding smoked) | 0.0 | 0.9 | 1.7 | 17.1 | 75.8 | 4.5 |
| Sardine (excluding canned) | 0.3 | 5.0 | 4.3 | 23.2 | 65.1 | 2.2 |
| Scorpion fish | - | - | - | 36.0 | 64.0 | - |
| Seabass* | 27.2 | 18.2 | 1.3 | 13.1 | 27.9 | 12.2 |
| Sea bream* | 2.3 | 24.2 | 5.0 | 16.8 | 48.9 | 2.7 |
| Smelt | 2.1 | 24.2 | - | 36.6 | 37.1 | - |
| Sole | 5.9 | 30.6 | 0.6 | 17.7 | 34.0 | 11.2 |
| Sprat | 37.5 | - | 12.5 | 12.5 | 37.5 | - |
| Swordfish* | - | 1.4 | - | 5.1 | 93.4 | - |
| Tuna (excluding canned)* | - | 1.3 | 1.5 | 7.3 | 87.2 | 2.7 |
| Turbot | 4.8 | 21.6 | - | 39.2 | 29.2 | 5.2 |
| Whiting | 1.4 | 9.8 | 3.1 | 21.0 | 64.0 | 0.7 |
| Other** | 19.6 | - | - | - | 80.4 | - |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 <br> ** Other: Perch, trout, scabbard-fish, sea wrasse, conger eel, croaker, shark, bassbu |  |  |  |  |  |  |

Fish

| Lorient | Fished or collected | Bought at the port | Bought at the market | Bought from a fishmonger | Bought from a supermarket | Consumed only outside the home |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchovy (excluding canned) | 1.5 | 2.1 | 12.5 | 68.3 | 14.6 | 1.0 |
| Angler fish* | 5.6 | 9.9 | 13.3 | 34.7 | 33.9 | 2.6 |
| Catshark* | 1.3 | - | 3.6 | 57.8 | 37.3 | - |
| Cod | 2.7 | 13.4 | 7.1 | 30.6 | 44.8 | 1.4 |
| Dab | 2.6 | 14.0 | 7.3 | 49.0 | 24.0 | 3.1 |
| Eel* | 52.8 | 1.6 | 7.6 | 17.3 | 20.7 | - |
| Emperor* | 1.8 | 4.0 | 10.1 | 26.4 | 55.2 | 2.4 |
| Goatfish | 8.5 | 21.6 | 14.5 | 42.0 | 12.6 | 0.7 |
| Grenadier / Hoki* | 3.6 | 2.4 | 2.9 | 38.2 | 45.4 | 7.5 |
| Grouper | - | - | - | 100.0 | - | - |
| Gurnard | 3.9 | 20.1 | 4.3 | 41.6 | 30.0 | - |
| Haddock | 3.9 | 0.8 | 2.5 | 35.1 | 57.7 | - |
| Hake | 3.4 | 26.4 | 11.7 | 29.8 | 25.9 | 2.7 |
| Halibut* | 3.2 | 1.8 | 22.4 | 29.6 | 37.6 | 5.3 |
| Herring (excluding smoked) | 30.4 | 15.4 | 12.8 | 26.2 | 13.9 | 1.3 |
| John Dory | 5.9 | - | 27.9 | 31.4 | 8.3 | 26.5 |
| Ling | 1.4 | 3.8 | 5.0 | 45.3 | 43.5 | 0.9 |
| Mackerel | 19.0 | 30.6 | 8.6 | 26.6 | 15.3 | - |
| Mullet | 55.9 | 44.1 | - | - | - | - |
| Plaice | 7.6 | 3.7 | 10.6 | 34.2 | 43.9 | - |
| Pollack | 13.2 | 12.7 | 5.4 | 40.6 | 27.4 | 0.5 |
| Pout | 15.8 | 38.1 | 26.9 | 12.0 | 5.5 | 1.7 |
| Ray* | 2.5 | 2.3 | 8.8 | 44.8 | 38.5 | 3.0 |
| Redfish | - | - | 33.5 | 66.5 | - | - |
| Saithe / Coalfish | 1.8 | 8.4 | 6.8 | 34.5 | 46.1 | 2.3 |
| Salmon (excluding smoked) | 0.8 | 1.1 | 6.7 | 35.5 | 53.2 | 2.8 |
| Sardine (excluding canned) | 3.4 | 9.9 | 15.3 | 40.8 | 28.6 | 1.8 |
| Scorpion fish | 26.5 | - | - | 28.5 | 19.9 | 25.2 |
| Seabass* | 31.7 | 21.7 | 7.8 | 21.6 | 14.2 | 3.0 |
| Sea bream* | 5.2 | 10.7 | 20.8 | 37.4 | 20.3 | 5.5 |
| Smelt | 16.3 | 13.2 | - | 53.6 | 16.9 | - |
| Sole | 16.3 | 20.4 | 3.9 | 46.4 | 10.2 | 2.6 |
| Sprat | - | - | 50.0 | - | 50.0 | - |
| Swordfish* | 2.6 | 8.8 | 27.1 | 31.4 | 23.9 | 6.3 |
| Tuna (excluding canned)* | 4.1 | 29.0 | 7.5 | 28.2 | 29.2 | 2.0 |
| Turbot | 12.8 | 0.7 | 17.9 | 22.8 | 13.2 | 32.7 |
| Whiting | 4.2 | 24.4 | 10.5 | 32.4 | 28.1 | 0.5 |
| Other** | 10.8 | - | - | 7.5 | 81.7 | - |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 <br> ** Other: Perch, trout, scabbard-fish, sea wrasse, conger eel, croaker, shark, bassbu |  |  |  |  |  |  |

Fish

| La Rochelle | Fished or collected | Bought at the port | Bought at the market | Bought from a fishmonger | Bought from a supermarket | Consumed only outside the home |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchovy (excluding canned) | 1.2 | 0.5 | 79.7 | 3.7 | 14.5 | 0.4 |
| Angler fish* | 1.1 | 2.1 | 33.9 | 20.7 | 33.5 | 8.7 |
| Catshark* | 5.9 | 18.2 | 21.9 | 15.0 | 29.2 | 9.7 |
| Cod | 0.4 | 4.8 | 29.3 | 17.3 | 47.3 | 0.8 |
| Dab | 5.2 | 23.2 | 19.2 | 17.7 | 31.6 | 3.2 |
| Eel | 24.1 | 1.9 | 30.6 | 11.0 | 27.8 | 4.6 |
| Emperor* | - | 8.5 | 36.1 | 20.9 | 29.1 | 5.5 |
| Goatfish | 3.3 | 6.1 | 49.4 | 16.7 | 22.7 | 1.8 |
| Grenadier / Hoki* | 0.4 | 16.7 | 21 | 19.9 | 34.4 | 7.6 |
| Grouper | - | - | 32.9 | 43.5 | 18.4 | 5.2 |
| Gurnard | - | 8.9 | 26.8 | 15.4 | 46.6 | 2.3 |
| Haddock | - | 24.8 | 19.7 | 14.0 | 37.3 | 4.1 |
| Hake | 2.5 | 10.2 | 29.5 | 24.1 | 33.4 | 0.3 |
| Halibut* | 0.0 | 2.1 | 34.2 | 25.1 | 37.6 | 1.0 |
| Herring (excluding smoked) | 1.0 | 13.5 | 25.0 | 17.1 | 43.4 | - |
| John Dory | 1.0 | 31.4 | 42.0 | 9.8 | 7.2 | 8.6 |
| Ling | 0.3 | 13.0 | 15.4 | 18.6 | 39.0 | 13.8 |
| Mackerel | 16.4 | 14.0 | 24.0 | 11.2 | 34.0 | 0.4 |
| Mullet | 38.7 | 30.4 | 18.5 | 4.1 | 8.3 | - |
| Plaice | 4.2 | 42 | 23.2 | 9.2 | 20.6 | 0.8 |
| Pollack | 5.5 | 19.0 | 27.3 | 16.7 | 31.0 | 0.5 |
| Pout | 33.1 | 15.1 | 21.4 | 10.1 | 16.6 | 3.7 |
| Ray* | 3.0 | 6.9 | 35.6 | 24.1 | 27.2 | 3.3 |
| Redfish | - | 50.3 | - | - | 34.8 | 14.8 |
| Saithe / Coalfish | 0.1 | 8.0 | 20.6 | 12.9 | 56.8 | 1.6 |
| Salmon (excluding smoked) | 0.2 | 2.4 | 24.8 | 18.8 | 45.8 | 7.9 |
| Sardine (excluding canned) | 1.2 | 6.2 | 33.5 | 21.4 | 36.4 | 1.3 |
| Scorpion fish | - | 14.3 | 39.5 | 27.4 | 12.4 | 6.4 |
| Seabass* | 34.8 | 2.1 | 36.0 | 13.9 | 8.6 | 4.6 |
| Sea bream* | 18.2 | 12.7 | 31.1 | 9.6 | 26.1 | 2.0 |
| Smelt | 17.6 | 1.2 | 19.2 | 22.6 | 32.8 | 6.7 |
| Sole | 12.4 | 5.3 | 35.8 | 14.5 | 28.2 | 3.9 |
| Sprat | - | 12.7 | 11.4 | 19.0 | 47.5 | 9.5 |
| Swordfish* | - | 5.6 | 60.2 | 19.4 | 11.3 | 4.1 |
| Tuna (excluding canned)* | 3.9 | 3.5 | 33.8 | 21.2 | 34.8 | 2.8 |
| Turbot | 2.2 | 8.7 | 19.8 | 8.8 | 35.4 | 25.0 |
| Whiting | 4.2 | 9.2 | 27.6 | 18.8 | 35.3 | 5.0 |
| Other** | 27.5 | 5.2 | 37.2 | 9.2 | 17.3 | 3.6 |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 <br> ** Other: Perch, trout, scabbard-fish, sea wrasse, conger eel, croaker, shark, bassbu |  |  |  |  |  |  |

Fish

| Toulon | Fished or collected | Bought at the port | Bought at the market | Bought from a fishmonger | Bought from a supermarket | Consumed only outside the home |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchovy (excluding canned) | - | 1.2 | 4.9 | 40.5 | 39.6 | 13.7 |
| Angler fish* | - | 0.9 | 2.2 | 30.5 | 58.6 | 7.8 |
| Catshark* | - | 2.2 | - | 38.7 | 55.0 | 3.7 |
| Cod | - | 0.6 | 2.3 | 16.9 | 77.7 | 2.6 |
| Dab | 1.0 | 0.9 | 4.9 | 38.4 | 50.6 | 4.2 |
| Eel* | 6.3 | 8.1 | 0.6 | 17.5 | 61.3 | 6.3 |
| Emperor* | - | 7.1 | - | 40.2 | 50.1 | 2.6 |
| Goatfish | 1.2 | 6.7 | 1.3 | 36.6 | 47.1 | 7.1 |
| Grenadier / Hoki* | - | 4.8 | - | 16.3 | 76.8 | 2.1 |
| Grouper | - | 3.2 | 4.3 | 31.1 | 29.1 | 32.3 |
| Gurnard | 2.7 | 8.9 | 3.7 | 13.2 | 58.2 | 13.3 |
| Haddock | - | 1.2 | - | 13.2 | 83.1 | 2.4 |
| Hake | - | - | - | 10.3 | 86.8 | 2.9 |
| Halibut* | - | 1.7 | 0.5 | 26.2 | 68.8 | 2.8 |
| Herring (excluding smoked) | - | - | - | 31.4 | 68.6 | - |
| John Dory | - | 7.2 | 3.1 | 26.0 | 29.9 | 33.7 |
| Ling | - | 1.3 | 6.6 | 23.0 | 67.0 | 2.1 |
| Mackerel | 3.5 | 2.3 | 0.9 | 34.2 | 57.2 | 1.9 |
| Mullet | 21.2 | 34.6 | 1.3 | 10.9 | 28.3 | 3.8 |
| Plaice | 4.6 | 2.9 | - | 18.1 | 71 | 3.4 |
| Pollack | - | 0.8 | 1.6 | 35.0 | 62.5 | - |
| Pout | - | - | - | 34.1 | 65.9 | - |
| Ray* | - | 0.3 | 0.2 | 21.5 | 68.7 | 9.4 |
| Redfish | 14.8 | 3.4 | - | - | 72.1 | 9.7 |
| Saithe / Coalfish | - | 0.8 | 1.9 | 13.4 | 83.9 | - |
| Salmon (excluding smoked) | - | 0.2 | 0.4 | 15.5 | 80.6 | 3.2 |
| Sardine (excluding canned) | 0.2 | 6.3 | 5.2 | 34.1 | 51.9 | 2.2 |
| Scorpion fish | 4.4 | 3.4 | 8.9 | 30.3 | 36.2 | 16.7 |
| Seabass* | 2.0 | 2.5 | 5.7 | 30.8 | 45.7 | 13.2 |
| Sea bream* | 6.9 | 6.3 | 5.6 | 32.2 | 40.6 | 8.4 |
| Smelt | - | - | 0.5 | 22.2 | 60.9 | 16.4 |
| Sole | - | 2.9 | 3.4 | 34.3 | 54.7 | 4.7 |
| Sprat | - | - | - | 63.2 | 13.8 | 23.0 |
| Swordfish* | - | 2.1 | 7.6 | 15.0 | 57.8 | 17.5 |
| Tuna (excluding canned)* | 0.4 | 3.6 | 5.0 | 30.7 | 57.4 | 2.9 |
| Turbot | - | 7.3 | 5.0 | 32.8 | 37.8 | 17.2 |
| Whiting | - | 3.8 | 4.5 | 27.5 | 63.2 | 1.0 |
| Other** | 22.0 | 8.6 | - | 27.6 | 41.8 | - |
| * Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005 <br> ** Other : Bogue, roquier, trout, conger eel, sard, shark, coriphaire, rainbow wrasse, gobie |  |  |  |  |  |  |

## Mollusc, crustacean

| Le Havre | Fished or collected | Bought at the port | Bought at the market | Bought from a fishmonger | Bought from a supermarket | Consumed only outside the home |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Abalone | 27.3 | - | 10.9 | - | 25.5 | 36.4 |
| Calico scallop | 2.0 | 9.7 | 1.9 | 6.6 | 77.8 | 2.0 |
| Carpet shell | 19.9 | 3.3 | - | 37.7 | 36.4 | 2.6 |
| Clam | 62.5 | - | - | - | 37.5 | - |
| Cockle | 16.5 | 1.4 | 14.7 | 22.0 | 43.2 | 2.1 |
| Crab | 2.6 | 17.1 | 3.1 | 18.3 | 57.3 | 1.6 |
| Crayfish | 12.6 | - | 0.5 | 26.0 | 59.3 | 1.6 |
| Cuttle fish | - | 28.5 | 2.1 | 23.5 | 43.4 | 2.4 |
| Donax clam | 71.4 | - | - | - | 28.6 | - |
| Great scallop | 2.8 | 13.8 | 2.8 | 19.1 | 56.2 | 5.3 |
| Grooved sea squirt | - | - | - | - | - | - |
| Hard clam | 0.3 | 0.3 | 3.8 | 33.5 | 59.9 | 2.2 |
| Limpet | 100.0 | - | - | - | - | - |
| Lobster | 3.6 | 3.1 | 1.0 | 10.2 | 69.8 | 12.3 |
| Mussel | 0.4 | 0.3 | 8.4 | 25.4 | 61.4 | 4.0 |
| Octopus | - | 2.2 | 10.7 | 17.7 | 65.1 | 4.4 |
| Oyster | 2.0 | 4.6 | 24.4 | 27.6 | 39.9 | 1.5 |
| Periwinkle | 4.8 | 1.8 | 6.7 | 23.8 | 61.4 | 1.4 |
| Queen scallop | - | - | - | - | - | 100.0 |
| Razor clam | - | - | - | - | - | - |
| Scampi | 1.9 | 1.2 | 4.0 | 24.7 | 62.7 | 5.4 |
| Sea urchin | - | - | - | - | 100.0 | - |
| Shrimp | 1.7 | 2.3 | 5.0 | 18.7 | 71.1 | 1.2 |
| Slipper lobster | - | - | - | - | - | - |
| Spider crab | 21.5 | 17.8 | 3.8 | 24.5 | 31.5 | 0.9 |
| Spiny lobster | - | 5.2 | 0.4 | 11.2 | 76.1 | 7.1 |
| Squid | - | 7.1 | 1.1 | 11.8 | 77.3 | 2.7 |
| Swimcrab | 12.2 | 25.9 | 5.6 | 22.0 | 34.0 | 0.3 |
| Whelk | 0.2 | 3.5 | 9.4 | 22.3 | 64.2 | 0.4 |

Mollusc, crustacean

| Lorient | Fished or collected | Bought at the port | Bought at the market | Bought from a fishmonger | Bought from a supermarket | Consumed only outside the home |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Abalone | - | 100.0 | - | - | - | - |
| Calico scallop | 9.7 | 7.8 | 12.3 | 19.7 | 48.6 | 1.9 |
| Carpet shell | 49.1 | 11.4 | 12.2 | 21.9 | 3.3 | 2.0 |
| Clam | 41.9 | - | 21.3 | 27.1 | 4.8 | 4.8 |
| Cockle | 65.0 | 0.9 | 12.1 | 12.8 | 9.2 | - |
| Crab | 7.8 | 15.7 | 13.9 | 41.5 | 20.2 | 0.8 |
| Crayfish | 6.1 | - | 13.3 | 13.9 | 48.6 | 18.1 |
| Cuttle fish | 7.3 | 4.9 | 1.2 | 29.5 | 57.2 | - |
| Donax clam | 68.8 | - | - | 31.3 | - | - |
| Great scallop | 0.5 | 4.2 | 6.8 | 23.1 | 62.3 | 2.8 |
| Grooved sea squirt | 100.0 | - | - | - | - | - |
| Hard clam | 33.6 | 13.6 | 11.7 | 23.1 | 15.5 | 2.6 |
| Limpet | 100.0 | - | - | - | - | - |
| Lobster | 17.5 | 4.8 | 6.0 | 42.8 | 20.5 | 8.4 |
| Mussel | 5.0 | 2.2 | 17.1 | 36.9 | 33.9 | 4.9 |
| Octopus | - | - | 14.7 | 49.2 | 36.1 | - |
| Oyster | 7.5 | 9.5 | 34.3 | 39.5 | 8.4 | 1.0 |
| Periwinkle | 24.6 | 2.9 | 22.6 | 35.3 | 14.3 | 0.4 |
| Queen scallop | - | - | - | - | - | 100.0 |
| Razor clam | 88.2 | - | - | 11.8 | - | - |
| Scampi | 3.2 | 14.2 | 15.2 | 46.5 | 20.2 | 0.7 |
| Sea urchin | 23.6 | 21.8 | 5.5 | 21.8 | - | 27.3 |
| Shrimp | 12.3 | 2.4 | 10.6 | 34.9 | 38.9 | 1.5 |
| Slipper lobster | - | - | - | - | - | - |
| Spider crab | 14.3 | 15.3 | 13.1 | 37.5 | 18.3 | 1.5 |
| Spiny lobster | 7.4 | 3.7 | 7.4 | 56.3 | 17.8 | 7.4 |
| Squid | 10.2 | 7.6 | 5.7 | 30.4 | 43.4 | 2.8 |
| Swimcrab | 46.9 | 6.6 | 13.7 | 21.0 | 11.8 | - |
| Whelk | 1.7 | 3.6 | 9.6 | 48.9 | 32.7 | 3.4 |

## Mollusc, crustacean

| La Rochelle | Fished or collected | Bought at the port | Bought at the market | Bought from a fishmonger | Bought from a supermarket | Consumed only outside the home |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Abalone | - | - | 7.7 | - | 84.6 | 7.7 |
| Calico scallop | 21.9 | 8.8 | 20.2 | 19.9 | 27.9 | 1.3 |
| Carpet shell | 32.3 | 20.8 | 21.1 | 9.7 | 11.8 | 4.2 |
| Clam | - | 33.8 | 12.8 | 15.5 | 30.4 | 7.4 |
| Cockle | 18.3 | 17.2 | 28.9 | 21.2 | 12.2 | 2.2 |
| Crab | 4.3 | 9.7 | 22.1 | 25.8 | 33.7 | 4.5 |
| Crayfish | 33.2 | 7.5 | 25.8 | 15.4 | 5.5 | 12.6 |
| Cuttle fish | 7.2 | 11.3 | 20.5 | 28.1 | 32.3 | 0.6 |
| Donax clam | - | - | - | - | - | - |
| Great scallop | 3.0 | 13.7 | 21.2 | 22.7 | 35.9 | 3.6 |
| Grooved sea squirt | - | - | - | - | - | - |
| Hard clam | 6.5 | 26.0 | 28.8 | 17.7 | 19.9 | 1.1 |
| Limpet | 10.2 | 15.5 | 37.2 | - | 37.2 | - |
| Lobster | 3.0 | 10.0 | 21.4 | 29.4 | 21.2 | 15.0 |
| Mussel | 3.6 | 7.0 | 33.9 | 28.2 | 22.8 | 4.6 |
| Octopus | 6.4 | - | 53.6 | 15.5 | 20.6 | 3.9 |
| Oyster | 10.2 | 13.2 | 39.5 | 17.8 | 18.3 | 1.0 |
| Periwinkle | 24.0 | 0.8 | 36.0 | 15.3 | 19.1 | 4.8 |
| Queen scallop | - | 42.4 | 24.6 | 10.6 | 22.5 | - |
| Razor clam | 72.6 | 4.3 | - | 5.1 | 5.1 | 12.8 |
| Scampi | 1.7 | 4.5 | 28.6 | 36.8 | 23.5 | 4.9 |
| Sea urchin | 39.1 | - | 26.1 | 6.5 | 21.7 | 6.5 |
| Shrimp | 5.8 | 3.7 | 27.4 | 28.1 | 33.6 | 1.3 |
| Slipper lobster | - | - | 100.0 | - | - | - |
| Spider crab | 17.0 | 16.4 | 29.3 | 13.5 | 23.8 | - |
| Spiny lobster | 2.4 | 27.4 | 14.3 | 11.9 | 28.6 | 15.5 |
| Squid | 3.4 | 11.9 | 24.8 | 18.5 | 37.7 | 4.1 |
| Swimcrab | 26.3 | 0.3 | 28.2 | 35.8 | 8.5 | 0.9 |
| Whelk | 1.6 | 11.1 | 34.6 | 15.6 | 25.9 | 11.1 |

Mollusc, crustacean

| Toulon | Fished or collected | Bought at the port | Bought at the market | Bought from a fishmonger | Bought from a supermarket | Consumed only outside the home |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Abalone | 25.0 | - | - | 4.2 | 12.5 | 58.3 |
| Calico scallop | - | 2.0 | 0.7 | 24.3 | 59.6 | 13.5 |
| Carpet shell | 4.2 | 3.3 | 2.8 | 26.5 | 35.4 | 27.8 |
| Clam | 1.3 | 5.0 | 2.5 | 31.7 | 19.6 | 40.0 |
| Cockle | 7.5 | - | - | 35.2 | 45.2 | 12.1 |
| Crab | 6.2 | 0.9 | 0.9 | 35.9 | 37.1 | 18.9 |
| Crayfish | 5.3 | - | 0.8 | 12.0 | 48.1 | 33.8 |
| Cuttle fish | 5.5 | 2.5 | 1.3 | 18.3 | 67.7 | 4.7 |
| Donax clam | 10.0 | - | - | 37.6 | 36.4 | 16.0 |
| Great scallop | 0.3 | 0.8 | - | 12.5 | 79.2 | 7.2 |
| Grooved sea squirt | 3.6 | 14.9 | 9.2 | 34.4 | 19.0 | 19.0 |
| Hard clam | 2.1 | 5.1 | 1.7 | 33.9 | 21.4 | 35.8 |
| Limpet | - | - | - | 33.3 | - | 66.7 |
| Lobster | 1.9 | 2.1 | 2.7 | 17.8 | 48.8 | 26.7 |
| Mussel | 0.2 | 1.3 | 3.3 | 26.5 | 51.4 | 17.2 |
| Octopus | 16.0 | 3.2 | 1.8 | 17.1 | 52.0 | 9.9 |
| Oyster | 1.1 | 2.1 | 7.1 | 29.8 | 43.2 | 16.7 |
| Periwinkle | 9.0 | 1.0 | - | 26.9 | 39.4 | 24.2 |
| Queen scallop | - | - | - | - | - | 100.0 |
| Razor clam | - | - | - | - | 54.5 | 45.5 |
| Sea urchin | 59.6 | 16.9 | 1.5 | 9.8 | 9.2 | 3.0 |
| Scampi | - | 1.0 | 0.7 | 25.6 | 59.3 | 13.4 |
| Shrimp | 0.2 | 1.0 | 1.4 | 17.7 | 75.5 | 4.2 |
| Slipper lobster | 7.9 | 15.8 | - | 31.6 | 28.9 | 15.8 |
| Spider crab | 17.6 | 5.9 | 1.5 | 36.8 | 35.3 | 2.9 |
| Spiny lobster | - | 2.6 | 0.9 | 16.1 | 57.4 | 23.9 |
| Squid | 4.8 | 2.7 | 0.9 | 12.9 | 74.4 | 4.4 |
| Swimcrab | 0.3 | 1.0 | 1.0 | 14.8 | 55.2 | 27.6 |
| Whelk | - | 25.4 | 3.4 | 24.2 | 33.6 | 13.5 |

## Appendix 5:

Contributors to the total exposure to omega 3 and contaminants (\%) - All subjects, all areas

| Species | $\begin{aligned} & \text { OMEGA } 3 \\ & \mathrm{n}-3^{* *} \end{aligned}$ | TRACE ELEMENTS |  |  |  |  |  | PERSISTENT ORGANIC POLLUTANTS |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | MeHg | Cd | Pb | OTC | $\mathrm{As}_{\text {T }}$ | As inorg | PBDE | PCDD/F | PCB-DL | Tot diox | iPCB |
| Anchovy | 4.85 | 0.59 | 9.93 | 5.20 | 3.66 | 0.57 | 1.74 | 3.59 | 0.62 | 0.97 | 0.88 | 0.81 |
| Angler fish* | 0.11 | 2.23 | 0.01 | 0.26 | 0.96 | 1.96 | 0.67 | 0.85 | 0.27 | 0.22 | 0.23 | 0.40 |
| Calico scallop | 0.43 | 0.07 | 4.76 | 1.61 | 0.53 | 0.41 | 0.23 | 0.19 | 0.54 | 0.18 | 0.28 | 0.29 |
| Catshark* | 0.27 | 1.70 | 2.68 | 1.00 | 1.13 | 4.55 | 1.40 | 0.33 | 0.18 | 0.18 | 0.18 | 0.38 |
| Cockle | 0.04 | 0.06 | 0.12 | 0.25 | 0.47 | 0.08 | 0.65 | 0.07 | 0.07 | 0.05 | 0.05 | 0.03 |
| Cod | 1.38 | 6.89 | 0.10 | 1.59 | 5.89 | 10.93 | 6.04 | 6.26 | 1.79 | 2.17 | 2.02 | 1.78 |
| Crab | 1.90 | 2.47 | 14.61 | 0.91 | 1.17 | 2.49 | 4.89 | 1.05 | 5.78 | 2.88 | 3.60 | 2.81 |
| Cuttle fish | 0.24 | 0.38 | 0.46 | 1.20 | 0.41 | 0.89 | 0.57 | 0.15 | 0.17 | 0.07 | 0.09 | 0.15 |
| Dab | 0.38 | 1.29 | 0.01 | 0.13 | 0.40 | 3.75 | 1.19 | 0.82 | 1.13 | 0.64 | 0.75 | 0.40 |
| Eel* | 0.66 | 1.01 | 0.12 | 0.42 | 0.33 | 0.06 | 0.10 | 4.39 | 1.56 | 7.00 | 6.44 | 8.28 |
| Emperor* | 0.69 | 1.24 | 0.02 | 0.18 | 0.27 | 0.06 | 0.09 | 0.37 | 1.36 | 1.50 | 1.44 | 1.29 |
| Fish soup | 1.90 | 0.57 | 2.05 | 2.12 | 1.39 | 0.99 | 4.13 | 1.79 | 1.19 | 1.05 | 1.07 | 0.82 |
| Goatfish | 1.08 | 1.07 | 0.01 | 0.12 | 0.21 | 2.48 | 1.41 | 0.70 | 2.13 | 2.23 | 2.18 | 1.69 |
| Great scallop | 1.10 | 1.41 | 9.04 | 10.67 | 6.10 | 2.40 | 8.55 | 1.31 | 3.25 | 1.29 | 1.80 | 2.36 |
| Grenadier/Hoki* | 0.23 | 1.57 | 0.07 | 0.25 | 0.96 | 1.08 | 0.51 | 0.74 | 0.43 | 0.20 | 0.26 | 0.49 |
| Gurnard | 0.02 | 0.18 | 0.0004 | 0.001 | 0.04 | 0.11 | 0.05 | 0.08 | 0.19 | 0.15 | 0.16 | 0.15 |
| Haddock | 0.07 | 0.39 | 0.06 | 0.11 | 0.28 | 0.47 | 0.28 | 0.28 | 0.12 | 0.10 | 0.10 | 0.11 |
| Hake | 0.38 | 4.56 | 0.01 | 1.19 | 2.31 | 2.57 | 1.60 | 1.47 | 0.53 | 1.23 | 1.03 | 1.33 |
| Halibut* | 3.37 | 0.96 | 0.95 | 2.81 | 2.89 | 1.31 | 0.61 | 1.77 | 3.35 | 2.00 | 2.33 | 1.76 |
| Herring | 4.66 | 0.25 | 0.002 | 0.13 | 0.32 | 0.14 | 0.18 | 0.58 | 0.82 | 0.35 | 0.46 | 0.32 |
| John dory | 0.09 | 0.27 | 0.10 | 0.13 | 0.44 | 0.06 | 0.15 | 0.16 | 0.09 | 0.16 | 0.15 | 0.19 |
| Ling | 0.24 | 5.73 | 0.17 | 0.03 | 1.19 | 1.89 | 0.73 | 1.14 | 0.39 | 0.38 | 0.37 | 0.48 |
| Lobster | 0.16 | 0.15 | 1.13 | 0.03 | 0.03 | 0.23 | 0.18 | 0.08 | 0.65 | 0.25 | 0.34 | 0.11 |
| Mackerel | 11.52 | 2.29 | 1.17 | 0.59 | 6.22 | 1.42 | 2.80 | 8.55 | 6.27 | 6.84 | 6.62 | 6.84 |
| Mussel | 0.98 | 1.15 | 1.86 | 10.47 | 1.34 | 2.89 | 4.61 | 1.34 | 2.32 | 1.17 | 1.45 | 1.08 |
| Octopus | 0.05 | 0.74 | 0.26 | 0.87 | 0.28 | 2.71 | 1.69 | 0.09 | 0.15 | 0.15 | 0.15 | 0.12 |
| Oyster | 0.68 | 0.33 | 3.22 | 5.08 | 2.54 | 1.28 | 7.04 | 0.91 | 3.29 | 1.51 | 1.91 | 1.06 |
| Paella | 2.55 | 0.04 | 1.72 | 3.85 | 1.38 | 0.17 | 1.09 | 1.48 | 0.96 | 0.23 | 0.42 | 0.17 |
| Periwinkle | 0.23 | 0.04 | 0.73 | 1.31 | 0.26 | 0.50 | 1.56 | 0.28 | 0.09 | 0.05 | 0.06 | 0.05 |
| Pilchard | 0.44 | 0.04 | 0.03 | 0.03 | 0.10 | 0.06 | 0.29 | 0.51 | 0.52 | 0.20 | 0.28 | 0.16 |
| Plaice | 0.11 | 0.79 | 0.00 | 0.17 | 0.58 | 1.63 | 0.18 | 0.61 | 0.70 | 0.55 | 0.58 | 0.50 |
| Pollack | 0.15 | 0.98 | 0.02 | 0.02 | 0.71 | 0.79 | 1.55 | 0.59 | 0.09 | 0.51 | 0.40 | 0.60 |
| Pout | 0.02 | 0.21 | 0.001 | 0.02 | 0.10 | 0.28 | 0.20 | 0.06 | 0.02 | 0.03 | 0.03 | 0.03 |
| Ray* | 0.52 | 2.89 | 2.36 | 2.95 | 1.08 | 9.94 | 5.15 | 1.27 | 1.04 | 0.49 | 0.62 | 0.44 |
| Sea bream* | 3.45 | 2.23 | 0.02 | 0.07 | 1.82 | 1.77 | 2.55 | 2.88 | 4.10 | 5.40 | 5.04 | 4.88 |
| Saithe / Coalfish | 1.76 | 3.03 | 4.04 | 7.17 | 6.12 | 1.95 | 3.48 | 4.73 | 0.77 | 1.18 | 1.06 | 1.02 |
| Salmon | 26.59 | 3.72 | 0.04 | 0.82 | 6.45 | 3.10 | 4.78 | 18.76 | 14.33 | 14.18 | 13.99 | 12.86 |
| Sardine | 9.53 | 2.75 | 5.24 | 17.23 | 4.24 | 3.86 | 4.17 | 6.52 | 15.84 | 20.10 | 18.91 | 20.39 |
| Scorpion fish | 0.30 | 0.48 | 0.001 | 0.002 | 0.06 | 0.10 | 0.10 | 0.18 | 0.60 | 0.63 | 0.62 | 0.47 |
| Seabass* | 2.72 | 3.26 | 0.02 | 1.37 | 4.39 | 0.96 | 1.57 | 4.82 | 5.25 | 8.09 | 7.35 | 8.01 |
| Sea urchin | 0.15 | 0.03 | 1.24 | 3.42 | 0.68 | 0.61 | 3.01 | 0.29 | 0.25 | 0.43 | 0.39 | 0.21 |
| Scampi | 0.33 | 1.93 | 1.23 | 2.19 | 1.04 | 2.55 | 4.40 | 0.60 | 1.46 | 0.49 | 0.74 | 0.39 |
| Shrimp | 0.69 | 1.51 | 15.53 | 1.83 | 1.67 | 1.22 | 1.56 | 1.57 | 0.93 | 0.39 | 0.52 | 0.25 |
| Sole | 0.59 | 5.71 | 0.18 | 1.10 | 1.85 | 9.69 | 1.71 | 1.92 | 1.11 | 1.05 | 1.04 | 2.46 |
| Spider crab | 0.59 | 0.10 | 1.03 | 0.65 | 0.10 | 1.49 | 0.97 | 0.95 | 2.08 | 1.11 | 1.36 | 0.64 |
| Squid | 0.50 | 1.16 | 1.47 | 0.62 | 2.57 | 1.84 | 0.48 | 1.19 | 1.79 | 1.18 | 1.32 | 1.03 |
| Surimi | 2.29 | 0.88 | 0.74 | 0.77 | 4.39 | 0.39 | 1.22 | 2.57 | 0.31 | 0.19 | 0.22 | 0.76 |
| Swimcrab | 0.78 | 0.53 | 0.38 | 2.61 | 1.01 | 0.89 | 0.79 | 0.54 | 4.84 | 4.61 | 4.65 | 4.47 |
| Swordfish* | 1.83 | 3.78 | 0.80 | 0.01 | 1.43 | 0.11 | 0.42 | 0.53 | 0.21 | 0.36 | 0.32 | 0.30 |
| Tarama | 1.40 | 0.01 | 0.004 | 0.01 | 0.09 | 0.03 | 0.12 | 0.77 | 0.10 | 0.08 | 0.09 | 0.12 |
| Tuna* | 4.35 | 19.21 | 3.26 | 0.85 | 13.34 | 2.38 | 2.10 | 5.07 | 1.01 | 1.87 | 1.62 | 2.20 |
| Whelk | 0.37 | 0.64 | 6.95 | 3.35 | 1.24 | 3.81 | 2.24 | 0.57 | 2.15 | 0.39 | 0.91 | 0.28 |
| Whiting | 0.28 | 4.50 | 0.05 | 0.22 | 1.53 | 2.09 | 2.21 | 1.66 | 0.79 | 1.27 | 1.15 | 1.77 |

[^30]
## Contributors to the total exposure to omega 3 and contaminants (\%) - All subjects - Le Havre

| Species | OMEGA 3n-3** | TRACE ELEMENTS |  |  |  |  |  | PERSISTENT ORGANIC POLLUTANTS |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | MeHg | Cd | Pb | OTC | $\mathrm{As}_{\text {T }}$ | As ${ }_{\text {inorg }}$ | PBDE | PCDD/F | PCB-DL | Tot diox | iPCB |
| Anchovy | 1.94 | 0.29 | 1.68 | 2.43 | 1.10 | 0.21 | 0.64 | 1.41 | 0.21 | 0.37 | 0.33 | 0.31 |
| Angler fish* | 0.12 | 1.66 | 0.00 | 0.25 | 0.32 | 1.05 | 0.91 | 0.62 | 0.15 | 0.11 | 0.12 | 0.08 |
| Catshark* | 0.83 | 1.38 | 10.28 | 3.71 | 2.98 | 6.76 | 2.70 | 0.80 | 0.58 | 0.51 | 0.52 | 0.86 |
| Cod | 1.72 | 8.04 | 0.03 | 1.70 | 4.90 | 20.45 | 13.26 | 9.36 | 1.08 | 2.13 | 1.83 | 1.70 |
| Crab | 2.15 | 6.98 | 0.47 | 1.35 | 2.74 | 2.69 | 5.96 | 1.56 | 9.92 | 7.45 | 8.01 | 9.00 |
| Dab | 0.41 | 1.13 | 0.00 | 0.13 | 0.40 | 3.79 | 0.99 | 0.93 | 1.59 | 1.10 | 1.23 | 0.59 |
| Eel* | 0.38 | 0.78 | 0.34 | 0.46 | 0.17 | 0.04 | 0.06 | 2.09 | 0.85 | 2.97 | 2.76 | 3.44 |
| Fish soup | 0.71 | 0.34 | 0.21 | 0.86 | 0.41 | 0.42 | 1.50 | 0.71 | 0.34 | 0.31 | 0.32 | 0.22 |
| Great scallop | 2.12 | 0.59 | 13.84 | 22.21 | 13.30 | 2.53 | 15.33 | 1.18 | 5.65 | 2.71 | 3.52 | 4.71 |
| Grenadier/Hoki* | 0.13 | 1.22 | 0.00 | 0.89 | 0.30 | 0.63 | 0.27 | 0.93 | 0.36 | 0.15 | 0.21 | 0.29 |
| Hake | 0.24 | 0.93 | 0.00 | 0.02 | 0.39 | 0.22 | 0.49 | 0.41 | 0.10 | 0.24 | 0.20 | 0.26 |
| Halibut* | 3.71 | 1.99 | 0.00 | 0.02 | 3.64 | 1.82 | 0.46 | 2.73 | 3.27 | 3.09 | 3.12 | 2.58 |
| Herring | 7.23 | 0.28 | 0.00 | 0.14 | 0.19 | 0.13 | 0.15 | 0.41 | 0.57 | 0.25 | 0.33 | 0.21 |
| Ling | 0.35 | 7.62 | 0.00 | 0.02 | 0.46 | 1.66 | 0.55 | 1.09 | 0.20 | 0.21 | 0.20 | 0.23 |
| Mackerel | 16.10 | 3.58 | 0.54 | 0.60 | 8.00 | 1.44 | 2.16 | 16.35 | 13.38 | 13.42 | 13.32 | 13.31 |
| Mussel | 0.97 | 0.31 | 0.02 | 14.29 | 1.29 | 0.69 | 10.18 | 0.80 | 2.62 | 1.52 | 1.80 | 1.41 |
| Oyster | 0.45 | 0.37 | 0.01 | 0.02 | 3.31 | 0.98 | 9.13 | 0.32 | 2.64 | 1.18 | 1.55 | 0.58 |
| Paella | 2.43 | 0.05 | 0.49 | 4.04 | 1.01 | 0.16 | 0.99 | 1.38 | 0.86 | 0.22 | 0.39 | 0.14 |
| Periwinkle | 0.30 | 0.06 | 0.37 | 1.65 | 0.04 | 0.30 | 1.18 | 0.35 | 0.07 | 0.04 | 0.05 | 0.04 |
| Pilchard | 0.53 | 0.07 | 0.01 | 0.05 | 0.09 | 0.07 | 0.31 | 0.69 | 0.51 | 0.21 | 0.29 | 0.16 |
| Plaice | 0.30 | 2.97 | 0.00 | 0.62 | 1.93 | 5.86 | 0.43 | 2.20 | 2.61 | 2.05 | 2.18 | 1.83 |
| Pollack | 0.20 | 1.91 | 0.00 | 0.01 | 0.58 | 0.94 | 0.67 | 0.73 | 0.11 | 0.49 | 0.40 | 0.58 |
| Ray* | 0.51 | 2.91 | 0.03 | 0.31 | 1.20 | 9.69 | 5.24 | 1.30 | 0.80 | 0.47 | 0.55 | 0.44 |
| Saithe / Coalfish | 1.10 | 3.95 | 0.01 | 0.72 | 4.07 | 1.65 | 4.42 | 4.07 | 0.59 | 0.95 | 0.84 | 0.86 |
| Salmon | 31.46 | 5.57 | 0.02 | 0.48 | 7.33 | 3.21 | 3.03 | 18.29 | 9.60 | 12.13 | 11.36 | 11.01 |
| Sardine | 6.22 | 1.25 | 1.92 | 16.93 | 3.19 | 2.30 | 2.12 | 4.30 | 9.25 | 14.02 | 12.89 | 13.28 |
| Scampi | 0.15 | 0.50 | 0.67 | 1.10 | 1.07 | 1.49 | 0.90 | 0.18 | 1.03 | 0.22 | 0.45 | 0.14 |
| Seabass* | 1.60 | 4.97 | 0.00 | 1.81 | 2.03 | 0.64 | 0.47 | 5.31 | 4.90 | 6.57 | 6.14 | 5.37 |
| Sea bream* | 1.76 | 0.95 | 0.00 | 0.06 | 0.81 | 0.49 | 0.26 | 1.24 | 0.72 | 1.12 | 1.01 | 0.92 |
| Shrimp | 0.91 | 1.26 | 59.89 | 4.91 | 0.73 | 1.47 | 1.97 | 2.11 | 0.51 | 0.30 | 0.35 | 0.31 |
| Sole | 0.48 | 3.90 | 0.00 | 0.04 | 0.35 | 8.69 | 2.37 | 1.55 | 0.76 | 1.30 | 1.14 | 2.64 |
| Squid | 0.69 | 2.59 | 0.07 | 0.42 | 6.56 | 2.92 | 0.66 | 2.80 | 4.99 | 3.07 | 3.55 | 2.44 |
| Surimi | 3.44 | 1.51 | 0.41 | 1.23 | 5.04 | 0.58 | 1.70 | 3.73 | 0.47 | 0.31 | 0.35 | 1.07 |
| Swimcrab | 1.54 | 1.59 | 0.88 | 8.36 | 3.84 | 2.41 | 2.88 | 1.04 | 15.48 | 16.26 | 15.97 | 16.39 |
| Swordfish* | 0.72 | 1.72 | 0.07 | 0.00 | 0.52 | 0.04 | 0.07 | 0.23 | 0.09 | 0.09 | 0.09 | 0.10 |
| Tarama | 0.76 | 0.01 | 0.00 | 0.00 | 0.03 | 0.01 | 0.05 | 0.37 | 0.03 | 0.03 | 0.03 | 0.04 |
| Tuna* | 4.47 | 19.82 | 1.29 | 1.12 | 10.29 | 2.14 | 2.01 | 4.24 | 0.94 | 1.35 | 1.24 | 1.63 |
| Whelk | 0.76 | 1.69 | 6.41 | 6.93 | 3.87 | 7.50 | 1.61 | 1.35 | 1.95 | 0.86 | 1.16 | 0.64 |
| Whiting | 0.10 | 3.24 | 0.01 | 0.03 | 1.51 | 1.89 | 1.81 | 0.80 | 0.22 | 0.22 | 0.21 | 0.19 |

[^31]
## Contributors to the total exposure to omega 3 and contaminants (\%) - All subjects - Lorient

| Species | OMEGA 3n-3** | TRACE ELEMENTS |  |  |  |  |  | PERSISTENT ORGANIC POLLUTANTS |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | MeHg | Cd | Pb | OTC | $\mathrm{As}_{\text {T }}$ | As ${ }_{\text {inorg }}$ | PBDE | PCDD/F | PCB-DL | Tot diox | iPCB |
| Anchovy | 5.43 | 0.58 | 7.30 | 4.76 | 4.80 | 0.47 | 1.50 | 3.83 | 0.58 | 1.07 | 0.93 | 1.06 |
| Angler fish* | 0.11 | 2.97 | 0.01 | 0.25 | 1.72 | 2.43 | 0.64 | 1.13 | 0.17 | 0.15 | 0.15 | 0.15 |
| Catshark* | 0.03 | 0.37 | 0.00 | 0.05 | 0.20 | 0.93 | 0.10 | 0.08 | 0.01 | 0.01 | 0.01 | 0.03 |
| Cockle | 0.10 | 0.21 | 0.36 | 0.29 | 1.61 | 0.22 | 2.07 | 0.22 | 0.25 | 0.17 | 0.19 | 0.09 |
| Cod | 0.87 | 6.94 | 0.15 | 0.37 | 4.43 | 6.35 | 1.99 | 4.52 | 0.91 | 1.58 | 1.39 | 1.48 |
| Crab | 2.24 | 1.02 | 53.11 | 1.32 | 0.79 | 3.50 | 9.30 | 0.97 | 8.64 | 3.05 | 4.55 | 1.67 |
| Dab | 0.12 | 2.88 | 0.00 | 0.01 | 0.70 | 3.89 | 1.23 | 0.54 | 0.17 | 0.13 | 0.14 | 0.18 |
| Eel* | 0.22 | 0.49 | 0.01 | 0.17 | 0.16 | 0.02 | 0.02 | 2.16 | 0.60 | 3.78 | 3.42 | 4.70 |
| Emperor* | 0.74 | 2.48 | 0.04 | 0.00 | 0.44 | 0.09 | 0.12 | 0.66 | 2.25 | 3.15 | 2.90 | 2.63 |
| Fish soup | 2.00 | 0.52 | 1.04 | 1.91 | 1.80 | 0.81 | 3.12 | 1.81 | 1.01 | 0.94 | 0.95 | 0.80 |
| Goatfish | 1.35 | 1.06 | 0.00 | 0.24 | 0.22 | 2.01 | 2.17 | 0.79 | 1.53 | 1.88 | 1.78 | 1.72 |
| Great scallop | 0.80 | 4.10 | 0.01 | 1.31 | 4.79 | 0.93 | 2.56 | 1.54 | 1.06 | 0.37 | 0.55 | 0.44 |
| Grenadier/Hoki* | 0.19 | 0.81 | 0.01 | 0.08 | 1.35 | 1.63 | 0.57 | 1.09 | 0.50 | 0.31 | 0.35 | 0.91 |
| Gurnard | 0.03 | 0.71 | 0.00 | 0.00 | 0.15 | 0.44 | 0.20 | 0.31 | 0.78 | 0.59 | 0.63 | 0.61 |
| Haddock | 0.06 | 0.56 | 0.00 | 0.20 | 0.54 | 1.27 | 0.42 | 0.36 | 0.14 | 0.10 | 0.11 | 0.14 |
| Hake | 0.42 | 11.26 | 0.01 | 0.97 | 6.10 | 6.71 | 3.95 | 2.87 | 0.67 | 1.59 | 1.35 | 2.01 |
| Halibut* | 1.31 | 0.38 | 0.00 | 0.00 | 0.91 | 0.44 | 0.23 | 0.74 | 1.84 | 0.92 | 1.16 | 0.68 |
| Herring | 3.60 | 0.23 | 0.00 | 0.14 | 0.43 | 0.12 | 0.15 | 0.64 | 0.75 | 0.34 | 0.44 | 0.35 |
| John dory | 0.16 | 0.24 | 0.39 | 0.53 | 0.42 | 0.08 | 0.07 | 0.34 | 0.20 | 0.28 | 0.25 | 0.41 |
| Ling | 0.21 | 9.17 | 0.67 | 0.03 | 2.33 | 2.01 | 0.96 | 1.33 | 0.21 | 0.34 | 0.30 | 0.53 |
| Mackerel | 12.92 | 2.42 | 0.86 | 0.49 | 7.22 | 1.82 | 5.00 | 8.60 | 5.15 | 6.67 | 6.23 | 6.42 |
| Mussel | 0.98 | 1.27 | 2.02 | 11.31 | 1.10 | 7.30 | 4.46 | 0.82 | 2.64 | 1.33 | 1.66 | 0.79 |
| Oyster | 0.61 | 0.31 | 0.01 | 9.16 | 2.19 | 1.13 | 0.56 | 1.20 | 2.79 | 0.98 | 1.44 | 0.79 |
| Paella | 2.99 | 0.04 | 1.36 | 3.65 | 2.00 | 0.14 | 0.90 | 1.64 | 0.86 | 0.22 | 0.38 | 0.18 |
| Periwinkle | 0.47 | 0.08 | 1.75 | 2.35 | 0.58 | 1.02 | 2.14 | 0.59 | 0.20 | 0.12 | 0.14 | 0.13 |
| Pilchard | 0.45 | 0.03 | 0.02 | 0.03 | 0.13 | 0.06 | 0.26 | 0.52 | 0.57 | 0.21 | 0.30 | 0.18 |
| Plaice | 0.05 | 0.20 | 0.00 | 0.05 | 0.38 | 0.66 | 0.30 | 0.22 | 0.20 | 0.14 | 0.15 | 0.16 |
| Pollack | 0.25 | 1.19 | 0.06 | 0.07 | 1.42 | 1.48 | 5.21 | 1.27 | 0.19 | 1.38 | 1.07 | 1.67 |
| Pout | 0.05 | 0.83 | 0.00 | 0.07 | 0.40 | 1.15 | 0.80 | 0.26 | 0.09 | 0.11 | 0.10 | 0.10 |
| Ray* | 0.54 | 3.19 | 0.01 | 0.03 | 0.74 | 11.84 | 9.15 | 1.70 | 0.90 | 0.39 | 0.52 | 0.41 |
| Saithe / Coalfish | 0.98 | 4.50 | 15.32 | 28.00 | 5.90 | 2.03 | 3.82 | 3.55 | 0.82 | 1.56 | 1.36 | 1.64 |
| Salmon | 25.18 | 3.51 | 0.03 | 0.31 | 2.28 | 2.66 | 7.23 | 17.04 | 11.06 | 11.56 | 11.29 | 11.42 |
| Sardine | 17.45 | 0.73 | 4.36 | 17.51 | 4.24 | 3.45 | 1.83 | 8.99 | 25.52 | 29.25 | 28.05 | 30.20 |
| Seabass* | 2.10 | 2.92 | 0.03 | 0.81 | 1.48 | 0.74 | 2.39 | 5.86 | 4.18 | 6.91 | 6.16 | 7.09 |
| Scampi | 0.67 | 3.49 | 2.64 | 5.72 | 1.32 | 6.76 | 6.77 | 1.33 | 2.84 | 1.00 | 1.47 | 0.79 |
| Sea bream* | 2.28 | 1.07 | 0.00 | 0.12 | 1.96 | 0.82 | 4.92 | 2.02 | 1.98 | 6.35 | 5.35 | 7.58 |
| Shrimp | 0.74 | 1.98 | 0.58 | 0.04 | 3.82 | 0.84 | 1.54 | 1.46 | 2.04 | 0.79 | 1.13 | 0.35 |
| Sole | 0.43 | 4.41 | 0.01 | 0.72 | 3.41 | 7.52 | 0.89 | 1.84 | 0.74 | 0.66 | 0.67 | 1.37 |
| Spider crab | 1.67 | 0.42 | 4.15 | 2.61 | 0.41 | 6.02 | 3.90 | 3.84 | 8.38 | 4.49 | 5.48 | 2.58 |
| Squid | 0.17 | 0.48 | 0.01 | 1.14 | 1.45 | 1.86 | 0.39 | 0.55 | 0.77 | 0.56 | 0.60 | 0.65 |
| Surimi | 1.42 | 0.45 | 0.23 | 0.33 | 3.98 | 0.17 | 0.53 | 1.59 | 0.14 | 0.09 | 0.11 | 0.47 |
| Swimcrab | 0.94 | 0.52 | 0.64 | 2.09 | 0.18 | 1.16 | 0.29 | 1.12 | 3.90 | 2.19 | 2.65 | 1.51 |
| Swordfish* | 1.52 | 3.68 | 0.59 | 0.00 | 2.35 | 0.16 | 1.23 | 0.60 | 0.17 | 0.37 | 0.32 | 0.24 |
| Tarama | 1.40 | 0.01 | 0.00 | 0.01 | 0.12 | 0.03 | 0.10 | 0.77 | 0.09 | 0.07 | 0.07 | 0.11 |
| Tuna* | 3.57 | 12.30 | 2.20 | 0.71 | 15.60 | 2.21 | 1.58 | 5.03 | 0.56 | 0.75 | 0.69 | 1.16 |
| Whiting | 0.19 | 2.98 | 0.01 | 0.03 | 1.46 | 2.62 | 2.42 | 1.68 | 0.92 | 1.11 | 1.05 | 1.41 |

[^32]| Species | $\begin{gathered} \text { OMEGA } 3 \\ \mathrm{n}-\mathbf{3}^{* *} \end{gathered}$ | TRACE ELEMENTS |  |  |  |  |  | PERSISTENT ORGANIC POLLUTANTS |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | MeHg | Cd | Pb | OTC | $\mathrm{As}_{\text {T }}$ | As ${ }_{\text {inorg }}$ | PBDE | PCDD/F | PCB-DL | Tot diox | iPCB |
| Anchovy | 3.78 | 0.40 | 7.69 | 3.25 | 2.65 | 0.41 | 1.54 | 2.43 | 0.35 | 0.59 | 0.52 | 0.51 |
| Angler fish* | 0.11 | 1.88 | 0.01 | 0.01 | 0.57 | 2.02 | 0.59 | 0.83 | 0.15 | 0.08 | 0.09 | 0.05 |
| Calico scallop | 1.16 | 0.27 | 19.12 | 6.45 | 2.12 | 1.65 | 0.91 | 0.75 | 2.17 | 0.71 | 1.12 | 1.15 |
| Catshark* | 0.16 | 2.53 | 0.03 | 0.21 | 0.83 | 4.88 | 1.57 | 0.30 | 0.05 | 0.13 | 0.10 | 0.49 |
| Cockle | 0.03 | 0.02 | 0.13 | 0.70 | 0.27 | 0.09 | 0.56 | 0.04 | 0.05 | 0.01 | 0.02 | 0.01 |
| Cod | 1.50 | 6.92 | 0.06 | 0.10 | 8.80 | 8.58 | 4.16 | 3.78 | 1.20 | 2.11 | 1.78 | 2.04 |
| Crab | 2.21 | 0.58 | 1.21 | 0.28 | 0.44 | 3.34 | 2.59 | 0.81 | 4.25 | 0.92 | 1.70 | 0.55 |
| Cuttle fish | 0.77 | 1.47 | 0.53 | 0.01 | 1.41 | 3.38 | 1.73 | 0.44 | 0.17 | 0.06 | 0.08 | 0.29 |
| Dab | 0.25 | 0.37 | 0.00 | 0.15 | 0.20 | 1.29 | 0.21 | 0.69 | 0.94 | 0.51 | 0.58 | 0.31 |
| Eel* | 1.93 | 2.46 | 0.13 | 0.87 | 0.89 | 0.15 | 0.28 | 11.55 | 3.90 | 18.23 | 16.73 | 21.55 |
| Emperor* | 0.69 | 1.34 | 0.03 | 0.69 | 0.30 | 0.10 | 0.18 | 0.40 | 1.63 | 1.34 | 1.36 | 1.16 |
| Fish soup | 2.47 | 0.66 | 2.38 | 2.22 | 1.80 | 1.31 | 6.09 | 2.03 | 1.20 | 1.20 | 1.16 | 0.95 |
| Goatfish | 1.15 | 1.37 | 0.02 | 0.22 | 0.46 | 2.99 | 2.29 | 0.78 | 1.67 | 1.72 | 1.65 | 1.05 |
| Great scallop | 1.02 | 0.48 | 2.28 | 11.03 | 4.37 | 2.08 | 2.04 | 1.46 | 5.30 | 1.89 | 2.73 | 4.04 |
| Grenadier/Hoki* | 0.52 | 3.75 | 0.06 | 0.01 | 1.63 | 1.46 | 0.98 | 0.64 | 0.55 | 0.20 | 0.29 | 0.38 |
| Haddock | 0.13 | 0.98 | 0.24 | 0.10 | 0.53 | 0.55 | 0.54 | 0.69 | 0.31 | 0.30 | 0.28 | 0.30 |
| Hake | 0.51 | 5.25 | 0.02 | 3.16 | 2.08 | 3.08 | 1.65 | 2.08 | 0.86 | 2.83 | 2.26 | 2.87 |
| Halibut* | 4.67 | 1.30 | 0.01 | 0.01 | 0.49 | 0.74 | 0.48 | 1.35 | 2.47 | 1.39 | 1.62 | 1.89 |
| Herring | 4.85 | 0.25 | 0.00 | 0.10 | 0.40 | 0.16 | 0.23 | 0.65 | 0.71 | 0.32 | 0.40 | 0.30 |
| Ling | 0.21 | 1.10 | 0.01 | 0.02 | 1.28 | 1.68 | 0.88 | 1.26 | 0.22 | 0.39 | 0.33 | 0.44 |
| Mackerel | 10.39 | 1.90 | 1.69 | 0.82 | 6.39 | 1.65 | 3.12 | 5.02 | 3.31 | 3.43 | 3.24 | 3.14 |
| Mussel | 1.64 | 1.91 | 5.41 | 16.41 | 2.34 | 1.90 | 2.76 | 2.44 | 1.62 | 0.47 | 0.76 | 0.36 |
| Oyster | 1.51 | 0.63 | 11.29 | 11.13 | 2.57 | 1.89 | 14.42 | 1.41 | 5.62 | 3.14 | 3.61 | 2.32 |
| Paella | 1.71 | 0.03 | 1.08 | 1.84 | 1.01 | 0.12 | 0.84 | 0.89 | 0.46 | 0.12 | 0.21 | 0.10 |
| Periwinkle | 0.11 | 0.03 | 0.84 | 1.26 | 0.42 | 0.68 | 2.95 | 0.16 | 0.08 | 0.05 | 0.06 | 0.04 |
| Pilchard | 0.58 | 0.04 | 0.04 | 0.04 | 0.14 | 0.07 | 0.42 | 0.52 | 0.58 | 0.24 | 0.32 | 0.20 |
| Pollack | 0.11 | 0.86 | 0.00 | 0.01 | 0.87 | 0.74 | 0.36 | 0.38 | 0.06 | 0.16 | 0.13 | 0.14 |
| Ray* | 0.67 | 4.42 | 9.41 | 11.48 | 1.63 | 15.38 | 5.89 | 1.23 | 1.53 | 0.46 | 0.72 | 0.33 |
| Saithe / Coalfish | 3.87 | 2.26 | 0.58 | 0.06 | 8.62 | 1.73 | 1.26 | 8.43 | 0.75 | 1.08 | 0.95 | 0.65 |
| Salmon | 27.14 | 3.61 | 0.04 | 0.27 | 4.65 | 2.81 | 5.28 | 20.90 | 21.52 | 17.12 | 17.69 | 14.78 |
| Sardine | 8.66 | 6.11 | 5.92 | 15.79 | 6.27 | 4.29 | 3.63 | 7.54 | 16.57 | 20.58 | 19.16 | 18.71 |
| Seabass* | 1.14 | 2.97 | 0.01 | 0.20 | 3.93 | 1.28 | 1.56 | 1.85 | 4.05 | 6.17 | 5.48 | 5.98 |
| Sea bream* | 1.70 | 2.69 | 0.01 | 0.02 | 1.55 | 1.30 | 1.49 | 1.27 | 3.38 | 3.96 | 3.67 | 3.49 |
| Scampi | 0.46 | 3.79 | 1.64 | 2.00 | 1.80 | 2.02 | 10.01 | 0.89 | 1.99 | 0.73 | 1.05 | 0.62 |
| Shrimp | 0.82 | 1.37 | 0.02 | 0.03 | 1.35 | 0.59 | 2.01 | 1.52 | 0.24 | 0.36 | 0.31 | 0.26 |
| Sole | 0.70 | 7.52 | 0.57 | 1.68 | 1.02 | 11.83 | 2.36 | 2.66 | 1.09 | 0.86 | 0.87 | 2.43 |
| Squid | 0.40 | 0.62 | 3.10 | 0.25 | 1.56 | 2.28 | 0.29 | 0.63 | 1.21 | 0.97 | 0.99 | 0.95 |
| Surimi | 1.56 | 0.51 | 0.63 | 0.42 | 3.50 | 0.23 | 0.93 | 1.55 | 0.14 | 0.10 | 0.11 | 0.44 |
| Swordfish* | 2.30 | 2.16 | 0.58 | 0.01 | 2.21 | 0.10 | 0.24 | 0.60 | 0.24 | 0.30 | 0.27 | 0.11 |
| Tarama | 1.22 | 0.01 | 0.00 | 0.00 | 0.07 | 0.02 | 0.10 | 0.56 | 0.05 | 0.06 | 0.05 | 0.08 |
| Tuna* | 4.43 | 20.67 | 2.33 | 0.38 | 15.37 | 1.67 | 2.56 | 5.28 | 1.39 | 4.06 | 3.29 | 4.06 |
| Whelk | 0.59 | 0.65 | 20.69 | 6.18 | 0.90 | 6.07 | 6.15 | 0.65 | 5.82 | 0.47 | 2.11 | 0.35 |
| Whiting | 0.18 | 1.86 | 0.16 | 0.14 | 0.27 | 1.40 | 1.83 | 0.64 | 0.17 | 0.14 | 0.14 | 0.12 |

[^33]
## Contributors to the total exposure to omega 3 and contaminants (\%) - All subjects - Toulon

| Species | OMEGA 3n-3** | TRACE ELEMENTS |  |  |  |  |  | PERSISTENT ORGANIC POLLUTANTS |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | MeHg | Cd | Pb | OTC | $\mathrm{As}_{\text {T }}$ | As ${ }_{\text {inorg }}$ | PBDE | PCDD/F | PCB-DL | Tot diox | iPCB |
| Anchovy | 8.41 | 1.08 | 22.86 | 10.29 | 6.06 | 1.17 | 3.27 | 6.65 | 1.33 | 1.83 | 1.70 | 1.37 |
| Angler fish* | 0.09 | 2.43 | 0.01 | 0.51 | 1.23 | 2.34 | 0.53 | 0.83 | 0.59 | 0.54 | 0.55 | 1.30 |
| Catshark* | 0.07 | 2.52 | 0.41 | 0.01 | 0.51 | 5.59 | 1.24 | 0.12 | 0.08 | 0.07 | 0.07 | 0.13 |
| Cod | 1.50 | 5.66 | 0.17 | 4.15 | 5.43 | 8.33 | 4.73 | 7.36 | 3.93 | 2.84 | 3.06 | 1.91 |
| Crab | 1.09 | 1.30 | 4.04 | 0.71 | 0.71 | 0.49 | 1.76 | 0.88 | 0.41 | 0.12 | 0.19 | 0.04 |
| Cuttle fish | 0.09 | 0.07 | 1.30 | 4.74 | 0.25 | 0.18 | 0.55 | 0.17 | 0.49 | 0.23 | 0.29 | 0.30 |
| Dab | 0.75 | 0.79 | 0.02 | 0.23 | 0.32 | 6.00 | 2.31 | 1.13 | 1.80 | 0.82 | 1.05 | 0.53 |
| Eel* | 0.19 | 0.32 | 0.02 | 0.19 | 0.08 | 0.02 | 0.03 | 1.80 | 0.89 | 3.10 | 2.92 | 3.52 |
| Emperor* | 0.80 | 1.16 | 0.03 | 0.04 | 0.35 | 0.07 | 0.06 | 0.41 | 1.57 | 1.51 | 1.51 | 1.36 |
| Fish soup | 2.51 | 0.75 | 4.54 | 3.45 | 1.56 | 1.42 | 5.78 | 2.59 | 2.17 | 1.75 | 1.83 | 1.31 |
| Goatfish | 1.44 | 1.84 | 0.01 | 0.02 | 0.16 | 4.90 | 1.19 | 1.24 | 5.28 | 5.27 | 5.24 | 3.97 |
| Great scallop | 0.51 | 0.47 | 19.81 | 8.10 | 1.97 | 4.02 | 14.13 | 1.08 | 0.99 | 0.23 | 0.40 | 0.25 |
| Grenadier/Hoki* | 0.11 | 0.52 | 0.19 | 0.01 | 0.55 | 0.59 | 0.23 | 0.32 | 0.30 | 0.16 | 0.19 | 0.36 |
| Haddock | 0.04 | 0.01 | 0.00 | 0.06 | 0.02 | 0.02 | 0.06 | 0.04 | 0.03 | 0.01 | 0.01 | 0.01 |
| Hake | 0.38 | 0.91 | 0.02 | 0.62 | 0.72 | 0.34 | 0.34 | 0.56 | 0.50 | 0.28 | 0.33 | 0.21 |
| Halibut* | 3.97 | 0.17 | 3.76 | 11.07 | 6.47 | 2.22 | 1.25 | 2.23 | 5.78 | 2.60 | 3.41 | 1.86 |
| Herring | 3.21 | 0.25 | 0.00 | 0.15 | 0.26 | 0.16 | 0.19 | 0.63 | 1.25 | 0.48 | 0.65 | 0.42 |
| John dory | 0.09 | 0.84 | 0.00 | 0.01 | 1.34 | 0.14 | 0.52 | 0.31 | 0.16 | 0.38 | 0.33 | 0.37 |
| Lobster | 0.17 | 0.58 | 4.46 | 0.11 | 0.11 | 0.90 | 0.70 | 0.32 | 2.58 | 0.97 | 1.36 | 0.42 |
| Ling | 0.18 | 5.04 | 0.01 | 0.07 | 0.71 | 2.21 | 0.56 | 0.90 | 0.92 | 0.58 | 0.65 | 0.72 |
| Mackerel | 7.23 | 1.27 | 1.58 | 0.44 | 3.32 | 0.80 | 0.97 | 4.25 | 3.26 | 3.87 | 3.70 | 4.50 |
| Mussel | 0.39 | 1.10 | 0.02 | 0.03 | 0.63 | 1.73 | 1.06 | 1.29 | 2.41 | 1.36 | 1.58 | 1.74 |
| Octopus | 0.15 | 2.91 | 1.03 | 3.43 | 1.10 | 10.69 | 6.69 | 0.36 | 0.58 | 0.58 | 0.58 | 0.46 |
| Oyster | 0.20 | 0.03 | 1.60 | 0.14 | 2.09 | 1.13 | 4.07 | 0.72 | 2.15 | 0.77 | 1.07 | 0.56 |
| Paella | 3.18 | 0.05 | 3.92 | 5.85 | 1.51 | 0.27 | 1.62 | 2.01 | 1.66 | 0.37 | 0.68 | 0.27 |
| Pilchard | 0.23 | 0.02 | 0.03 | 0.02 | 0.06 | 0.04 | 0.17 | 0.32 | 0.43 | 0.15 | 0.21 | 0.11 |
| Ray* | 0.37 | 1.09 | 0.01 | 0.02 | 0.73 | 2.97 | 0.44 | 0.86 | 0.95 | 0.63 | 0.69 | 0.58 |
| Saithe / Coalfish | 1.25 | 1.44 | 0.36 | 0.13 | 5.91 | 2.38 | 4.40 | 2.91 | 0.90 | 1.14 | 1.07 | 0.93 |
| Salmon | 23.87 | 2.22 | 0.07 | 2.20 | 11.45 | 3.70 | 3.60 | 18.81 | 15.13 | 15.88 | 15.61 | 14.23 |
| Sardine | 6.14 | 2.90 | 8.70 | 18.69 | 3.30 | 5.38 | 9.02 | 5.29 | 12.14 | 16.67 | 15.64 | 19.48 |
| Scorpion fish | 0.82 | 1.90 | 0.01 | 0.01 | 0.25 | 0.40 | 0.41 | 0.70 | 2.38 | 2.49 | 2.45 | 1.86 |
| Seabass* | 6.10 | 2.19 | 0.03 | 2.62 | 10.05 | 1.17 | 1.85 | 6.25 | 7.84 | 12.65 | 11.56 | 13.50 |
| Sea bream* | 8.17 | 4.19 | 0.06 | 0.08 | 2.92 | 4.44 | 3.53 | 6.95 | 10.23 | 10.09 | 10.05 | 7.51 |
| Sea urchin | 0.58 | 0.12 | 4.90 | 13.53 | 2.70 | 2.42 | 11.90 | 1.14 | 0.99 | 1.72 | 1.55 | 0.83 |
| Shrimp | 0.34 | 1.43 | 1.62 | 2.34 | 0.80 | 1.97 | 0.74 | 1.21 | 0.91 | 0.13 | 0.30 | 0.08 |
| Sole | 0.76 | 6.99 | 0.12 | 1.95 | 2.64 | 10.70 | 1.23 | 1.63 | 1.83 | 1.37 | 1.47 | 3.39 |
| Squid | 0.77 | 0.93 | 2.67 | 0.69 | 0.73 | 0.31 | 0.58 | 0.80 | 0.21 | 0.14 | 0.16 | 0.10 |
| Surimi | 2.83 | 1.03 | 1.66 | 1.09 | 5.04 | 0.57 | 1.72 | 3.37 | 0.50 | 0.27 | 0.32 | 1.08 |
| Swordfish* | 2.86 | 7.50 | 1.95 | 0.01 | 0.65 | 0.16 | 0.13 | 0.68 | 0.35 | 0.69 | 0.61 | 0.76 |
| Tarama | 2.29 | 0.02 | 0.01 | 0.01 | 0.14 | 0.06 | 0.21 | 1.37 | 0.24 | 0.17 | 0.19 | 0.26 |
| Tuna* | 5.11 | 23.93 | 7.16 | 1.21 | 12.14 | 3.48 | 2.24 | 5.73 | 1.16 | 1.33 | 1.28 | 1.96 |
| Whelk | 0.10 | 0.23 | 0.79 | 0.32 | 0.19 | 1.66 | 1.22 | 0.29 | 0.84 | 0.23 | 0.36 | 0.14 |
| Whiting | 0.64 | 9.83 | 0.04 | 0.69 | 2.86 | 2.44 | 2.75 | 3.52 | 1.86 | 3.57 | 3.17 | 5.32 |

[^34]Appendix 6: Contributors to the recommendation for EPA and DHA and to the tolerable intake of contaminants (\%) - All subjects, all areas

|  | OMEGA 3EPA+DHA | TRACE ELEMENTS |  |  |  |  |  | PERSISTENT ORGANIC POLLUTANTS Tot diox iPCB |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | MeHg | Cd | Pb | OTC | $\mathrm{As}_{\text {T }}$ | As ${ }_{\text {inorg }}$ |  |  |
| Anchovy | 13.3 | 0.49 | 1.90 | 0.09 | 0.33 | 0.11 | 0.07 | 0.86 | 1.92 |
| Angler fish* | 0.27 | 2.15 | 0.00 | 0.00 | 0.05 | 0.45 | 0.03 | 0.17 | 0.60 |
| Calico scallop | 0.89 | 0.06 | 2.04 | 0.05 | 0.05 | 0.09 | 0.01 | 0.26 | 0.56 |
| Catshark* | 0.66 | 1.58 | 1.77 | 0.01 | 0.10 | 1.24 | 0.05 | 0.12 | 0.63 |
| Cockle | 0.08 | 0.05 | 0.02 | 0.00 | 0.05 | 0.02 | 0.03 | 0.05 | 0.05 |
| Cod | 2.76 | 5.08 | 0.01 | 0.01 | 0.38 | 2.13 | 0.20 | 1.19 | 2.41 |
| Crab | 4.65 | 1.79 | 7.76 | 0.01 | 0.10 | 0.59 | 0.24 | 4.20 | 8.18 |
| Cuttle fish | 0.54 | 0.37 | 0.06 | 0.02 | 0.03 | 0.19 | 0.02 | 0.07 | 0.26 |
| Dab | 0.83 | 1.42 | 0.00 | 0.00 | 0.03 | 1.39 | 0.05 | 0.62 | 0.70 |
| Eel* | 1.16 | 1.01 | 0.00 | 0.00 | 0.03 | 0.01 | 0.00 | 9.79 | 55.6 |
| Emperor* | 1.31 | 1.79 | 0.00 | 0.00 | 0.03 | 0.02 | 0.00 | 2.18 | 4.11 |
| Fish soup | 3.29 | 0.41 | 0.20 | 0.03 | 0.10 | 0.18 | 0.15 | 0.74 | 1.21 |
| Goatfish | 3.15 | 1.20 | 0.00 | 0.00 | 0.03 | 0.69 | 0.07 | 2.40 | 4.04 |
| Great scallop | 2.47 | 1.22 | 2.60 | 0.16 | 0.80 | 0.46 | 0.40 | 1.44 | 4.36 |
| Grenadier / hoki* | 0.46 | 1.57 | 0.01 | 0.00 | 0.05 | 0.24 | 0.02 | 0.21 | 0.79 |
| Gurnard | 0.07 | 0.24 | 0.00 | 0.00 | 0.00 | 0.05 | 0.00 | 0.26 | 0.51 |
| Haddock | 0.17 | 0.45 | 0.00 | 0.00 | 0.00 | 0.12 | 0.01 | 0.13 | 0.28 |
| Hake | 0.78 | 4.99 | 0.00 | 0.02 | 0.13 | 0.68 | 0.07 | 0.95 | 2.60 |
| Halibut* | 8.06 | 0.95 | 0.09 | 0.07 | 0.55 | 0.29 | 0.02 | 2.36 | 3.97 |
| Herring | 14.0 | 0.19 | 0.00 | 0.00 | 0.03 | 0.03 | 0.01 | 0.38 | 0.57 |
| John dory | 0.24 | 0.27 | 0.02 | 0.00 | 0.05 | 0.01 | 0.01 | 0.13 | 0.37 |
| Ling | 0.60 | 5.54 | 0.02 | 0.00 | 0.05 | 0.38 | 0.03 | 0.25 | 0.68 |
| Lobster | 0.29 | 0.14 | 0.15 | 0.00 | 0.00 | 0.05 | 0.01 | 0.23 | 0.15 |
| Mackerel | 28.0 | 1.88 | 0.14 | 0.01 | 0.40 | 0.31 | 0.12 | 8.45 | 20.8 |
| Mussel | 2.00 | 0.86 | 0.20 | 0.15 | 0.10 | 0.64 | 0.18 | 1.05 | 1.64 |
| Octopus | 0.12 | 0.75 | 0.03 | 0.01 | 0.03 | 0.67 | 0.08 | 0.08 | 0.16 |
| Oyster | 1.33 | 0.24 | 0.40 | 0.09 | 0.25 | 0.26 | 0.29 | 1.72 | 1.92 |
| Paella | 3.88 | 0.03 | 0.15 | 0.04 | 0.10 | 0.03 | 0.03 | 0.23 | 0.20 |
| Periwinkle | 0.40 | 0.03 | 0.16 | 0.02 | 0.03 | 0.11 | 0.07 | 0.05 | 0.09 |
| Pilchard | 1.13 | 0.03 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.27 | 0.33 |
| Plaice | 0.32 | 0.60 | 0.00 | 0.00 | 0.05 | 0.52 | 0.01 | 0.83 | 1.62 |
| Pollack | 0.34 | 0.86 | 0.00 | 0.00 | 0.05 | 0.19 | 0.08 | 0.33 | 1.03 |
| Pout | 0.04 | 0.20 | 0.00 | 0.00 | 0.00 | 0.08 | 0.01 | 0.03 | 0.06 |
| Ray* | 1.22 | 2.58 | 0.28 | 0.05 | 0.10 | 2.79 | 0.23 | 0.49 | 0.76 |
| Saithe | 3.78 | 2.21 | 0.94 | 0.15 | 0.45 | 0.35 | 0.12 | 0.64 | 1.34 |
| Salmon | 66.2 | 1.33 | 0.00 | 0.00 | 0.23 | 0.27 | 0.09 | 6.15 | 11.6 |
| Sardine | 23.8 | 2.63 | 0.65 | 0.27 | 0.25 | 0.77 | 0.16 | 24.0 | 57.4 |
| Scampi | 0.73 | 1.63 | 0.24 | 0.03 | 0.08 | 0.64 | 0.18 | 0.67 | 0.67 |
| Scorpion fish | 0.63 | 0.41 | 0.00 | 0.00 | 0.00 | 0.02 | 0.00 | 0.45 | 0.77 |
| Sea bream* | 8.96 | 2.29 | 0.00 | 0.00 | 0.13 | 0.39 | 0.14 | 5.93 | 13.9 |
| Sea urchin | 0.34 | 0.03 | 0.17 | 0.11 | 0.10 | 0.17 | 0.27 | 0.30 | 0.35 |
| Seabass* | 7.12 | 2.96 | 0.00 | 0.02 | 0.45 | 0.20 | 0.07 | 7.94 | 19.0 |
| Shrimp | 1.59 | 1.14 | 10.97 | 0.02 | 0.10 | 0.22 | 0.06 | 0.40 | 0.40 |
| Sole | 1.02 | 5.16 | 0.02 | 0.01 | 0.10 | 2.55 | 0.06 | 0.77 | 4.34 |
| Spider crab | 1.03 | 0.10 | 0.30 | 0.01 | 0.00 | 0.48 | 0.06 | 1.55 | 1.27 |
| Squid | 1.10 | 0.85 | 0.18 | 0.01 | 0.40 | 0.41 | 0.02 | 1.30 | 2.19 |
| Surimi | 3.99 | 0.53 | 0.06 | 0.01 | 0.40 | 0.05 | 0.03 | 0.10 | 0.84 |
| Swimcrab | 2.22 | 0.41 | 0.18 | 0.05 | 0.13 | 0.26 | 0.04 | 12.1 | 29.0 |
| Swordfish* | 5.89 | 6.48 | 0.11 | 0.00 | 0.13 | 0.03 | 0.02 | 0.32 | 0.68 |
| Tarama | 2.15 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.05 | 0.16 |
| Tuna* | 8.62 | 17.5 | 0.27 | 0.01 | 1.05 | 0.45 | 0.07 | 1.35 | 3.73 |
| Whelk | 0.72 | 0.44 | 2.80 | 0.06 | 0.15 | 0.97 | 0.10 | 0.86 | 0.53 |
| Whiting | 0.63 | 4.52 | 0.00 | 0.00 | 0.10 | 0.45 | 0.09 | 0.81 | 2.86 |
| TOTAL | 239 | 91.7 | 34.9 | 1.62 | 8.23 | 23.7 | 4.19 | 108 | 274 |

In bold: mean contributors (>5\%)

* Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005

For certain species the different packaging are taken into account: Herring: fresh and canned, Mackerel: fresh, canned and smoked, Sardine: fresh and canned, Salmon: fresh and smoked, Tuna :fresh and canned, Anchovy: fresh and canned, Crab: fresh and canned, Haddock: fresh and smoked
COT=TBT+DBT+TPT+DOT; Tot diox=PCDD/F+DL-PCB

Contributors to the recommendation for EPA and DHA and to the tolerable intake of contaminants (\%) - All subjects, Le Havre


In bold: mean contributors (>5\%)

* Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005

For certain species the different packaging are taken into account: Herring: fresh and canned, Mackerel: fresh, canned and smoked,
Sardine: fresh and canned, Salmon: fresh and smoked, Tuna :fresh and canned, Anchovy: fresh and canned, Crab: fresh and canned,
Haddock: fresh and smoked
COT=TBT+DBT+TPT+DOT; Tot diox=PCDD/F+DL-PCB

Contributors to the recommendation for EPA and DHA and to the tolerable intake of contaminants (\%) - All subjects, Lorient


In bold: mean contributors (>5\%)

* Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005

For certain species the different packaging are taken into account: Herring: fresh and canned, Mackerel: fresh, canned and smoked, Sardine: fresh and canned, Salmon: fresh and smoked, Tuna :fresh and canned, Anchovy: fresh and canned, Crab: fresh and canned, Haddock: fresh and smoked
COT=TBT+DBT+TPT+DOT; Tot diox=PCDD/F+DL-PCB

Contributors to the recommendation for EPA and DHA and to the tolerable intake of contaminants (\%) - All subjects, La Rochelle

|  | OMEGA 3EPA+DHA | TRACE ELEMENTS |  |  |  |  |  | PERSISTENT ORGANIC POLLUTANTS Tot diox iPCB |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | MeHg | Cd | Pb | OTC | $\mathrm{As}_{\text {T }}$ | As ${ }_{\text {inorg }}$ |  |  |
| Anchovy | 10.9 | 0.40 | 1.20 | 0.06 | 0.23 | 0.09 | 0.05 | 0.90 | 2.18 |
| Angler fish* | 0.27 | 1.99 | 0.00 | 0.00 | 0.03 | 0.45 | 0.02 | 0.10 | 0.12 |
| Calico scallop | 2.27 | 0.23 | 8.20 | 0.19 | 0.23 | 0.35 | 0.03 | 1.06 | 2.27 |
| Catshark* | 0.36 | 2.48 | 0.00 | 0.00 | 0.05 | 1.32 | 0.05 | 0.08 | 0.83 |
| Cockle | 0.05 | 0.01 | 0.02 | 0.01 | 0.03 | 0.02 | 0.02 | 0.02 | 0.02 |
| Cod | 2.89 | 5.73 | 0.00 | 0.00 | 0.43 | 1.55 | 0.11 | 1.37 | 3.19 |
| Crab | 4.86 | 0.52 | 0.14 | 0.00 | 0.03 | 0.61 | 0.08 | 1.66 | 1.11 |
| Cuttle fish | 1.59 | 1.41 | 0.06 | 0.00 | 0.08 | 0.74 | 0.06 | 0.08 | 0.56 |
| Dab | 0.58 | 0.35 | 0.00 | 0.00 | 0.03 | 0.30 | 0.01 | 0.67 | 0.74 |
| Eel* | 3.15 | 2.70 | 0.01 | 0.01 | 0.08 | 0.03 | 0.01 | 26.3 | 149 |
| Emperor* | 1.39 | 2.42 | 0.00 | 0.02 | 0.03 | 0.03 | 0.01 | 3.38 | 7.31 |
| Fish soup | 4.07 | 0.49 | 0.24 | 0.03 | 0.13 | 0.21 | 0.18 | 0.89 | 1.45 |
| Goatfish | 3.21 | 1.80 | 0.00 | 0.00 | 0.05 | 0.83 | 0.08 | 2.81 | 4.00 |
| Great scallop | 2.16 | 0.38 | 0.25 | 0.20 | 0.35 | 0.36 | 0.06 | 2.51 | 7.82 |
| Grenadier / hoki* | 0.95 | 4.38 | 0.01 | 0.00 | 0.10 | 0.34 | 0.03 | 0.27 | 0.73 |
| Haddock | 0.27 | 1.31 | 0.01 | 0.00 | 0.03 | 0.14 | 0.02 | 0.44 | 0.95 |
| Hake | 0.97 | 4.77 | 0.00 | 0.05 | 0.13 | 0.57 | 0.05 | 2.27 | 6.08 |
| Halibut* | 12.0 | 1.56 | 0.00 | 0.00 | 0.03 | 0.18 | 0.02 | 2.23 | 5.55 |
| Herring | 10.3 | 0.19 | 0.00 | 0.00 | 0.03 | 0.03 | 0.01 | 0.38 | 0.57 |
| Ling | 0.46 | 0.87 | 0.00 | 0.00 | 0.05 | 0.32 | 0.03 | 0.27 | 0.73 |
| Mackerel | 18.3 | 1.49 | 0.15 | 0.01 | 0.33 | 0.26 | 0.09 | 3.59 | 8.08 |
| Mussel | 2.89 | 1.38 | 0.51 | 0.27 | 0.18 | 0.29 | 0.07 | 0.52 | 0.48 |
| Oyster | 2.35 | 0.44 | 1.51 | 0.19 | 0.15 | 0.31 | 0.47 | 3.08 | 3.93 |
| Paella | 2.78 | 0.02 | 0.10 | 0.03 | 0.08 | 0.02 | 0.02 | 0.16 | 0.14 |
| Periwinkle | 0.04 | 0.02 | 0.13 | 0.02 | 0.03 | 0.13 | 0.11 | 0.05 | 0.06 |
| Pilchard | 1.23 | 0.03 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.29 | 0.35 |
| Pollack | 0.24 | 0.98 | 0.00 | 0.00 | 0.05 | 0.19 | 0.01 | 0.13 | 0.28 |
| Ray* | 1.31 | 4.25 | 1.10 | 0.21 | 0.08 | 3.99 | 0.18 | 0.58 | 0.57 |
| Saithe | 8.11 | 1.79 | 0.04 | 0.00 | 0.75 | 0.29 | 0.03 | 0.67 | 0.95 |
| Salmon | 68.8 | 1.51 | 0.00 | 0.00 | 0.18 | 0.26 | 0.08 | 10.1 | 15.9 |
| Sardine | 19.1 | 6.65 | 0.60 | 0.27 | 0.33 | 0.79 | 0.10 | 28.2 | 59.1 |
| Scampi | 0.89 | 3.08 | 0.16 | 0.03 | 0.08 | 0.33 | 0.33 | 0.78 | 0.85 |
| Sea bream* | 4.57 | 3.24 | 0.00 | 0.00 | 0.10 | 0.32 | 0.05 | 6.02 | 12.3 |
| Seabass* | 2.42 | 3.18 | 0.00 | 0.00 | 0.18 | 0.28 | 0.05 | 6.98 | 17.2 |
| Shrimp | 1.70 | 1.07 | 0.00 | 0.00 | 0.10 | 0.10 | 0.06 | 0.24 | 0.38 |
| Sole | 1.28 | 7.14 | 0.06 | 0.03 | 0.05 | 2.57 | 0.06 | 0.75 | 4.89 |
| Squid | 0.84 | 0.53 | 0.46 | 0.00 | 0.15 | 0.43 | 0.01 | 0.97 | 1.97 |
| Surimi | 2.86 | 0.37 | 0.04 | 0.00 | 0.28 | 0.04 | 0.02 | 0.07 | 0.59 |
| Swordfish* | 6.03 | 2.87 | 0.06 | 0.00 | 0.20 | 0.02 | 0.01 | 0.24 | 0.18 |
| Tarama | 1.41 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.03 | 0.09 |
| Tuna* | 8.43 | 21.6 | 0.19 | 0.00 | 1.18 | 0.27 | 0.07 | 3.67 | 9.46 |
| Whelk | 0.97 | 0.51 | 6.38 | 0.13 | 0.05 | 1.38 | 0.25 | 2.12 | 0.61 |
| Whiting | 0.43 | 1.72 | 0.01 | 0.00 | 0.00 | 0.24 | 0.05 | 0.13 | 0.22 |
| TOTAL | 219 | 97.9 | 21.7 | 1.78 | 6.55 | 21.0 | 3.05 | 117 | 334 |

In bold: mean contributors (>5\%)

* Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005

For certain species the different packaging are taken into account: Herring: fresh and canned, Mackerel: fresh, canned and smoked,
Sardine: fresh and canned, Salmon: fresh and smoked, Tuna :fresh and canned, Anchovy: fresh and canned, Crab: fresh and canned,
Haddock: fresh and smoked
COT=TBT+DBT+TPT+DOT; Tot diox=PCDD/F+DL-PCB

Contributors to the recommendation for EPA and DHA and to the tolerable intake of contaminants (\%) - All subjects, Toulon

|  | OMEGA 3EPA+DHA | TRACE ELEMENTS |  |  |  |  |  | PERSISTENT ORGANIC POLLUTANTS Tot diox iPCB |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | MeHg | Cd | Pb | OTC | $\mathrm{As}_{\text {T }}$ | As ${ }_{\text {inorg }}$ |  |  |
| Anchovy | 20.6 | 0.79 | 3.69 | 0.17 | 0.60 | 0.18 | 0.12 | 1.02 | 1.96 |
| Angler fish* | 0.17 | 2.53 | 0.00 | 0.01 | 0.13 | 0.47 | 0.02 | 0.31 | 1.80 |
| Catshark* | 0.19 | 2.68 | 0.03 | 0.00 | 0.05 | 1.82 | 0.05 | 0.05 | 0.21 |
| Cod | 2.77 | 3.99 | 0.00 | 0.03 | 0.38 | 1.19 | 0.12 | 1.31 | 1.97 |
| Crab | 2.71 | 1.23 | 0.35 | 0.01 | 0.05 | 0.09 | 0.07 | 0.09 | 0.05 |
| Cuttle fish | 0.27 | 0.07 | 0.16 | 0.09 | 0.03 | 0.04 | 0.03 | 0.20 | 0.47 |
| Dab | 1.44 | 0.66 | 0.00 | 0.00 | 0.00 | 1.95 | 0.08 | 0.49 | 0.46 |
| Eel* | 0.52 | 0.50 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 4.84 | 27.5 |
| Emperor* | 1.38 | 1.75 | 0.00 | 0.00 | 0.05 | 0.02 | 0.00 | 1.60 | 3.02 |
| Fish soup | 4.24 | 0.53 | 0.26 | 0.03 | 0.13 | 0.23 | 0.20 | 0.96 | 1.57 |
| Goatfish | 3.74 | 1.72 | 0.00 | 0.00 | 0.00 | 1.16 | 0.04 | 4.15 | 7.01 |
| Great scallop | 1.20 | 0.39 | 2.72 | 0.10 | 0.23 | 0.74 | 0.65 | 0.22 | 0.32 |
| Grenadier / hoki* | 0.17 | 0.37 | 0.01 | 0.00 | 0.03 | 0.08 | 0.01 | 0.08 | 0.38 |
| Haddock | 0.11 | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.01 | 0.01 |
| Hake | 0.69 | 0.85 | 0.00 | 0.01 | 0.05 | 0.08 | 0.01 | 0.16 | 0.22 |
| Halibut* | 7.26 | 0.10 | 0.34 | 0.29 | 1.13 | 0.38 | 0.03 | 2.31 | 2.62 |
| Herring | 6.75 | 0.19 | 0.00 | 0.00 | 0.03 | 0.03 | 0.01 | 0.39 | 0.57 |
| John dory | 0.22 | 0.82 | 0.00 | 0.00 | 0.15 | 0.03 | 0.02 | 0.22 | 0.57 |
| Ling | 0.33 | 4.93 | 0.00 | 0.00 | 0.05 | 0.32 | 0.01 | 0.31 | 0.77 |
| Lobster | 0.36 | 0.56 | 0.60 | 0.00 | 0.00 | 0.20 | 0.03 | 0.90 | 0.61 |
| Mackerel | 19.3 | 1.49 | 0.12 | 0.01 | 0.30 | 0.17 | 0.04 | 3.07 | 10.8 |
| Mussel | 0.79 | 0.81 | 0.00 | 0.00 | 0.05 | 0.27 | 0.03 | 0.80 | 2.05 |
| Octopus | 0.37 | 2.98 | 0.10 | 0.05 | 0.10 | 2.63 | 0.34 | 0.33 | 0.61 |
| Oyster | 0.44 | 0.02 | 0.08 | 0.00 | 0.20 | 0.18 | 0.14 | 0.56 | 0.68 |
| Paella | 4.61 | 0.03 | 0.18 | 0.05 | 0.10 | 0.03 | 0.04 | 0.28 | 0.24 |
| Pilchard | 0.64 | 0.02 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.17 | 0.21 |
| Ray* | 1.03 | 1.38 | 0.00 | 0.00 | 0.10 | 0.87 | 0.02 | 0.53 | 1.00 |
| Saithe | 2.41 | 0.98 | 0.01 | 0.00 | 0.58 | 0.34 | 0.12 | 0.48 | 0.96 |
| Salmon | 50.0 | 0.88 | 0.00 | 0.01 | 0.20 | 0.28 | 0.05 | 5.12 | 10.6 |
| Sardine | 13.3 | 2.50 | 0.65 | 0.26 | 0.20 | 0.88 | 0.37 | 10.5 | 35.9 |
| Scorpion fish | 1.59 | 1.63 | 0.00 | 0.00 | 0.03 | 0.07 | 0.01 | 1.79 | 3.04 |
| Sea bream* | 19.3 | 3.76 | 0.00 | 0.00 | 0.20 | 0.83 | 0.11 | 7.35 | 12.2 |
| Sea urchin | 1.28 | 0.13 | 0.66 | 0.42 | 0.43 | 0.66 | 1.06 | 1.20 | 1.36 |
| Seabass* | 15.8 | 1.65 | 0.00 | 0.02 | 1.38 | 0.20 | 0.05 | 8.07 | 22.5 |
| Shrimp | 0.68 | 1.05 | 0.07 | 0.02 | 0.05 | 0.29 | 0.02 | 0.15 | 0.09 |
| Sole | 0.93 | 6.28 | 0.00 | 0.01 | 0.18 | 2.43 | 0.03 | 0.59 | 3.65 |
| Squid | 1.33 | 0.75 | 0.22 | 0.01 | 0.08 | 0.05 | 0.02 | 0.08 | 0.13 |
| Surimi | 4.99 | 0.67 | 0.07 | 0.01 | 0.50 | 0.07 | 0.04 | 0.13 | 1.07 |
| Swordfish* | 10.6 | 13.9 | 0.23 | 0.00 | 0.05 | 0.04 | 0.01 | 0.54 | 1.59 |
| Tarama | 3.62 | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.09 | 0.28 |
| Tuna* | 9.65 | 23.8 | 0.36 | 0.01 | 1.05 | 0.61 | 0.06 | 0.55 | 1.89 |
| Whelk | 0.21 | 0.25 | 0.08 | 0.00 | 0.03 | 0.44 | 0.06 | 0.29 | 0.27 |
| Whiting | 1.40 | 11.0 | 0.00 | 0.01 | 0.23 | 0.43 | 0.09 | 1.92 | 8.15 |
| TOTAL | 220 | 101 | 11.0 | 1.61 | 9.13 | 20.8 | 4.22 | 64.2 | 171 |

In bold: mean contributors (>5\%)

* Predatory fish as described in the Commission Regulation (EC) No 78/2005 of 19 January 2005

For certain species the different packaging are taken into account: Herring: fresh and canned, Mackerel: fresh, canned and smoked, Sardine: fresh and canned, Salmon: fresh and smoked, Tuna :fresh and canned, Anchovy: fresh and canned, Crab: fresh and canned, Haddock: fresh and smoked
COT=TBT+DBT+TPT+DOT; Tot diox=PCDD/F+DL-PCB

## CALIPSO

Fish and seafood consumption study
and biomarker of exposure to
trace elements, pollutants and omega 3.


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    For certain species the different packaging are taken into account: Herring: fresh and canned, Mackerel: fresh, canned and smoked,
    Sardine: fresh and canned, Salmon: fresh and smoked, Tuna :fresh and canned, Anchovy: fresh and canned, Crab: fresh and canned,
    Haddock: fresh and smoked
    ** omega 3 (ALA, C18:4 n-3, EPA, DPA and DHA)

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